Elsevier Editorial System(tm) for Applied Catalysis B: Environmental Manuscript Draft

Manuscript Number: APCATB-D-13-01690R1

Title: Zero-valent iron supported on nitrogen-containing activated carbon for catalytic wet peroxide oxidation of phenol

Article Type: Research Paper

Keywords: Phenol; Activated carbon; Surface chemistry; Catalytic wet peroxide oxidation

Corresponding Author: Dr. Josep --- Font, Dr

Corresponding Author's Institution: Universitat Rovira i Virgili

First Author: Selamawit Ashagre Messele, MSc

Order of Authors: Selamawit Ashagre Messele, MSc; Olivia Salome G. P. Soares, PhD; José J. M. Órfão, PhD; Frank Stüber, PhD; Christophe Bengoa, PhD; Agustin Fortuny, PhD; Azael Fabregat, PhD; Josep Font, PhD



VNIVERSITAT ROVIRA I VIRGILI

ESCOLA TÈCNICA SUPERIOR D'ENGINYERIA QUÍMICA DEPARTAMENT D'ENGINYERIA QUÍMICA

Avinguda dels Països Catalans, 26 Campus Sescelades 43007 Tarragona (Spain) Tel. 34 977 55 96 46 Fax 34 977 55 96 21 e-mail: jose.font@urv.cat http://www.etseq.urv.cat/DEQ

February 5th, 2014

Editor-in-Chief Journal of Applied Catalysis B: Environmental

Dear Xenophon E. Verykios,

With this cover letter, we will submit the revised manuscript (APCATB-D-13-01690) entitled, "Zero-valent iron supported on nitrogen-containing activated carbon for catalytic wet peroxide oxidation of phenol" for publication in the journal of Applied Catalysis B: Environmental. We would like to thank you and the reviewers for the careful and constructive reviews. We do appreciate the comments and suggestions made by the reviewers and we do think that they contributed to improve the quality of the manuscript. Based on the comments and suggestions from the reviewers, we have revised the manuscript. Please find attached our responses to the questions and suggestions given by the reviewers. Each comment of the reviewers is listed in italic style followed by our responses.

Looking forward to hearing from you,

Best regards,

Tait

Dr. Josep Font Capafons

Associate Professor Department of Chemical Engineering University Rovira i Virgili Tel.: +34 977 559646 Fax: +34 977 559667 E-mail: jose.font@urv.net

Chemical Reaction Engineering & Process Intensification Group http://www.etseq.urv.es/CREPI/





ESCOLA TÈCNICA SUPERIOR D'ENGINYERIA QUÍMICA DEPARTAMENT D'ENGINYERIA QUÍMICA

Avinguda dels Països Catalans, 26 Campus Sescelades 43007 Tarragona (Spain) Tel. 34 977 55 96 46 Fax 34 977 55 96 21 e-mail: jose.font@urv.cat http://www.etseq.urv.cat/DEQ

Tarragona, November 11th, 2013

Editor-in-Chief Journal of Applied Catalysis B: Environmental

Dear Sir,

We are pleased to submit our manuscript entitled "Zero-valent iron supported on nitrogencontaining activated carbon for catalytic wet peroxide oxidation of phenol" for consideration towards the journal of Applied Catalysis B: Environmental.

The paper shows the preparation, characterization and activity test zero-valent iron supported catalysts on modified activated carbon support with different nitrogen containing precursors. The results from the different characterization techniques demonstrate that the nitrogen-containing groups are successfully introduced into the carbon surface via all the precursors used. The tests of the different modified carbons as adsorbents/catalysts indicated that the adsorption capacity and the efficiency in phenol oxidation are governed by the specific surface area and functional groups present. The modified carbons supported iron catalysts revealed significantly enhanced phenol removal efficiency, reaching over 85% conversion after 3 h. We believe that these findings will be of great interest to researchers working on heterogeneous catalysis for water treatment and your journal is the perfect platform for sharing these results with the international research community.

Finally, we confirm that this manuscript has not been previously published elsewhere and is not under consideration by another journal. All authors have approved the manuscript and agree with submission to this special issue of Applied Catalysis B: Environmental.

Looking forward to hearing from you,

Best regards,

ai Tait

Dr. Josep Font Capafons

Associate Professor Department of Chemical Engineering University Rovira i Virgili Tel.: +34 977 559646 Fax: +34 977 559667 E-mail: jose.font@urv.net

Chemical Reaction Engineering & Process Intensification Group http://www.etseq.urv.es/CREPI/



Our answers to reviewers' comments

First of all, we would like to thank the reviewers for reading the paper critically. We do appreciate the comments and suggestions made by the reviewers and we do think that they contributed to improve the quality of the manuscript. Listed below are our responses to the questions and suggestions given by the reviewers. Each comment of the reviewers is listed in italic style followed by our responses.

Reviewer #1:

Thank you very much for your constructive comments and suggestions to our manuscript submission. We have incorporated all your feedbacks into our revised manuscript. We do think that they have resulted in improving the quality of the manuscript. The following summarizes how we responded to your comments and suggestions. Each comment is listed in italic style and followed by our responses.

Page 12

Results and discussion:

The subchapter 3.1 is not needed. This information is in the experimental part.

The reviewer is right and we have removed the paragraph in the subchapter 3.1 from the manuscript.

Page 13

Why impregnation with iron did not result in any property changes?

In previous works, we observed that for small amounts of metals supported on activated carbon the textural properties of the catalysts remained practically unchanged compared to the unloaded carbon. In the preparation of the ZVI catalysts, we have used 3% (wt %) of iron, and therefore we can realistically assume that the textural properties of the supported metal catalysts are not significantly different from those of the activated carbon. In fact, it can be observed in Table 3 that the textural properties of the ZVI supported on the original AC are similar to the unloaded carbon. This information was included in section 3.1.1 of the manuscript.

Nitrogen containing carbons are known of having oxidizing properties and the authors should use this and support from the literature for their discussion.

We have included some references which have used nitrogen containing carbons as catalyst and catalyst support (refer introduction section paragraph 5).

The authors should try to clarify about the role of iron

ZVI acts as a heterogeneous catalyst for the degradation of phenol in aqueous solution by effectively decomposing hydrogen peroxide and generating hydroxyl radical, which reacts at high rate with phenol and its intermediates. This information was included in section 3.2.2.2 of the manuscript.

Reviewer #2:

We thank you very much for your careful analysis and thoughtful comments and suggestions. We have carefully taken them into consideration in preparing the revised

manuscript. We do think that your comments and suggestions contributed to the improvement of the quality of the manuscript. Below are our responses to your comments and suggestions. Each comment of the reviewer is listed in italic style followed by our responses.

Abstract

The interesting stability of the catalyst should be emphasized in the Abstract Section.

We have emphasized the stability of the catalyst in the Abstract section based on the comment made by the reviewer. We have added the last sentence of the paragraph to the previous Abstract.

Page 3, lines 3-6: "Fenton process (Fe²⁺/H₂O₂) belongs to catalytic wet peroxide oxidation (CWPO) technologies and operates under relatively mild operating conditions (room temperature and atmospheric pressure) using hydrogen peroxide as oxidant and iron as a catalyst". This can be a quite controversial statement. Fenton was developed before CWPO. The latter is an adaptation of Fenton process to overcome some process drawbacks. In fact, a wide number of researchers consider CWPO as a different AOP than Fenton, both using H₂O₂ as oxidant but, in the former, in presence of a heterogeneous catalyst and, in the latter, using a salt of Fe²⁺ (S. Perathoner, G. Centi, Top. Catal. 33 (2005) 207- 224). In both treatments, temperatures above 100 °C have been reported.

We agree with the suggestion and in order to avoid the controversy we re-wrote the sentence in the previous manuscript as: "Fenton process mainly operates under relatively mild operating conditions using hydrogen peroxide as oxidant and iron as a catalyst".

2. Page 3, lines 17-21: "On the other hand, the catalytic activity of AC is far from well understood. Besides the direct relationship with the physical properties (surface area, pore volume, etc.) of activated carbon [15], the surface chemistry can play an important role [16-18]. Moreover, it has been reported [19] that the mineral content (in particular Fe content) of AC is a key to display catalytic activities." The recent work of Dominguez et al., (2013, Carbon, 76-83) should be referred, and in general an update of the references is highly recommended. In this work, thanks to the use of cyclic voltammetry, the effects of physicochemical characteristics of carbon materials on the rate of hydrogen peroxide decomposition were elucidated. The results indicate that the most important factor in the catalytic activity was the content of metals, in particular iron, this was followed by the specific surface area and finally the content of surface oxygen groups.

We have included (in the introduction of the manuscript) the proposed reference and the main findings of the work about the importance of the presence of metals, in particular iron, on the catalytic activity.

3. Page 4, line 2: Why does the attachment of a N-functional group ligand help the oxidative power for the H_2O_2 ?, An explanation should be given, more considering

that opposite results were found in the CWPO experiments since the BET area is reduced in presence of these groups.

We revised the whole paragraph with the justification of why the attachment of Nfunctional groups helps to increase the catalytic activity of the system. (Refer paragraph 5 in the introduction section).

4. Page 4, line 3-4: "However, care should be given to the load of iron or ligand to be incorporated on the AC". Why?, This statement should be better explain.

We decided to remove this sentence to avoid ambiguity. In fact, we did not compare different iron or ligand loads in the current study.

5. In my opinion, the approach of the study should be better introduced in the Introduction Section (a). The oxidation and activation steps before Nfunctionalization are not expected at all. A good explanation about the use of thionyl chloride should be given before the Section 2.2.3(b). Can the activate carbons contain Cl on their surface?, can Cl affect the AC activity?, Shouldn't Cl be measured?(c). According to this paragraph in pag 15: "...it is important to have high density of oxygenated groups on the surface of carbon before functionalization with the N containing compounds. The N-containing groups (EDA, urea and melamine) strongly interact with carboxylic acids, anhydrides and lactones [24, 34].." why is the thionyl chloride activation required then?(d)

Several sentences were included in the manuscript:

- (a) We revised the last paragraph of the introduction section and added a sentence stating about the oxidation and activation steps before Nfunctionalization.
- (b) An explanation about the role of thionyl chloride is given in the first sentence of the paragraph in section 2.2.2. "Thionyl chloride was used as a linking agent on the surface of AC for the attachment of N-functional groups."
- (c) We didn't quantify the amount of Cl on the modified AC. However, selected modified activated carbons were characterized using scanning electron microscopy energy dispersive spectroscopy (SEM-EDS) and surface characterization confirmed that the activated carbon treated with thionyl chloride (AC2) contains small amount of Cl, but after the treatment with N-containing precursors the amount is minimized and become negligible. In addition, the ZVI catalyst was calcined and reduced at 400°C and at this temperature all the Cl that still remained in the carbon surface was released.
- (d) The thionyl chloride activation is required just for a comparison purpose, in order to see whether AC1 (oxidized with HNO₃) or AC2 (activated by SOCl₂) can be functionalized more efficiently with the N-functional groups on the surface. As we discussed in section 2.2.3, in the direct method, using AC1, the N-functional groups condense with the carboxyl group on the surface, while in the case of AC2 the N-functional groups are anchored on the surface via a linkage agent (SOCl₂). In both approaches the N-functional groups were successfully attached with a similar amount of N-content, except melamine.

Based on the reviewer's comment, we added a sentence in the last paragraph of section 3.1.3.1 as: "...whereas, in the case of AC2, the N-functional groups were better introduced on the carbon surface via SOCl₂ linkage".

6. Why was the ultrasonic mixing used for the Fe catalyst preparation by impregnation?, can it be expected differences from mechanical mixing?

In previous works carried out in the research group, we observed that metal supported catalysts present higher dispersions when ultrasonic mixing is used, although it is not a critical factor.

7. Page 12: "These oxygen containing groups possibly block the entry of N₂ inside the small pores [30] or the carbon structure of activated carbon is partially destroyed and some micropores are vanished by treatment with HNO₃ [24]. " According to the results given in Table 3, I do not understand this comment. AC1 have the same Vmicro as AC0, but the Smeso diminishes, therefore, one can deduce that HNO₃ treatment apparently affects the mesostructure but no the microstructure, do you agree?

We agree with the comments and we re-wrote the previous sentence as: "In spite of having similar micropore volumes, AC0 and AC1 present different mesopore areas, suggesting that the mesostructure of activated carbon was partially destroyed after treatment with HNO_3 ". (Refer section 3.1.1)

8. Melamine also affects the mesostructure according to the Smeso values.

This observation is included in section 3.1.1.

9. Explanation about the textural properties of the urea-supports is missed, why?

We have included explanation about the textural properties of the urea-supports in section 3.1.1 line 312-314 of the manuscript.

10. Why do N-containing surface groups promote the adsorption of phenol?

The presence of N-containing groups on the surface, increases the electronic density, and therefore the basicity of samples (see pH_{PZC} in Table 4), which generally favors adsorption of aromatic compounds. We included the reason and two recent references in section 3.2.1, paragraph 5.

11. By using 150 mg/L of initial phenol concentration and a mass phenol/catalyst ratio of 0.15 the results obtained in the CWPO with modified activated carbons is expected (C.M. Dominguez et al., 2013, Appl. Catal. B Environ. Page 663-670). I guess that in spite of the disappearance of phenol by adsorption, H₂O₂ was completed consumed. Some comments should be given on this respect.

We did not measure the H_2O_2 concentration. Nevertheless, we included some comments in the second paragraph of section 3.2.2.1. "According to Dominguez et al [49], as the initial phenol concentration and the phenol/catalyst ratio used in the present work are relatively low, a large fraction of active sites on the carbon surface is available for hydrogen peroxide decomposition into hydroxyl radicals, which may be consumed by non-effective reactions. This probably explains the poor performance of activated carbon in CWPO compared to adsorption."

12. From the graphical abstract, it is not possible to deduce the application of the prepared Fe-catalysts in CWPO process.

Changed as requested.

Reviewer #3:

First of all, we would like to thank the reviewer for his /her time in critically reviewing the manuscript and for his/her valuable suggestions for the improvement of the manuscript. We are so grateful and very much appreciate that. We have carefully taken them into consideration in preparing our response. We do hope that we properly address the main doubts and comments of the reviewer in this report and in the revised manuscript. Below are our responses to comments and suggestions of the reviewer. Each comment of the reviewer is listed in italic style followed by our responses.

Minor revision

Graphical abstract

I think it is a good presentation of the prepared AC materials, but the activation with thionyl chloride and subsequent nitrogen-containing AC2 and iron-containing nitrogen modified AC2 materials should be also included.

Changed as requested. As it was suggested by other reviewer, we also included the application of the prepared Fe-catalysts in the CWPO process.

Introduction:

It is not clearly justified what are the benefits of incorporating nitrogen containing groups into the carbon surface. Are these functional groups especially active as supports in wet peroxide oxidation or wet air oxidation? Are these functional groups especially interesting for the anchoring of iron species? These issues should be mentioned including also the state of the art for both of them.

It is right that the benefit of incorporating nitrogen containing groups into the carbon surface is not explicitly stated or discussed in the introduction of the previous manuscript. Some text was added to the 5th paragraph of the introduction section, in order to justify the interest of N-containing groups on carbon surface.

Major revision

Result and discussion

Textural properties.

In order to make a better discussion of the catalytic performance of Fe-AC materials, textural properties of all nitrogen-containing activated carbons after iron impregnation should be included. As it is discussed in the paper for nitrogen-containing activated carbons, the textural properties had a crucial role in the adsorption of phenol. I consider that this information is crucial for the evaluation of its catalytic performance. On the other hand, I also recommend giving Smicro instead of Vmicro.

In previous works, we observed that for small amounts of metals supported on activated carbon the textural properties of the catalysts remained practically unchanged compared to the unloaded carbon. In the preparation of the ZVI catalysts we have used 3% (wt%) of iron, and therefore we can reasonably assume that the textural properties of the supported metal catalysts are not significantly different from those of the activated carbon. In fact, it can be observed in Table 3 that the textural properties of the ZVI supported on the original AC are similar to the unloaded carbon. This information was included in section 3.1.1 of the manuscript.

On the other hand, we prefer to maintain Vmicro, since this parameter gives more information about the microporosity of the activated carbons than Smicro.

CWPO experiments with iron supported nitrogen-containing ACs.

In my opinion the characterization of all nitrogen-containing activated carbons after iron impregnations is crucial for a deeper discussion of results and better interpretation of the catalytic behavior of each Fe-AC material. The study of the adsorption and CWPO of nitrogen-containing activated carbons as supports for phenol removal was very rigorous and well-detailed according to the presented characterization data. Unfortunately, these materials showed a negligible catalytic performance in the operation conditions of CWPO that they were tested. In terms of adsorption is undoubtedly that the different strategies of incorporating nitrogen functional groups drastically affect their adsorption properties. However, I consider that the most important objective of this work is the evaluation of ZVI supported on nitrogencontaining activated carbons for CWPO of phenol, and this issue was briefly discussed in the last point 6.3.2.2 of CWPO experiments for iron supported catalysts. As characterization data of Fe-AC are not presented (at least textural properties should be given), the discussion is very general, and in some cases, hypothesis are given to support the results. It was only mentioned that impregnation does not change significantly the textural properties of the original AC material (ACO), but the resultant nitrogen-functionalized ACs could suffer some changes in its textural properties or even in the elemental composition due to the thermal reduction treatment that it is needed for the generation of ZVI. By this reason, textural properties of the materials after impregnation should be included. This information also allowed elucidating the influence of the different nitrogen functional groups on the effective dispersion of iron and its catalytic activity in CWPO. This point is also vaguely discussed in the manuscript.

As explained previously, the textural properties of the ZVI catalysts supported on the activated carbon are not significantly different from the unloaded carbon, and therefore we included in the manuscript only the textural properties of one of the ZVI catalysts to illustrate this fact. All the nitrogen-functionalized ACs were heat treated at 450°C after the functionalization and were characterized only after this step. After the impregnation with iron, the samples were reduced at even lower temperature (400°C). Thus, significant changes both in textural properties and elemental composition are neither expected not probable.

ZVI acts as a heterogeneous catalyst for the degradation of phenol in aqueous solution by effectively decomposing hydrogen peroxide and generating hydroxyl radicals, which react with phenol and its intermediates. This information was included in section 3.2.2.2.

1	Zero-valent iron supported on nitrogen-containing activated carbon for catalytic
2	wet peroxide oxidation of phenol
3	
4	S.A. Messele ^a , O.S.G.P. Soares ^b , J.J.M. Órfão ^b , F. Stüber ^a , C. Bengoa ^a , A. Fortuny ^c , A.
5	Fabregat ^a , J. Font ^{a*}
6	
7	^a Departament d'Enginyeria Química, Universitat Rovira i Virgili, Av. Països Catalans
8	26, 43007 Tarragona, Catalunya, Spain
9	^b Laboratório de Catálise e Materiais (LCM), Laboratório Associado LSRE/LCM,
10	Departamento de Engenharia Química, Faculdade de Engenharia, Universidade do
11	Porto, Rua Dr. Roberto Frias, 4200-465 Porto, Portugal
12	^c Departament d'Enginyeria Química, Universitat Politècnica de Catalunya, EUPVG,
13	Av. Víctor Balaguer, s/n, 08800 Vilanova i la Geltrú, Catalunya, Spain
14	(*) corresponding author: jose.font@urv.cat , Tel: 34977558567, Fax: 34977559621
15	
16	Abstract
17	Zero-valent iron supported catalysts were prepared through modifying an
18	activated carbon (AC) support with different nitrogen containing precursors;
19	(ethylenediamine, urea and melamine) and impregnating it with 3 wt% of iron. The
20	supports were characterized by N_2 adsorption at -196 °C, elemental analysis (EA), the
21	pH at the point of zero charge (pH_{PZC}) and temperature programmed desorption (TPD).
22	The iron catalysts were also characterized by temperature programmed reduction (TPR).
23	Subsequently, the catalysts were tested in the adsorption and wet peroxidation of
24	phenol. The results from the different characterization techniques demonstrate that the
25	nitrogen-containing groups are successfully introduced into the carbon surface via all

26	the precursors used. The tests of the different modified carbons as adsorbents/catalysts
27	indicated that the adsorption capacity and the efficiency in phenol oxidation are
28	governed by the specific surface area and functional groups present. Both surface
29	chemistry and textural properties of carbons are influenced by the nitrogen source and
30	the type of oxygen functionalities preexisting on the surface. The modified carbons
31	supported iron catalysts revealed significantly enhanced phenol removal efficiency,
32	reaching over 85% conversion after 3 h, and showed interesting catalytic stability.
33	
34	Keywords: Phenol; Activated carbon; Surface chemistry; Catalytic wet peroxide
35	oxidation
36	
37	1. Introduction
37 38	1. Introduction
37 38 39	1. Introduction The ever increasing demand for water has caused considerable attention to be
 37 38 39 40 	1. Introduction The ever increasing demand for water has caused considerable attention to be focused towards recovery and re-use of wastewaters [1]. Besides, due to the increasing
 37 38 39 40 41 	1. Introduction The ever increasing demand for water has caused considerable attention to be focused towards recovery and re-use of wastewaters [1]. Besides, due to the increasing complexity and toxicity of organic pollutants in industrial wastewaters and the more
 37 38 39 40 41 42 	1. Introduction The ever increasing demand for water has caused considerable attention to be focused towards recovery and re-use of wastewaters [1]. Besides, due to the increasing complexity and toxicity of organic pollutants in industrial wastewaters and the more increasingly strict environmental regulations, the toxicity nature must be reduced.
 37 38 39 40 41 42 43 	1. Introduction The ever increasing demand for water has caused considerable attention to be focused towards recovery and re-use of wastewaters [1]. Besides, due to the increasing complexity and toxicity of organic pollutants in industrial wastewaters and the more increasingly strict environmental regulations, the toxicity nature must be reduced. Phenol is among the most common water pollutants [2, 3] because it is presented in the
 37 38 39 40 41 42 43 44 	1. Introduction The ever increasing demand for water has caused considerable attention to be focused towards recovery and re-use of wastewaters [1]. Besides, due to the increasing complexity and toxicity of organic pollutants in industrial wastewaters and the more increasingly strict environmental regulations, the toxicity nature must be reduced. Phenol is among the most common water pollutants [2, 3] because it is presented in the effluent of numerous industrial processes such as oil refineries, petrochemical and
 37 38 39 40 41 42 43 44 45 	1. Introduction The ever increasing demand for water has caused considerable attention to be focused towards recovery and re-use of wastewaters [1]. Besides, due to the increasing complexity and toxicity of organic pollutants in industrial wastewaters and the more increasingly strict environmental regulations, the toxicity nature must be reduced. Phenol is among the most common water pollutants [2, 3] because it is presented in the effluent of numerous industrial processes such as oil refineries, petrochemical and pharmaceutical industries [4, 5]. This draws the attention of policymakers and scientists
 37 38 39 40 41 42 43 44 45 46 	1. Introduction The ever increasing demand for water has caused considerable attention to be focused towards recovery and re-use of wastewaters [1]. Besides, due to the increasing complexity and toxicity of organic pollutants in industrial wastewaters and the more increasingly strict environmental regulations, the toxicity nature must be reduced. Phenol is among the most common water pollutants [2, 3] because it is presented in the effluent of numerous industrial processes such as oil refineries, petrochemical and pharmaceutical industries [4, 5]. This draws the attention of policymakers and scientists to take the necessary measures by studying alternative technologies, such as Advanced

48 AOPs are technologies based on the action of hydroxyl and other radicals to 49 oxidize recalcitrant, toxic and non-biodegradable compounds towards by-products and 50 eventually inert end products [6, 7]. Among others, hydrogen peroxide, ozone or Fenton process achieves good results so far. However, its application is mainly dependent on the narrow operational pH range, the loss of reagent activity and the need for a subsequent separation step to remove the homogeneous catalyst from the effluent, which significantly increases the cost of the operation [12]. Therefore, taking into consideration the aforementioned inconveniences of Fenton system, a strategy is proposed to prepare a heterogeneous catalyst containing both the catalytic metal (iron) and N-functional groups on the surface of activated carbon (AC).

62 On one hand, it is well known that AC is widely used as a good adsorbent and 63 supporting material due to its excellent properties in mechanical strength and porous 64 structures [13, 14]. On the other hand, the catalytic activity of AC is far from well 65 understood. Besides the direct relationship with the physical properties (surface area, 66 pore volume, etc.) of activated carbon [15], the surface chemistry can play an important 67 role [16-18]. According to the recent work of Dominguez et al. [19], the effects of physicochemical characteristics of carbon materials on the rate of hydrogen peroxide 68 69 decomposition were explained using cyclic voltammetry, and the main results indicated 70 that the most important factor in the catalytic activity was the content of metals, in 71 particular iron, this was followed by the specific surface area and finally the content of 72 surface oxygen groups.

In addition, the presence of a chelating agent in homogeneous Fenton has demonstrated to enhance the oxidation capacity [20], either by powering the oxidation potential or keeping the iron in solution at higher pH, although the addition of a new

76	compound in solution is clearly an inconvenience [21]. Although, some interest has
77	been showed in using nitrogen-containing carbon materials as metal catalyst supports
78	[22], to the best of our knowledge, there are no data about their application in the
79	catalytic wet peroxide oxidation (CWPO) process, but there are some recent studies on
80	catalytic wet air oxidation (CWAO). Aminated activated carbon as a catalyst in the
81	CWAO of cooking wastewater has been studied by Chen et al. [23]. The COD removal
82	in CWAO of organic compounds by nitrogen containing AC was higher than that of the
83	untreated AC and suggested that nitrogen-containing functional groups contributed to
84	the enhanced activities of ACs. A similar study was reported by Ayusheev et al. [24],
85	focusing on the effect of nitrogen content in N-doped carbon nanofibers (N-CNFs) on
86	the catalytic activity of Ru/N-CNFs in the wet air oxidation of phenol. Ru-containing
87	catalysts and nitrogen in N-CNFs was found to be responsible for both the increased
88	activity and stability of the catalysts. Moreover, this study illustrates that the increase in
89	catalytic activity is related with several factors: (1) introducing a heteroatom into carbon
90	materials changes the acid-base properties of the support surface; (2) it makes possible
91	to control the size of the supported metal particles; (3) using a support of higher
92	conductivity leads to enhanced chemical reactivity for electron transfer process in a
93	catalytic system.
94	Therefore, the main objective of this work is to prepare surface modified
95	activated carbon materials (for utilisation as supports/catalysts) using different N-
96	containing groups, i.e. ethylenediamine (EDA), urea or melamine. Prior to introducing
97	the N-functional groups on the AC surface, the activated carbon was oxidized by HNO_3
98	and/or subsequently activated by thionyl chloride treatment. Later, iron impregnated and
99	then their performances where evaluated on phenol adsorption and wet peroxide

100 oxidation.

102

103 2.1. Materials

104

105 Activated carbon Norit ROX 0.8 was used as starting material for further 106 treatments. Phenol (99%) and sodium chloride (99%) were purchased from Panreac. 107 Nitric acid ($\geq 65\%$), hydrochloric acid ($\geq 37\%$), hydrogen peroxide (30% wt), urea (98%), 108 sodium hydroxide (≥97%) and iron nitrate nonahydrate (>98%) were purchased from 109 Sigma-Aldrich. Ethylenediamine (>99%) was obtained from Merck. Thionyl chloride 110 (>99%), melamine (≥99%) and toluene (>99%) were purchased from Fluka. The main 111 physico-chemical properties of these N-containing compounds are presented in Table 1. 112 All chemicals were used as received without further purification. Deionised water was 113 used throughout the work.

114

115 2.2. Modification of activated carbon supports and catalyst preparation

116

Prior to use, the Norit ROX 0.8 activated carbon was crushed and granules of 25 -50 mesh size (0.3 – 0.7 mm) were separated and used as starting material (Sample ACO). Activated carbons were modified using a combination of surface modification protocols described below, in order to obtain supports having N-containing surface groups. Iron was then supported on these modified activated carbons by incipient wetness impregnation.

123

124 2.2.1. Oxidation with nitric acid in liquid phase

125

126	The starting activated carbon (sample AC0) was oxidized with HNO_3 using a
127	125 mL Soxhlet extraction apparatus containing 9 g of activated carbon, connected to a
128	boiling flask and to a condenser. A volume of 250 mL of 6 M HNO3 was introduced
129	into a 500 mL Pyrex round-bottom flask and heated to boiling temperature with a
130	heating mantle. The reflux was stopped after 3 h. The oxidized activated carbon was
131	extensively washed with distilled water to neutral pH and then dried in an oven at 110
132	°C for 24 h (sample AC1) [18].

133

134 2.2.2. Activation with thionyl chloride

135

Thionyl chloride was used as a linking agent on the surface of AC for the attachment of N-functional groups [25]. About 12 g of the oxidized activated carbon (AC1) was subsequently activated with 40 mL of 5% solution of thionyl chloride in toluene for 5 h at 70 °C. The carbon was then rinsed at least two times with toluene, and then purified by Soxhlet extraction with toluene for 2 h, and dried in an oven at 110 °C for 24 h (sample AC2) [26-28].

142

143 2.2.3. Functionalization with ethylenediamine, urea and melamine

144

Samples AC1 and AC2 were the starting materials for functionalization with Ncontaining compounds. The functionalization of activated carbons with EDA, urea or melamine can be conducted directly (using sample AC1) or indirectly (using sample AC2). In the direct method, amine groups of these compounds condense with carboxyl groups on AC to generate surface amide groups. In the indirect method, they are anchored on the surface via a linkage agent thionyl chloride (SOC1₂). In this case, first the linking agent reacts with a carboxylic group of the surface to convert the surface carboxylic groups into acyl chloride functionalities [25]; and in the second stage, the amine groups of EDA/urea/melamine condense with anchored chlorine atoms [29].

154

Ethylenediamine: 2 g of AC1 (or AC2) were refluxed with 100 mL solution of EDA in toluene (1 M) for 24 h. The amine grafted samples were washed in toluene under sonication for 10 min and then were purified in a Soxhlet extraction unit for 2 h to remove unattached or free EDA from activated carbon. The resulting activated carbons were dried in an oven at 110 °C for 24 h (samples AC1_EDA and AC2_EDA).

160

161 *Urea*: 2 g of AC1 (or AC2) were added into 100 mL of an aqueous urea solution (1 M), 162 and stirred at room temperature for 24 h. Then, the material was filtered and dried in the 163 oven. The sample treated with urea was carbonized under N_2 flow (100 cm³ min⁻¹) at 10 164 °C min⁻¹ up to 450 °C and held at this temperature for 50 min (samples AC1_Urea and 165 AC2_Urea)

166

167 *Melamine*: 2 g of AC1 (or AC2) were mixed with a melamine suspension (1.3 g of 168 melamine in 100 mL of 80% ethanol) and stirred at 70 °C for 5 h. Then the mixture was 169 boiled to evaporate the solvent and the slurry was dried at 110 °C for 24 h. The sample 170 impregnated with melamine was carbonized under N₂ flow (100 cm³ min⁻¹) at 10 °C 171 min⁻¹ up to 450 °C and held at this temperature for 50 min [30, 31]. After this treatment 172 the samples are labeled as AC1_Melamine and AC2_Melamine, respectively.

173

174 2.2.4. Preparation of catalysts

175

176 The active metal was supported on the original (AC0) and in all the modified 177 activated carbons by incipient wetness impregnation of an aqueous solution of 178 iron(III) nitrate nonahydrate. The impregnation was always conducted under vacuum 179 and ultrasonic mixing. The precursor solutions with calculated concentration were 180 added drop wise using a peristaltic pump and the slurry was left at room temperature 181 under ultrasonic mixing for 90 min. After impregnation, the samples were dried at 110 182 °C for 24 h and calcined under a nitrogen flow at 400 °C for 1 h, then finally reduced at 183 400 °C in hydrogen flow for 3 h (sample ZVI/support). The content of metal was 184 maintained constant at 3 wt%. The reduction temperature for the iron catalysts was 185 determined by temperature programmed reduction (TPR). The treatment methods for 186 the activated carbon used in this work are summarized in Table 2.

187

188 2.3. Characterization of supports and catalysts

189

190 The supports were characterized by N_2 adsorption at -196 °C, temperature 191 programmed desorption (TPD), elemental analysis and determination of the pH at the 192 point of zero charge (pH_{PZC}). The catalysts were also characterized by temperature 193 programmed reduction (TPR).

194

195 2.3.1. Textural characterization

196

197 The textural characterization of all the materials was checked in order to 198 evaluate if there had been significant textural changes after the modification of surface 199 chemistry. This characterization was based on the N_2 adsorption isotherms, determined 200 at -196 °C with a Quanthachrome NOVA 4200e multi-station instrument. Prior to the 201 measurements, the samples were outgassed at 120 °C for 5 h under vacuum. The specific surface area of the mesopores (S_{meso}) and the micropore volume (V_{micro}) were calculated by the t-method. Moreover, the surface area (S_{BET}) of the samples was calculated by the B.E.T. method.

205

206 2.3.2. Temperature programmed reduction (TPR)

207

208 Temperature programmed reduction (TPR) analysis allows finding the most 209 appropriate reduction temperature of the metal and to evaluate the effect of modified 210 activated carbons on the reducibility. TPR profiles were obtained with a fully automated 211 AMI-200 (Altamira Instruments). The catalyst (0.15 g) was placed in a U-shaped quartz tube inside an electrical furnace and heated at 5 °C min⁻¹ up to 600 °C under a flow of 212 5% (v/v) H₂ in Ar at 30 cm³ min⁻¹. The H₂ consumption was monitored by a thermal 213 214 conductivity detector (TCD). The temperature range where reduction occurs could be 215 indicated directly from the H₂ consumption peaks.

216

217 2.3.3. Surface chemistry characterization

218

219 The surface chemistry of the starting and modified activated carbons was 220 characterized by temperature-programmed desorption (TPD) [18]. The TPD spectra of 221 CO and CO₂ were obtained with a fully automated AMI-300 (Altamira instruments). 222 The carbon sample (0.10 g) was placed in a U-shaped quartz tube inside an electrical furnace and heated at 5 °C min⁻¹ up to 1100 °C under a constant flow rate of He at 25 223 cm³ min⁻¹ (STP). The amounts of CO (m/z = 28) and CO₂ (m/z = 44) released from the 224 carbon samples were monitored with a mass spectrometer (Dymaxion, Ametek). CO 225 226 and CO_2 were calibrated at the end of each analysis.

227 The pH at the point of zero charge (pH_{PZC}) was determined by mixing 0.05 g of 228 each sample with 20 mL of 0.01 M NaCl solution with pH values adjusted between 2 229 and 11, by adding 0.1 M HCl or 0.1 M NaOH solutions. The final pH was measured 230 after 48 h of shaking at room temperature. Blank experiments (without addition of 231 carbon) were also performed for each pH and the values measured after 48 h are 232 considered as the initial pH, in order to avoid the variation of pH caused by the effect of 233 CO_2 present in head space. The pH_{PZC} value of each carbon sample was determined by 234 intercepting the obtained final pH vs. initial pH curve with the straight line final pH =235 initial pH [32].

Carbon, hydrogen, nitrogen and oxygen (by difference) contents were
determined using a Carlo Erba EA 1108 Elemental Analyzer. Prior to the analysis,
water adsorbed on the surface of the supports was removed by drying at 110 °C in the
oven overnight.

240

241 2.4. Adsorption and Catalytic tests

242

Both adsorption and CWPO experiments were carried out in a magnetically stirred tank (batch reactor). The reactor was filled with 100 mL of a phenol aqueous solution (150 mg/L) and heated by immersion in a water bath at controlled temperature.

The solution pH was adjusted to a value of 3.0 using H_2SO_4 solution (1 M). When the desired temperature was reached (30 °C), a calculated volume of H_2O_2 (0.25 mL) was added into the system after catalyst addition (100 mg), in order to reach a H_2O_2 concentration of 750 mg/L in the CWPO runs (the theoretical stoichiometric amount needed to completely mineralize phenol); at this moment the reaction was assumed to start. The CWPO oxidation runs were carried out both using the activated carbons (without iron) and iron supported carbon as catalysts, in order to evaluate the ability of the supports alone to decompose H_2O_2 , and subsequently to promote phenol oxidation. This may give an idea to discriminate between a potential catalytic surface chemistry and the catalytic activity owing to the presence of iron.

In addition, pure adsorption runs were performed in order to assess the contribution of adsorption to the removal of phenol. The experiments were carried out in the same conditions, but without H_2O_2 addition.

259 Selected experiments were performed in triplicate and the relative error of the 260 experimental results was below $\pm 4\%$.

261

262 2.5. Analytical methods

263

Both in pure adsorption and in CWPO, liquid samples were periodically
withdrawn from the reactor. Then, each sample was filtered with a syringe filter of 0.20
μm nylon (Teknokroma, ref._TR-200101) and placed in a glass vial (Agilent) for
immediate analysis.

268 The determination of phenol concentration was performed by High Performance Liquid Chromatograph (HPLC, model 1220 Infinity LC, Agilent Technologies) 269 270 equipped with UV detector and C18 reverse phase column (Hypersil ODS, 5 µm, 25 x 271 0.4 cm from Agilent Technologies). The analyses were carried out with a mobile phase 272 (flow rate 1 mL/min) of a 40/60% mixture of methanol and ultrapure water (Milli-Q 273 water). The pH of the water was adjusted at 1.41 with sulphuric acid (H_2SO_4). The 274 detection was performed by UV absorbance at a wavelength of 254 nm. The automatic 275 injection volume was 20 µL.

276 Total Organic Carbon (TOC) was also determined at the end of the experiments

11

in a TC Multi Analyzer 2100 N/C equipment from Analytic Jena with a non-diffractiveIR detector.

Fe leached to the reaction media was determined by using Atomic Absorption
Spectroscopy (AAS) at 249 nm.

The main parameters used to compare the results in the discussion section are the removal of phenol (X_{PhOH}) and TOC abatement (X_{TOC}), which are respectively defined as:

284
$$X_{PhOH}(\%) = \frac{[PhOH]_0 - [PhOH]_t}{[PhOH]_0} \times 100$$
 (2)

where $[PhOH]_0$ is the initial phenol concentration and $[PhOH]_t$ is the concentration at time t, and

$$X_{TOC}(\%) = \frac{TOC_0 - TOC_t}{TOC_0} \times 100$$
(3)

where TOC_0 is the initial TOC concentration and TOC_t is the concentration at time t.

289

290 **3. Results and discussion**

291

292 3.1. Characterization of activated carbons and catalysts

293 *3.1.1. Textural properties*

BET surface area (S_{BET}), mesopore surface area (S_{meso}) and micropore volume (V_{micro}) of the original AC, modified AC and one iron supported catalyst are summarized in Table 3. The results revealed that the original AC (AC0) is highly microporous and has high BET surface area, 984 m²/g. In the case of acid-treated (AC1) and HNO₃ plus SOCl₂-treated (AC2) activated carbons, the BET surface area decreased following the treatments by 13% and 35%, respectively. The slight change in the surface area of AC1 can be explained by the abundant presence of oxygen-containing 301 groups on the surface of the activated carbon, which are introduced by the HNO_3 302 treatment. In spite of having similar micropore volumes, AC0 and AC1 present different mesopore areas, suggesting that the mesostructure of activated carbon was partially 303 304 destroyed after treatment with HNO₃. In the case of AC2, which is treated with HNO₃ 305 and SOCl₂, the reduction in the surface area can be explained by the introduction of 306 both oxygen and sulfur-containing groups. Such sulfur moieties (sulfones, sulfides, 307 sulfoxides, and sulfur atoms) may lead to active sites occlusion inside the pores and on 308 the surface of AC, as reported by Khayoon et al. [33]. Treatments with EDA and 309 melamine lead to a drastic reduction of the surface area. This may be due to the 310 presence of numerous groups on the activated carbon surface, which may partially block 311 the access of N_2 molecules to the micropores. However, an appreciable increase of 312 surface area and pore volume can be noticed only for urea treated samples (compared to 313 AC1 and AC2). In general, the textural properties of carbons are highly influenced by 314 the nitrogen containing precursor; specially, for the AC2_Melamine sample, for which 315 the micropore volume decreased up to nearly zero and also affects the mesostructure 316 according to the S_{meso} value. On the other hand, the impregnation of iron did not 317 substantially change the textural properties of the original carbon material due to the 318 small load of iron (3 wt%) used, which is in agreement with the previous literature [34].

319

320 *3.1.2. TPR*

TPR profiles of iron catalyst on the different supports are shown in Fig. 1. For comparison purpose, the TPR analyses of selected modified activated carbons were also performed and no additional peaks were observed. As it can be seen from Fig. 1, the position, width and intensity of the peaks mainly depend on the nature of the support. The major reduction peak is at a temperature around 400 °C for original (AC0) and 326 oxidized AC (AC1). The second peak observed at temperature above 550 °C in AC0 327 profile could be related to iron oxide which is strongly interacting with the support. In 328 general, the reduction peak for samples treated with N-containing compounds is shifted 329 to the right irrespective of the starting material (AC1 or AC2). The reduction 330 temperature depends on the degree of interaction between the active species and the 331 support. For all samples treated with N-containing compounds using either AC1 or AC2 332 as a starting material, the iron reduction peaks shift to higher temperatures. Exceptions 333 are the samples treated with urea, which present a reduction peak at lower temperature.

334

335 *3.1.3. Surface chemistry characterization*

336 *3.1.3.1. TPD* and pH_{PZC}

337 TPD analyses were carried out to evaluate the surface chemistry of the different 338 AC supports. The nature of the groups can be assessed by the decomposition 339 temperature and the type of gas released. CO and CO₂ are released by thermal 340 decomposition of oxygen containing groups on the surface of carbon materials at 341 different temperatures [18]. The TPD profiles of the original and modified activated 342 carbons are depicted in Fig. 2. From the TPD profiles, it is possible to identify and 343 quantify the amounts of the oxygenated groups of each sample. CO_2 peaks result from 344 carboxylic acids at low temperatures (150 - 450 °C), or lactones at higher temperatures 345 (600 - 800 °C); carboxylic anhydrides originate both CO and CO₂ (400 - 650 °C); 346 groups such as phenols (600 - 800 °C) and carbonyls/quinones (750 - 1000 °C) 347 originate CO peaks [35].

348 It can be seen that the nitric acid treatment (AC1) increase the amount of 349 oxygenated surface groups, which is evidenced by the increase of CO₂ (Fig. 2a) and CO

14

350 (Fig. 2b) released. The samples treated with EDA originate CO peaks at lower
351 temperatures (200-400 °C) than the others.

The total amounts of CO and CO_2 released from the AC materials were calculated from the corresponding TPD spectra. This information as well as the point of zero charge of the samples with different surface chemistries are shown in Table 4.

355 As expected, AC1 has a low pH_{PZC} and the highest amount of CO and CO₂, due 356 to the introduction of oxygen-containing surface groups having acidic properties, 357 mainly carboxylic acids. After thionyl chloride treatment (sample AC2), the carboxyl 358 groups were mostly converted into acid chloride groups, resulting a decrease of the 359 amount of CO and CO₂, but this material (AC2) still has acidic properties with the 360 corresponding pH_{PZC} value similar to that of sample AC1. With the exception of 361 AC2_EDA, all materials obtained after the treatment with N-containing precursors 362 using sample AC1 or AC2 as starting materials have approximately neutral or slightly 363 basic properties. This can be explained by the presence of nitrogen groups having basic 364 properties.

In general, the TPD profile shows that it is important to have high density of oxygenated groups on the surface of carbon before functionalization with the Ncontaining compounds. The N-containing groups (EDA, urea and melamine) strongly interact with carboxylic acids, anhydrides and lactones [28, 36]. Whereas, in the case of AC2, the N-functional groups were effectively introduced on the carbon surface via SOCl₂ linkage.

371

372 *3.1.3.2. Elemental analysis*

The carbon, hydrogen, nitrogen and oxygen contents obtained by elemental analysis are summarized in Table 5. Considering the treatment carried out, it was 375 expected that AC2 and AC2-based samples may contain some sulphur, which is376 summed with the oxygen content as oxygen determined by difference.

The original AC contains a small amount (about 1%) of nitrogen. Significant amounts of nitrogen were introduced on the surface after treatment with N-containing compounds. The highest amount of nitrogen incorporated into the structure was obtained when melamine was the source of nitrogen, as its molecule contains up to six nitrogen atoms.

382 Thus, the differences in the amounts of nitrogen may be related to the content of 383 nitrogen in the precursors (67% in melamine; 47% in urea and EDA) and the promoting 384 effect of surface acidity, enhanced by oxidation/activation of SOCl₂ on the retention of 385 N-containing organic bases [31,37]. On the other hand, the EDA treated samples 386 contain a relatively high content of hydrogen compared to others; this is probably due to 387 the higher content of hydrogen found in the compound. It is interesting to note that in 388 the urea and EDA treated samples similar total contents of nitrogen are found regardless 389 the pretreatment applied, either using oxidized (AC1) or activated with thionyl chloride 390 (AC2). However, in the case of melamine treated samples, the nitrogen content is higher 391 when the AC2 sample is the starting material in comparison to the inactivated (only 392 oxidized) sample AC1.

393

394 *3.2. Catalytic activity*

Although it is difficult to compare the performance of catalysts having different textural properties and surface chemistries, we performed some tests using the activated carbons as adsorbents or catalysts and the iron supported on modified AC as catalysts for the removal of phenol in solution.

16

The phenol removal on the original and modified activated carbons was studied as a function of time under the following conditions: 100 mL of solution having 150 mg/L of phenol, adsorbent/catalyst amount = 1 g/L, initial pH of the solution = 3.0, T = $30 \,^{\circ}$ C, atmospheric pressure and stoichiometric amount of H₂O₂.

403

404 3.2.1. Adsorption experiments

405 Adsorption experiments were carried out to evaluate the effect of textural and 406 surface properties of the original and modified activated carbons on phenol removal by 407 adsorption. The results of phenol removal in adsorption runs performed with the 408 original and modified activated carbons are compiled in Fig. 3.

409 Even though the difference is not big compared to the original AC, it is clear that 410 urea treated samples (AC1_Urea and AC2_Urea) improve the adsorption capacity of the 411 original AC (AC0). These samples (AC0, AC1_Urea and AC2_Urea) present similar 412 surface areas and pore volumes, their difference being mainly related with their surface 413 chemistry. All the other materials showed lower adsorption capacities compared to the 414 original AC. These adsorption results confirm that the textural properties of the carbon 415 materials have a primary role in adsorption of phenol, irrespective of surface chemistry, 416 but closely following the available surface area.

In the adsorption of aromatic compounds in liquid phase on activated carbons, there are mainly two types of interactions: (1) electrostatic and (2) dispersive [38]. The first mechanism is involved when the adsorbate is dissociated under the experimental conditions. For example, if solution pH > pH_{PZC}, then the carbon surface is negatively charged, the opposite occurs when the solution pH < pH_{PZC} as described elsewhere [39]. In the case of the second mechanism, the existence of π - π dispersion interactions is commonly accepted [38, 40]. Taking this in consideration, at pH of 3.0, phenol 424 $(pK_a=9.89)$ is found in solution predominantly in the molecular form and only the 425 dispersive interactions are most probably involved in its adsorption on the carbon 426 surface [41, 42].

427 Among the materials tested, the best performance was obtained with AC2_Urea, 428 reaching 67% of phenol removal after 3 h. The BET surface area of this adsorbent is 429 slightly lower and the pH_{PZC} is higher compared to the original AC. The EDA treated 430 samples show moderate adsorption performance having moderate surface area and 431 pH_{PZC} . However, the melamine treated samples show a poor adsorption performance 432 due to the drastic decrease of surface area, though having a higher pH_{PZC} compared to 433 AC1 and AC2 (pH_{PZC} of 2.5).

434 The presence of N-containing groups on the surface, increases the electronic 435 density, and therefore the basicity of samples (see pH_{PZC} in Table 4), which generally 436 favors adsorption of aromatic compounds. Recently, ammonia-modified activated carbon was prepared for the adsorption of 2,4-dichlorophenol (2,4-DCP), which 437 438 enhances the adsorption capacity in comparison with the parent activated carbon [43]. 439 This was explained due to the basic surface functional groups created by nitrogen-440 incorporation. Similar results also presented by Yang et al [44] for phenol adsorption 441 using aminated activated carbon.

441 using aminated activated carbon.

442 Moreover, the analysis of the results shows that there is a correlation between 443 the phenol removal efficiency by adsorption and the S_{BET} of the tested activated 444 carbons. The correlation is shown in Fig. 4. The figure clearly shows that the phenol 445 removal efficiency increases linearly with the S_{BET} , irrespective of the surface 446 chemistry. This indicates that the surface area of activated carbons is the principal 447 responsible for the phenol adsorption capacity. Fierro and co-workers [45] also 448 observed that the amount of adsorbed phenol in microporous activated carbons449 increases linearly with the increase in micropore volume.

450 As mentioned above, the textural and chemical properties of the tested samples 451 are different. So, an additional study is needed to take conclusions about the effect of 452 textural properties and surface functionalities. For this purpose, the apparent rate 453 constants for the first-order (k_1) and second-order (k_2) adsorption models were 454 determined for all samples tested. The constants of the two models are listed in Table 6. 455 The ratio between k_1 or k_2 and S_{BET} is presented in the same table. Even though, both the first-order and the second-order model fit the experimental data quite well with R^2 456 457 values close to unity, the second-order model was suitable for the adsorption of lower 458 molecular weight adsorbates on smaller adsorbent particles as reported by Wu et al. [46] 459 and therefore the corresponding kinetic constants are used in the following discussion.

460 It is reported that phenol adsorption is dependent on both the surface area and 461 the presence of surface groups [45, 47]. In order to evaluate the influence of the 462 nitrogen content, the second-order apparent rate constants (k_2) are normalized by the 463 specific surface area (S_{BET}). It can be observed in Fig. 5 that high normalized adsorption 464 rate constants correspond to samples with high amounts of N-containing surface groups 465 (see Table 5).

466

467 *3.2.2. CWPO experiments*

-00

468 *3.2.2.1. Modified activated carbons as catalysts*

In order to evaluate the ability of the modified activated carbons to act as catalysts (without any supported metal) in the CWPO of phenol, runs were performed under the same conditions as the adsorption tests but now adding H_2O_2 . The corresponding phenol removal curves are shown in Fig. 6. 473 Phenol removal by CWPO in the presence of modified activated carbons as 474 catalysts was similar or slightly higher than those obtained by adsorption. However, the 475 results of phenol removal by adsorption (Fig. 3) and CWPO (Fig. 6) lead to the 476 conclusion that these modified carbon materials are not particularly active for CWPO of 477 phenol. For instance, the phenol removal at 60 min for AC1_Urea is 43% without H₂O₂ 478 and 42% in the presence of H_2O_2 , or 65% and 64%, respectively, at 180 min, which 479 suggests that both cases are governed by adsorption. So, the removal of phenol is 480 mainly due to pure adsorption. Similar conclusion was obtained in a previous work for 481 some dyes [48]. According to Dominguez et al [49], as the initial phenol concentration 482 and the phenol/catalyst ratio used in the present work are relatively low, a large fraction 483 of active sites on the carbon surface is available for hydrogen peroxide decomposition into hydroxyl radicals, which may be consumed by non-effective reactions. This 484 probably explains the poor performance of activated carbon in CWPO compared to 485 486 adsorption.

487

488 *3.2.2.2. Iron supported catalysts*

489 In order to study the influence of surface chemical characteristics of the support 490 on the activity of iron catalysts in phenol oxidation, a set of runs were performed in 491 CWPO for different ZVI supported catalysts. Fig. 7 shows the evolution of phenol 492 removal in the presence of the iron catalyst supported on different supports. 493 ZVI/AC2_Melamine and ZVI/AC2_EDA show a good and similar performance, 494 reaching above 80% phenol conversion after 60 min. On the contrary, the phenol 495 removal by pure adsorption and CWPO using the same material without the presence of 496 ZVI (i.e AC2_Melamine and AC2_EDA) shows low removal efficiency below 15 % 497 after 60 min.

498 It must be strongly noted that the presence of ZVI on N-containing catalysts 499 yields a better phenol removal, reaching values over 85% after 3 h. Here, ZVI acts as a 500 heterogeneous catalyst for the degradation of phenol in aqueous solution by effectively 501 decomposing hydrogen peroxide and generating hydroxyl radical, which reacts at high 502 rate with phenol and its intermediates [50]. These results revealed that there is a direct 503 and positive relation between the catalytic activity and the nitrogen content of the 504 materials. Therefore, the activity of iron supported catalysts based on N-containing AC 505 in CWPO of phenol is enhanced when compared to that of the untreated ones. It is thus 506 observed that the presence of nitrogen groups on the surface of activated carbon, in 507 addition to iron, clearly increases the phenol removal. This synergy can be probably 508 assigned to the ability of retaining the iron ions close to the carbon surface due to the 509 complexing properties of these N-containing groups.

510 Similar trends were observed for TOC removal at the end of adsorption and 511 oxidation tests. Fig. 8 shows that the EDA and melamine treated AC present lower TOC 512 removal performance in the adsorption or oxidation tests when these materials were 513 used as adsorbents or catalysts, respectively. However, the ZVI supported catalysts on 514 the EDA and melamine treated supports also promote the TOC removal. The main 515 reason for this can be explained by the synergetic effect of the presence of nitrogen 516 group and iron on the surface of activated carbon, promoting oxidation instead of 517 adsorption. In this case, the access of H_2O_2 to the iron on the carbon surface could be 518 easier than for the other samples and accelerates the hydroxyl radical generation for 519 deep oxidation of phenol, as observed in Fig. 7. However, additional studies are still 520 required to reveal the exact origin of these effects.

521 Leaching tests were performed in order to evaluate the stability and contribution 522 of homogeneous reaction for each catalyst after 120 and 180 min of reaction. Leaching
523 of iron corresponds to values below 0.01% of the metal initially present in almost all 524 catalysts; also limited values of 0.03 to 0.04% were measured in the case of AC1 and 525 AC2 supported catalysts. This fact can be related with the pH_{PZC} of the support. In fact, 526 AC1 and AC2 present the lowest pH_{PZC} and a relatively high iron leaching. In 527 conclusion, the contribution of homogeneous system is quite negligible, as the amount 528 of leached iron is quite low.

529 Since the stability of a catalyst is one important aspect in the general evaluation 530 of its performance, the most efficient catalysts of this study (ZVI/AC2 Melamine and 531 ZVI/AC2_EDA) were reused three times in consecutive CWPO reactions (Fig. 9). A slight loss of catalyst activity is observed with both catalysts from the 1st to the 2nd run, 532 but the activity being practically maintained from the 2nd to the 3rd run, and the actual 533 removal of phenol is around 80%. The decrease in the removal of phenol observed from 534 535 the first to the second run may be due to the lower adsorption capacity of the used 536 catalyst and/or the slight loss of the catalyst during the recovery process. In fact, at the 537 end of the experiment, a fraction of organic compounds will remain adsorbed onto the 538 catalyst under the employed operating conditions. These compounds can be either 539 phenol or intermediates formed during the oxidation reaction, as discussed elsewhere 540 [51]. In any case, both catalysts (ZVI/AC2 Melamine and ZVI/AC2 EDA) still present 541 a good efficiency under consecutive runs.

542

543 **4.** Conclusions

544

545 Activated carbons with different N-containing precursors were modified, 546 characterized and tested as adsorbents or catalysts for adsorption and peroxide oxidation of phenol. Besides, these modified carbons were also impregnated with iron and used ascatalysts for CWPO.

549 Treatments with EDA and melamine lead to a drastic surface area reduction. 550 This may be due to the presence of numerous groups on the activated carbon surface, 551 which may partially block the access of N₂ molecules to the micropores.

In the urea and EDA treated samples similar total contents of nitrogen are found regardless of the pretreatment applied. However, in the case of melamine treated samples, the nitrogen content is significantly higher when the oxidized activated carbon treated with thionyl chloride is the starting material.

556 Phenol removal by CWPO in the presence of carbon materials was slightly 557 higher than those obtained by adsorption. However, the results lead to the conclusion 558 that these materials are not particularly active for the reaction, the removal of phenol 559 being mainly due to adsorption.

The iron supported catalysts based on N-containing AC show the highest phenol removal efficiency, reaching values over 85% conversion after 3 h. This result revealed that there is a positive relation between the catalytic activity and the nitrogen content of the materials.

564

565 Acknowledgements

566

567 Financial support for this research was provided by the Spanish Ministerio de 568 Educación y Ciencia and FEDER, projects CTM2008-03338 and CTM2011-23069. The 569 Spanish Ministerio de Ciencia e Innovación is also thanked for providing a doctoral 570 scholarship (Programme FPI, BES-2009-017016) to carry out this research work. The 571 author's research group is recognised by the Comissionat per a Universitats i Recerca

572	del	DIUE de la Generalitat de Catalunya (2009SGR865) and supported by the
573	Univ	ersitat Rovira i Virgili (2010PFR-URV-B2-41). Funding from FCT and FEDER in
574	the fi	ramework of Program COMPETE (FCT grant SFRH/BPD/80435/2011 and Project
575	Pest-	C/EQB/ LA0020/2011) is acknowledged by the LCM group of authors.
576		
577	Refe	rences
578		
579	[1]	G. Annadurai, S.R. Babu, K.P.O. Mahesh, T. Murugesan, Bioprocess Eng., 2
580		(2000) 493–501.
581	[2]	P.S. Nayak, B.K. Singh, Desalination 207 (2007) 71–79.
582	[3]	C.E. Paisio, E. Agostini, P.S. González, M.L. Bertuzzi, J. Hazard. Mater. 167
583		(2009) 64–68.
584	[4]	A. Fortuny, C. Bengoa, J. Font, F. Castells, A. Fabregat, Catal. Today 53 (1999)
585		107–114.
586	[5]	A. Santos, P. Yustos, A. Quintanilla, S. Rodríguez, F. García-Ochoa, Appl. Catal.
587		B 39 (2002) 97–113.
588	[6]	C. Comninellis, A. Kapalka, S. Malato, S.A. Parsons, I. Poulios, D. Mantzavinos,
589		J Chem Technol Biotechnol, 83 (2008) 769–776.
590	[7]	D. Mantzavinos, E. Psillakis, J Chem Technol Biotechnol, 79 (2004) 431–454.
591	[8]	V. García-Molina, M. López-Arias, M. Florczyk, E. Chamarro, S. Esplugas, Water
592		Res. 39 (2005) 795–802.
593	[9]	F. Luck, Catal. Today 53 (1999) 81–91.
594	[10]	H. Debellefontaine, S. Crispel, P. Reilhac, F. Périé, J. Foussard, Chem. Eng. Sci.
595		54 (1999) 4953-4959.

- 596 [11] M. Klavarioti, D. Mantzavinos, D. Kassinos, Environ. Int. 35 (2009) 402–417.
- 597 [12] J.J. Pignatello, E.E. Oliveros, A. Mackay, Environ. Sci. Technol. 36 (2006) 1–84.
- 598 [13] E. Ahumada, H. Lizama, F. Orellana, C. Suárez, A. Huidobro, A. Sepúlveda599 Escribano, F. Rodríguez-Reinoso, Carbon 40 (2002) 2827–2834.
- 600 [14] F. Rodríguez-Reinoso, Carbon, 36 (1998) 159–175.
- 601 [15] M. Eugeni, F. Stüber, A. Fortuny, A. Fabregat, J. Carrera, J. Font, Appl. Catal. B:
 602 Environ. 58 (2005) 105–114.
- 603 [16] F. Stüber, J. Font, A. Fortuny, C. Bengoa, A. Eftaxias, and A. Fabregat, Top
 604 Catal., 33 (2005) 3–50.
- 605 [17] A. Quintanilla, J.A.Casas., J.J. Rodriguez, M.T. Kreutzer, F. Kapteijn, J.A.
 606 Moulijn, Int. J. Chem. Reactor Eng. 5 (2007) A62.
- 607 [18] J.L. Figueiredo, M.F.R. Pereira, M.M.A. Freitas, J.J.M. Órfão, Carbon 37 (1999)
 608 1379–1389.
- 609 [19] C.M. Domínguez, A. Quintanilla, P. Ocón, J.A. Casas, J.J. Rodriguez, Carbon 60
 610 (2013) 76–83.
- 611 [20] N. Klamerth, S. Malato, A. Agüera, A. Fernández-Alba, Water Res. 47 (2013)
 612 833–840.
- 613 [21] Z. Yuan, J.M. VanBriesen, Environ. Eng. Sci. 23 (2006) 533–544.
- 614 [22] L. Mabena, S. Ray, S. Mhlanga, N. Coville, Applied Nanoscience 1 (2011) 67–77.
- 615 [23] H. Chen, G. Yang, Y. Feng, C. Shi, S. Xu, W. Cao, X. Zhang, Chem. Eng. J, 198–
- 616 **199 (2012) 45–51.**

- 617 [24] A.B. Ayusheev, O.P. Taran, I.A. Seryak, O.Y. Podyacheva, C. Descorme, M.
- Besson, L.S. Kibis, A.I. Boronin, A.I. Romanenko, Z.R. Ismagilov, V. Parmona,
 Appl. Catal. B: Environ. 146 (2014) 177–185.
- 620 [25] A.R. Silva, M. Martins, M.M.A. Freitas, A. Valente, C. Freire, B. De Castro, J.L.
 621 Figueiredo, Micropor Mesopor Mat. 55 (2002) 275–284.
- 622 [26] J.C. Lennox, R.W. Murray, J. Electroanal. Chem. 78 (1977) 395–401.
- 623 [27] W. Yantasee, Y. Lin, G.E. Fryxell, K.L. Alford, B.J. Busche, C.D. Johnson, Ind.
 624 Eng.Chem. Res. 43 (2004) 2759–2764.
- 625 [28] H. Tamai, K. Shiraki, T. Shiono, H. Yasuda, J. Colloid Interface Sci. 295 (2006)
 626 299–302.
- 627 [29] A. Houshmand, W.M.A. Wan Daud, M. Lee, M.S. Shafeeyan, Water Air Soil
 628 Pollut 223 (2012) 827–835.
- 629 [30] J.P.S. Sousa, M.F.R. Pereira, J.L. Figueiredo, Fuel Process Technol 106 (2013)
 630 727–733.
- 631 [31] A. Bagreev, J.A. Menendez, I. Dukhno, Y. Tarasenko, T.J. Bandosz, Carbon 42
 632 (2004) 469–476.
- [32] J.J.M. Órfão , A.I.M. Silva , J.C.V. Pereira , S.A. Barata , I.M. Fonseca, P.C.C.
 Faria ,M.F.R. Pereira, J. Colloid Interface Sci. 296 (2006) 480–489.
- 635 [33] M.S. Khayoon, B.H. Hameed Bioresour Technol. 102 (2011) 9229-9235.
- 636 [34] O.S.G.P. Soares, J.J.M. Órfão, M.F.R. Pereira, Appl. Catal. B: Environ. 91 (2009)
 637 441-448.
- 638 [35] J.L. Figueiredo, M.F.R. Pereira, M.A. Freitas, J.J.M. Órfão, Ind Eng Chem Res.,
 639 46 (2007) 4110–4115.

- 640 [36] Y. El-Sayed and T. J. Bandosz, Langmuir 21 (2005) 1282-1289.
- 641 [37] M. Seredych, D. Hulicova-Jurcakova, G.Q. Lu, T.J. Bandosz, Carbon 46 (2008)
 642 1475–1488.
- 643 [38] C. Moreno- Castilla, Carbon 42 (2004) 83–94.
- 644 [39] L.R. Radovic, I.F. Silva, J.I. Ume, J.A. Menéndez, C.A. Leon y Leon, A.W.
 645 Scaroni, Carbon 35 (1997) 1339–1348.
- 646 [40] S. Haydar, M.A. Ferro-Garcia, J. Rivera-Utrilla, J.P. Joly, Carbon 41(2003), 387–
 647 395.
- 648 [41] F. Villacañas, M.F.R. Pereira, J.J.M. Órfão, J.L. Figueiredo, J. Colloid Interface
 649 Sci. 293 (2006) 128–136.
- 650 [42] O. Taran, E. Polyanskaya, O. Ogorodnikova, V. Kuznetsov, V. Parmon, M.
 651 Besson, C. Descorme, Appl. Catal., A, 387 (2010) 55–66.
- 652 [43] F.W. Shaarani, B. Hameed, Chem. Eng. J., 169 (2011) 180–185.
- 653 [44] G. Yang, H. Chen, H. Qin, Y. Feng, Appl. Surf. Sci. 293 (2014) 299–305.
- [45] V. Fierro, V. Torne-Fernandez, D. Montane, A. Celzard, Micropor Mesopor Mat.
 111 (2008) 276–284.
- 656 [46] F.C. Wu, R.L. Tseng, S.C. Huang, R.S. Juang, Chem. Eng. J. 151 (2009) 1–9.
- 657 [47] G.G. Stavropoulos, P. Samaras, G.P. Sakellaropoulos, J. Hazard. Mater. 151
 658 (2008) 414–421.
- [48] V. P. Santos, M. F. R. Pereira, P. C. C. Faria and J. J. M. Órfão, J. Hazard. Mater.
 162 (2009) 736–742.
- 661 [49] C.M. Domínguez, P. Ocón, A. Quintanilla, J.A. Casas, J.J. Rodriguez, Appl. Catal.
- 662 **B**: Environ. 140–141 (2013) 663–670.

- 663 [50] T. Zhou, Y.Z. Li, J. Ji, F S. Wong, X.H. Lu, Sep. Purif. Technol., 62 (2008) 551-
- 664 <mark>558.</mark>
- [51] J.A. Melero, G. Calleja, F. Martínez, R. Molina, M.I. Pariente, Nanocomposite
 Fe₂O₃/SBA-15: Chem. Eng. J. 131 (2007) 245–256.

1	Zero-valent iron supported on nitrogen-containing activated carbon for catalytic
2	wet peroxide oxidation of phenol
3	
4	S.A. Messele ^a , O.S.G.P. Soares ^b , J.J.M. Órfão ^b , F. Stüber ^a , C. Bengoa ^a , A. Fortuny ^c , A.
5	Fabregat ^a , J. Font ^{a*}
6	
7	^a Departament d'Enginyeria Química, Universitat Rovira i Virgili, Av. Països Catalans
8	26, 43007 Tarragona, Catalunya, Spain
9	^b Laboratório de Catálise e Materiais (LCM), Laboratório Associado LSRE/LCM,
10	Departamento de Engenharia Química, Faculdade de Engenharia, Universidade do
11	Porto, Rua Dr. Roberto Frias, 4200-465 Porto, Portugal
12	^c Departament d'Enginyeria Química, Universitat Politècnica de Catalunya, EUPVG,
13	Av. Víctor Balaguer, s/n, 08800 Vilanova i la Geltrú, Catalunya, Spain
14	(*) corresponding author: jose.font@urv.cat , Tel: 34977558567, Fax: 34977559621
15	
16	Abstract
17	Zero-valent iron supported catalysts were prepared through modifying an
18	activated carbon (AC) support with different nitrogen containing precursors;
19	(ethylenediamine, urea and melamine) and impregnating it with 3 wt% of iron. The
20	supports were characterized by N_2 adsorption at -196 °C, elemental analysis (EA), the
21	pH at the point of zero charge (pH_{PZC}) and temperature programmed desorption (TPD).
22	The iron catalysts were also characterized by temperature programmed reduction (TPR).
23	Subsequently, the catalysts were tested in the adsorption and wet peroxidation of
24	phenol. The results from the different characterization techniques demonstrate that the
25	nitrogen-containing groups are successfully introduced into the carbon surface via all

26	the precursors used. The tests of the different modified carbons as adsorbents/catalysts
27	indicated that the adsorption capacity and the efficiency in phenol oxidation are
28	governed by the specific surface area and functional groups present. Both surface
29	chemistry and textural properties of carbons are influenced by the nitrogen source and
30	the type of oxygen functionalities preexisting on the surface. The modified carbons
31	supported iron catalysts revealed significantly enhanced phenol removal efficiency,
32	reaching over 85% conversion after 3 h, and showed interesting catalytic stability.
33	
34	Keywords: Phenol; Activated carbon; Surface chemistry; Catalytic wet peroxide
35	oxidation
36	
37	1. Introduction
38	
39	The ever increasing demand for water has caused considerable attention to be
40	focused towards recovery and re-use of wastewaters [1]. Besides, due to the increasing
41	complexity and toxicity of organic pollutants in industrial wastewaters and the more
42	increasingly strict environmental regulations, the toxicity nature must be reduced.
43	Phenol is among the most common water pollutants [2, 3] because it is presented in the
44	effluent of numerous industrial processes such as oil refineries, petrochemical and
45	pharmaceutical industries [4, 5]. This draws the attention of policymakers and scientists
46	to take the necessary measures by studying alternative technologies, such as Advanced
47	Oxidation Processes (AOPs).

48 AOPs are technologies based on the action of hydroxyl and other radicals to 49 oxidize recalcitrant, toxic and non-biodegradable compounds towards by-products and 50 eventually inert end products [6, 7]. Among others, hydrogen peroxide, ozone or 51 oxygen/air can be used as oxidizing agents, the processes being called wet peroxide 52 oxidation (WPO), ozonation and wet air oxidation (WAO), respectively [8-10]. Fenton 53 process mainly operates under relatively mild operating conditions using hydrogen 54 peroxide as oxidant and iron as a catalyst [11].

Fenton process achieves good results so far. However, its application is mainly dependent on the narrow operational pH range, the loss of reagent activity and the need for a subsequent separation step to remove the homogeneous catalyst from the effluent, which significantly increases the cost of the operation [12]. Therefore, taking into consideration the aforementioned inconveniences of Fenton system, a strategy is proposed to prepare a heterogeneous catalyst containing both the catalytic metal (iron) and N-functional groups on the surface of activated carbon (AC).

62 On one hand, it is well known that AC is widely used as a good adsorbent and 63 supporting material due to its excellent properties in mechanical strength and porous 64 structures [13, 14]. On the other hand, the catalytic activity of AC is far from well understood. Besides the direct relationship with the physical properties (surface area, 65 66 pore volume, etc.) of activated carbon [15], the surface chemistry can play an important 67 role [16-18]. According to the recent work of Dominguez et al. [19], the effects of 68 physicochemical characteristics of carbon materials on the rate of hydrogen peroxide 69 decomposition were explained using cyclic voltammetry, and the main results indicated 70 that the most important factor in the catalytic activity was the content of metals, in 71 particular iron, this was followed by the specific surface area and finally the content of 72 surface oxygen groups.

In addition, the presence of a chelating agent in homogeneous Fenton has demonstrated to enhance the oxidation capacity [20], either by powering the oxidation potential or keeping the iron in solution at higher pH, although the addition of a new

76 compound in solution is clearly an inconvenience [21]. Although, some interest has 77 been showed in using nitrogen-containing carbon materials as metal catalyst supports 78 [22], to the best of our knowledge, there are no data about their application in the 79 catalytic wet peroxide oxidation (CWPO) process, but there are some recent studies on 80 catalytic wet air oxidation (CWAO). Aminated activated carbon as a catalyst in the 81 CWAO of cooking wastewater has been studied by Chen et al. [23]. The COD removal 82 in CWAO of organic compounds by nitrogen containing AC was higher than that of the 83 untreated AC and suggested that nitrogen-containing functional groups contributed to 84 the enhanced activities of ACs. A similar study was reported by Ayusheev et al. [24], 85 focusing on the effect of nitrogen content in N-doped carbon nanofibers (N-CNFs) on 86 the catalytic activity of Ru/N-CNFs in the wet air oxidation of phenol. Ru-containing 87 catalysts and nitrogen in N-CNFs was found to be responsible for both the increased 88 activity and stability of the catalysts. Moreover, this study illustrates that the increase in 89 catalytic activity is related with several factors: (1) introducing a heteroatom into carbon 90 materials changes the acid-base properties of the support surface; (2) it makes possible 91 to control the size of the supported metal particles; (3) using a support of higher 92 conductivity leads to enhanced chemical reactivity for electron transfer process in a 93 catalytic system.

94 Therefore, the main objective of this work is to prepare surface modified 95 activated carbon materials (for utilisation as supports/catalysts) using different N-96 containing groups, i.e. ethylenediamine (EDA), urea or melamine. Prior to introducing 97 the N-functional groups on the AC surface, the activated carbon was oxidized by HNO₃ 98 and/or subsequently activated by thionyl chloride treatment. Later, iron impregnated and 99 then their performances where evaluated on phenol adsorption and wet peroxide 100 oxidation. 102

103 2.1. Materials

104

105 Activated carbon Norit ROX 0.8 was used as starting material for further 106 treatments. Phenol (99%) and sodium chloride (99%) were purchased from Panreac. 107 Nitric acid ($\geq 65\%$), hydrochloric acid ($\geq 37\%$), hydrogen peroxide (30% wt), urea (98%), 108 sodium hydroxide (≥97%) and iron nitrate nonahydrate (>98%) were purchased from 109 Sigma-Aldrich. Ethylenediamine (>99%) was obtained from Merck. Thionyl chloride 110 (>99%), melamine (≥99%) and toluene (>99%) were purchased from Fluka. The main 111 physico-chemical properties of these N-containing compounds are presented in Table 1. 112 All chemicals were used as received without further purification. Deionised water was 113 used throughout the work.

114

115 2.2. Modification of activated carbon supports and catalyst preparation

116

Prior to use, the Norit ROX 0.8 activated carbon was crushed and granules of 25 -50 mesh size (0.3 – 0.7 mm) were separated and used as starting material (Sample ACO). Activated carbons were modified using a combination of surface modification protocols described below, in order to obtain supports having N-containing surface groups. Iron was then supported on these modified activated carbons by incipient wetness impregnation.

123

124 2.2.1. Oxidation with nitric acid in liquid phase

The starting activated carbon (sample AC0) was oxidized with HNO₃ using a 125 mL Soxhlet extraction apparatus containing 9 g of activated carbon, connected to a boiling flask and to a condenser. A volume of 250 mL of 6 M HNO₃ was introduced into a 500 mL Pyrex round-bottom flask and heated to boiling temperature with a heating mantle. The reflux was stopped after 3 h. The oxidized activated carbon was extensively washed with distilled water to neutral pH and then dried in an oven at 110 °C for 24 h (sample AC1) [18].

133

126

127

128

129

130

131

132

134 2.2.2. Activation with thionyl chloride

135

Thionyl chloride was used as a linking agent on the surface of AC for the attachment of N-functional groups [25]. About 12 g of the oxidized activated carbon (AC1) was subsequently activated with 40 mL of 5% solution of thionyl chloride in toluene for 5 h at 70 °C. The carbon was then rinsed at least two times with toluene, and then purified by Soxhlet extraction with toluene for 2 h, and dried in an oven at 110 °C for 24 h (sample AC2) [26-28].

142

143 2.2.3. Functionalization with ethylenediamine, urea and melamine

144

Samples AC1 and AC2 were the starting materials for functionalization with Ncontaining compounds. The functionalization of activated carbons with EDA, urea or melamine can be conducted directly (using sample AC1) or indirectly (using sample AC2). In the direct method, amine groups of these compounds condense with carboxyl groups on AC to generate surface amide groups. In the indirect method, they are anchored on the surface via a linkage agent thionyl chloride (SOCl₂). In this case, first the linking agent reacts with a carboxylic group of the surface to convert the surface carboxylic groups into acyl chloride functionalities [25]; and in the second stage, the amine groups of EDA/urea/melamine condense with anchored chlorine atoms [29].

154

Ethylenediamine: 2 g of AC1 (or AC2) were refluxed with 100 mL solution of EDA in toluene (1 M) for 24 h. The amine grafted samples were washed in toluene under sonication for 10 min and then were purified in a Soxhlet extraction unit for 2 h to remove unattached or free EDA from activated carbon. The resulting activated carbons were dried in an oven at 110 °C for 24 h (samples AC1_EDA and AC2_EDA).

160

161 *Urea*: 2 g of AC1 (or AC2) were added into 100 mL of an aqueous urea solution (1 M), 162 and stirred at room temperature for 24 h. Then, the material was filtered and dried in the 163 oven. The sample treated with urea was carbonized under N_2 flow (100 cm³ min⁻¹) at 10 164 °C min⁻¹ up to 450 °C and held at this temperature for 50 min (samples AC1_Urea and 165 AC2_Urea)

166

167 *Melamine*: 2 g of AC1 (or AC2) were mixed with a melamine suspension (1.3 g of 168 melamine in 100 mL of 80% ethanol) and stirred at 70 °C for 5 h. Then the mixture was 169 boiled to evaporate the solvent and the slurry was dried at 110 °C for 24 h. The sample 170 impregnated with melamine was carbonized under N₂ flow (100 cm³ min⁻¹) at 10 °C 171 min⁻¹ up to 450 °C and held at this temperature for 50 min [30, 31]. After this treatment 172 the samples are labeled as AC1_Melamine and AC2_Melamine, respectively.

173

174 2.2.4. Preparation of catalysts

176 The active metal was supported on the original (AC0) and in all the modified 177 activated carbons by incipient wetness impregnation of an aqueous solution of 178 iron(III) nitrate nonahydrate. The impregnation was always conducted under vacuum 179 and ultrasonic mixing. The precursor solutions with calculated concentration were 180 added drop wise using a peristaltic pump and the slurry was left at room temperature 181 under ultrasonic mixing for 90 min. After impregnation, the samples were dried at 110 182 °C for 24 h and calcined under a nitrogen flow at 400 °C for 1 h, then finally reduced at 183 400 °C in hydrogen flow for 3 h (sample ZVI/support). The content of metal was 184 maintained constant at 3 wt%. The reduction temperature for the iron catalysts was 185 determined by temperature programmed reduction (TPR). The treatment methods for 186 the activated carbon used in this work are summarized in Table 2.

187

188 2.3. Characterization of supports and catalysts

189

190 The supports were characterized by N_2 adsorption at -196 °C, temperature 191 programmed desorption (TPD), elemental analysis and determination of the pH at the 192 point of zero charge (pH_{PZC}). The catalysts were also characterized by temperature 193 programmed reduction (TPR).

194

195 2.3.1. Textural characterization

196

197 The textural characterization of all the materials was checked in order to 198 evaluate if there had been significant textural changes after the modification of surface 199 chemistry. This characterization was based on the N_2 adsorption isotherms, determined 200 at -196 °C with a Quanthachrome NOVA 4200e multi-station instrument. Prior to the 201 measurements, the samples were outgassed at 120 °C for 5 h under vacuum. The specific surface area of the mesopores (S_{meso}) and the micropore volume (V_{micro}) were calculated by the t-method. Moreover, the surface area (S_{BET}) of the samples was calculated by the B.E.T. method.

205

206 2.3.2. Temperature programmed reduction (TPR)

207

208 Temperature programmed reduction (TPR) analysis allows finding the most 209 appropriate reduction temperature of the metal and to evaluate the effect of modified 210 activated carbons on the reducibility. TPR profiles were obtained with a fully automated 211 AMI-200 (Altamira Instruments). The catalyst (0.15 g) was placed in a U-shaped quartz tube inside an electrical furnace and heated at 5 °C min⁻¹ up to 600 °C under a flow of 212 5% (v/v) H₂ in Ar at 30 cm³ min⁻¹. The H₂ consumption was monitored by a thermal 213 214 conductivity detector (TCD). The temperature range where reduction occurs could be 215 indicated directly from the H₂ consumption peaks.

216

217 2.3.3. Surface chemistry characterization

218

219 The surface chemistry of the starting and modified activated carbons was 220 characterized by temperature-programmed desorption (TPD) [18]. The TPD spectra of 221 CO and CO₂ were obtained with a fully automated AMI-300 (Altamira instruments). 222 The carbon sample (0.10 g) was placed in a U-shaped quartz tube inside an electrical furnace and heated at 5 °C min⁻¹ up to 1100 °C under a constant flow rate of He at 25 223 cm³ min⁻¹ (STP). The amounts of CO (m/z = 28) and CO₂ (m/z = 44) released from the 224 carbon samples were monitored with a mass spectrometer (Dymaxion, Ametek). CO 225 226 and CO_2 were calibrated at the end of each analysis.

227 The pH at the point of zero charge (pH_{PZC}) was determined by mixing 0.05 g of 228 each sample with 20 mL of 0.01 M NaCl solution with pH values adjusted between 2 229 and 11, by adding 0.1 M HCl or 0.1 M NaOH solutions. The final pH was measured 230 after 48 h of shaking at room temperature. Blank experiments (without addition of 231 carbon) were also performed for each pH and the values measured after 48 h are 232 considered as the initial pH, in order to avoid the variation of pH caused by the effect of 233 CO_2 present in head space. The pH_{PZC} value of each carbon sample was determined by 234 intercepting the obtained final pH vs. initial pH curve with the straight line final pH =235 initial pH [32].

Carbon, hydrogen, nitrogen and oxygen (by difference) contents were
determined using a Carlo Erba EA 1108 Elemental Analyzer. Prior to the analysis,
water adsorbed on the surface of the supports was removed by drying at 110 °C in the
oven overnight.

240

241 2.4. Adsorption and Catalytic tests

242

Both adsorption and CWPO experiments were carried out in a magnetically stirred tank (batch reactor). The reactor was filled with 100 mL of a phenol aqueous solution (150 mg/L) and heated by immersion in a water bath at controlled temperature.

The solution pH was adjusted to a value of 3.0 using H_2SO_4 solution (1 M). When the desired temperature was reached (30 °C), a calculated volume of H_2O_2 (0.25 mL) was added into the system after catalyst addition (100 mg), in order to reach a H_2O_2 concentration of 750 mg/L in the CWPO runs (the theoretical stoichiometric amount needed to completely mineralize phenol); at this moment the reaction was assumed to start. The CWPO oxidation runs were carried out both using the activated carbons (without iron) and iron supported carbon as catalysts, in order to evaluate the ability of the supports alone to decompose H_2O_2 , and subsequently to promote phenol oxidation. This may give an idea to discriminate between a potential catalytic surface chemistry and the catalytic activity owing to the presence of iron.

In addition, pure adsorption runs were performed in order to assess the contribution of adsorption to the removal of phenol. The experiments were carried out in the same conditions, but without H_2O_2 addition.

259 Selected experiments were performed in triplicate and the relative error of the 260 experimental results was below $\pm 4\%$.

261

262 2.5. Analytical methods

263

Both in pure adsorption and in CWPO, liquid samples were periodically
withdrawn from the reactor. Then, each sample was filtered with a syringe filter of 0.20
μm nylon (Teknokroma, ref._TR-200101) and placed in a glass vial (Agilent) for
immediate analysis.

268 The determination of phenol concentration was performed by High Performance Liquid Chromatograph (HPLC, model 1220 Infinity LC, Agilent Technologies) 269 270 equipped with UV detector and C18 reverse phase column (Hypersil ODS, 5 µm, 25 x 271 0.4 cm from Agilent Technologies). The analyses were carried out with a mobile phase 272 (flow rate 1 mL/min) of a 40/60% mixture of methanol and ultrapure water (Milli-Q 273 water). The pH of the water was adjusted at 1.41 with sulphuric acid (H_2SO_4). The 274 detection was performed by UV absorbance at a wavelength of 254 nm. The automatic 275 injection volume was 20 µL.

276 Total Organic Carbon (TOC) was also determined at the end of the experiments

in a TC Multi Analyzer 2100 N/C equipment from Analytic Jena with a non-diffractiveIR detector.

Fe leached to the reaction media was determined by using Atomic Absorption
Spectroscopy (AAS) at 249 nm.

The main parameters used to compare the results in the discussion section are the removal of phenol (X_{PhOH}) and TOC abatement (X_{TOC}), which are respectively defined as:

284
$$X_{PhOH}(\%) = \frac{[PhOH]_0 - [PhOH]_t}{[PhOH]_0} \times 100$$
 (2)

where $[PhOH]_0$ is the initial phenol concentration and $[PhOH]_t$ is the concentration at time t, and

$$X_{TOC}(\%) = \frac{TOC_0 - TOC_t}{TOC_0} \times 100$$
(3)

where TOC_0 is the initial TOC concentration and TOC_t is the concentration at time t.

289

290 **3. Results and discussion**

291

292 3.1. Characterization of activated carbons and catalysts

293 *3.1.1. Textural properties*

BET surface area (S_{BET}), mesopore surface area (S_{meso}) and micropore volume (V_{micro}) of the original AC, modified AC and one iron supported catalyst are summarized in Table 3. The results revealed that the original AC (AC0) is highly microporous and has high BET surface area, 984 m²/g. In the case of acid-treated (AC1) and HNO₃ plus SOCl₂-treated (AC2) activated carbons, the BET surface area decreased following the treatments by 13% and 35%, respectively. The slight change in the surface area of AC1 can be explained by the abundant presence of oxygen-containing 301 groups on the surface of the activated carbon, which are introduced by the HNO₃ 302 treatment. In spite of having similar micropore volumes, AC0 and AC1 present different 303 mesopore areas, suggesting that the mesostructure of activated carbon was partially 304 destroyed after treatment with HNO₃. In the case of AC2, which is treated with HNO₃ 305 and SOCl₂, the reduction in the surface area can be explained by the introduction of 306 both oxygen and sulfur-containing groups. Such sulfur moieties (sulfones, sulfides, 307 sulfoxides, and sulfur atoms) may lead to active sites occlusion inside the pores and on 308 the surface of AC, as reported by Khayoon et al. [33]. Treatments with EDA and 309 melamine lead to a drastic reduction of the surface area. This may be due to the 310 presence of numerous groups on the activated carbon surface, which may partially block 311 the access of N₂ molecules to the micropores. However, an appreciable increase of 312 surface area and pore volume can be noticed only for urea treated samples (compared to 313 AC1 and AC2). In general, the textural properties of carbons are highly influenced by 314 the nitrogen containing precursor; specially, for the AC2_Melamine sample, for which 315 the micropore volume decreased up to nearly zero and also affects the mesostructure 316 according to the S_{meso} value. On the other hand, the impregnation of iron did not 317 substantially change the textural properties of the original carbon material due to the 318 small load of iron (3 wt%) used, which is in agreement with the previous literature [34].

319

320 *3.1.2. TPR*

TPR profiles of iron catalyst on the different supports are shown in Fig. 1. For comparison purpose, the TPR analyses of selected modified activated carbons were also performed and no additional peaks were observed. As it can be seen from Fig. 1, the position, width and intensity of the peaks mainly depend on the nature of the support. The major reduction peak is at a temperature around 400 °C for original (AC0) and

326 oxidized AC (AC1). The second peak observed at temperature above 550 °C in AC0 327 profile could be related to iron oxide which is strongly interacting with the support. In 328 general, the reduction peak for samples treated with N-containing compounds is shifted 329 to the right irrespective of the starting material (AC1 or AC2). The reduction 330 temperature depends on the degree of interaction between the active species and the 331 support. For all samples treated with N-containing compounds using either AC1 or AC2 332 as a starting material, the iron reduction peaks shift to higher temperatures. Exceptions 333 are the samples treated with urea, which present a reduction peak at lower temperature.

334

335 *3.1.3. Surface chemistry characterization*

336 *3.1.3.1. TPD* and pH_{PZC}

337 TPD analyses were carried out to evaluate the surface chemistry of the different 338 AC supports. The nature of the groups can be assessed by the decomposition 339 temperature and the type of gas released. CO and CO₂ are released by thermal 340 decomposition of oxygen containing groups on the surface of carbon materials at 341 different temperatures [18]. The TPD profiles of the original and modified activated 342 carbons are depicted in Fig. 2. From the TPD profiles, it is possible to identify and 343 quantify the amounts of the oxygenated groups of each sample. CO_2 peaks result from 344 carboxylic acids at low temperatures (150 - 450 °C), or lactones at higher temperatures 345 (600 - 800 °C); carboxylic anhydrides originate both CO and CO₂ (400 - 650 °C); 346 groups such as phenols (600 - 800 °C) and carbonyls/quinones (750 - 1000 °C) 347 originate CO peaks [35].

348 It can be seen that the nitric acid treatment (AC1) increase the amount of 349 oxygenated surface groups, which is evidenced by the increase of CO₂ (Fig. 2a) and CO

350 (Fig. 2b) released. The samples treated with EDA originate CO peaks at lower
351 temperatures (200-400 °C) than the others.

The total amounts of CO and CO_2 released from the AC materials were calculated from the corresponding TPD spectra. This information as well as the point of zero charge of the samples with different surface chemistries are shown in Table 4.

355 As expected, AC1 has a low pH_{PZC} and the highest amount of CO and CO₂, due 356 to the introduction of oxygen-containing surface groups having acidic properties, 357 mainly carboxylic acids. After thionyl chloride treatment (sample AC2), the carboxyl 358 groups were mostly converted into acid chloride groups, resulting a decrease of the 359 amount of CO and CO₂, but this material (AC2) still has acidic properties with the 360 corresponding pH_{PZC} value similar to that of sample AC1. With the exception of 361 AC2_EDA, all materials obtained after the treatment with N-containing precursors 362 using sample AC1 or AC2 as starting materials have approximately neutral or slightly 363 basic properties. This can be explained by the presence of nitrogen groups having basic 364 properties.

In general, the TPD profile shows that it is important to have high density of oxygenated groups on the surface of carbon before functionalization with the Ncontaining compounds. The N-containing groups (EDA, urea and melamine) strongly interact with carboxylic acids, anhydrides and lactones [28, 36]. Whereas, in the case of AC2, the N-functional groups were effectively introduced on the carbon surface via SOCl₂ linkage.

371

372 *3.1.3.2. Elemental analysis*

The carbon, hydrogen, nitrogen and oxygen contents obtained by elemental analysis are summarized in Table 5. Considering the treatment carried out, it was

375 expected that AC2 and AC2-based samples may contain some sulphur, which is376 summed with the oxygen content as oxygen determined by difference.

The original AC contains a small amount (about 1%) of nitrogen. Significant amounts of nitrogen were introduced on the surface after treatment with N-containing compounds. The highest amount of nitrogen incorporated into the structure was obtained when melamine was the source of nitrogen, as its molecule contains up to six nitrogen atoms.

382 Thus, the differences in the amounts of nitrogen may be related to the content of 383 nitrogen in the precursors (67% in melamine; 47% in urea and EDA) and the promoting 384 effect of surface acidity, enhanced by oxidation/activation of SOCl₂ on the retention of 385 N-containing organic bases [31,37]. On the other hand, the EDA treated samples 386 contain a relatively high content of hydrogen compared to others; this is probably due to 387 the higher content of hydrogen found in the compound. It is interesting to note that in 388 the urea and EDA treated samples similar total contents of nitrogen are found regardless 389 the pretreatment applied, either using oxidized (AC1) or activated with thionyl chloride 390 (AC2). However, in the case of melamine treated samples, the nitrogen content is higher 391 when the AC2 sample is the starting material in comparison to the inactivated (only 392 oxidized) sample AC1.

393

394 *3.2. Catalytic activity*

Although it is difficult to compare the performance of catalysts having different textural properties and surface chemistries, we performed some tests using the activated carbons as adsorbents or catalysts and the iron supported on modified AC as catalysts for the removal of phenol in solution.

The phenol removal on the original and modified activated carbons was studied as a function of time under the following conditions: 100 mL of solution having 150 mg/L of phenol, adsorbent/catalyst amount = 1 g/L, initial pH of the solution = 3.0, T = $30 \,^{\circ}$ C, atmospheric pressure and stoichiometric amount of H₂O₂.

403

404 3.2.1. Adsorption experiments

405 Adsorption experiments were carried out to evaluate the effect of textural and 406 surface properties of the original and modified activated carbons on phenol removal by 407 adsorption. The results of phenol removal in adsorption runs performed with the 408 original and modified activated carbons are compiled in Fig. 3.

409 Even though the difference is not big compared to the original AC, it is clear that 410 urea treated samples (AC1_Urea and AC2_Urea) improve the adsorption capacity of the 411 original AC (AC0). These samples (AC0, AC1_Urea and AC2_Urea) present similar 412 surface areas and pore volumes, their difference being mainly related with their surface 413 chemistry. All the other materials showed lower adsorption capacities compared to the 414 original AC. These adsorption results confirm that the textural properties of the carbon 415 materials have a primary role in adsorption of phenol, irrespective of surface chemistry, 416 but closely following the available surface area.

In the adsorption of aromatic compounds in liquid phase on activated carbons, there are mainly two types of interactions: (1) electrostatic and (2) dispersive [38]. The first mechanism is involved when the adsorbate is dissociated under the experimental conditions. For example, if solution pH > pH_{PZC}, then the carbon surface is negatively charged, the opposite occurs when the solution pH < pH_{PZC} as described elsewhere [39]. In the case of the second mechanism, the existence of π - π dispersion interactions is commonly accepted [38, 40]. Taking this in consideration, at pH of 3.0, phenol 424 $(pK_a=9.89)$ is found in solution predominantly in the molecular form and only the 425 dispersive interactions are most probably involved in its adsorption on the carbon 426 surface [41, 42].

427 Among the materials tested, the best performance was obtained with AC2_Urea, 428 reaching 67% of phenol removal after 3 h. The BET surface area of this adsorbent is 429 slightly lower and the pH_{PZC} is higher compared to the original AC. The EDA treated 430 samples show moderate adsorption performance having moderate surface area and 431 pH_{PZC} . However, the melamine treated samples show a poor adsorption performance 432 due to the drastic decrease of surface area, though having a higher pH_{PZC} compared to 433 AC1 and AC2 (pH_{PZC} of 2.5).

434 The presence of N-containing groups on the surface, increases the electronic 435 density, and therefore the basicity of samples (see pH_{PZC} in Table 4), which generally 436 favors adsorption of aromatic compounds. Recently, ammonia-modified activated 437 carbon was prepared for the adsorption of 2,4-dichlorophenol (2,4-DCP), which 438 enhances the adsorption capacity in comparison with the parent activated carbon [43]. 439 This was explained due to the basic surface functional groups created by nitrogen-440 incorporation. Similar results also presented by Yang et al [44] for phenol adsorption 441 using aminated activated carbon.

442 Moreover, the analysis of the results shows that there is a correlation between 443 the phenol removal efficiency by adsorption and the S_{BET} of the tested activated 444 carbons. The correlation is shown in Fig. 4. The figure clearly shows that the phenol 445 removal efficiency increases linearly with the S_{BET} , irrespective of the surface 446 chemistry. This indicates that the surface area of activated carbons is the principal 447 responsible for the phenol adsorption capacity. Fierro and co-workers [45] also 448 observed that the amount of adsorbed phenol in microporous activated carbons449 increases linearly with the increase in micropore volume.

450 As mentioned above, the textural and chemical properties of the tested samples 451 are different. So, an additional study is needed to take conclusions about the effect of 452 textural properties and surface functionalities. For this purpose, the apparent rate 453 constants for the first-order (k_1) and second-order (k_2) adsorption models were 454 determined for all samples tested. The constants of the two models are listed in Table 6. 455 The ratio between k_1 or k_2 and S_{BET} is presented in the same table. Even though, both the first-order and the second-order model fit the experimental data quite well with R^2 456 457 values close to unity, the second-order model was suitable for the adsorption of lower 458 molecular weight adsorbates on smaller adsorbent particles as reported by Wu et al. [46] 459 and therefore the corresponding kinetic constants are used in the following discussion.

460 It is reported that phenol adsorption is dependent on both the surface area and 461 the presence of surface groups [45, 47]. In order to evaluate the influence of the 462 nitrogen content, the second-order apparent rate constants (k_2) are normalized by the 463 specific surface area (S_{BET}). It can be observed in Fig. 5 that high normalized adsorption 464 rate constants correspond to samples with high amounts of N-containing surface groups 465 (see Table 5).

466

467 *3.2.2. CWPO experiments*

-00

468 *3.2.2.1. Modified activated carbons as catalysts*

In order to evaluate the ability of the modified activated carbons to act as catalysts (without any supported metal) in the CWPO of phenol, runs were performed under the same conditions as the adsorption tests but now adding H_2O_2 . The corresponding phenol removal curves are shown in Fig. 6. 473 Phenol removal by CWPO in the presence of modified activated carbons as 474 catalysts was similar or slightly higher than those obtained by adsorption. However, the 475 results of phenol removal by adsorption (Fig. 3) and CWPO (Fig. 6) lead to the 476 conclusion that these modified carbon materials are not particularly active for CWPO of 477 phenol. For instance, the phenol removal at 60 min for AC1_Urea is 43% without H₂O₂ 478 and 42% in the presence of H_2O_2 , or 65% and 64%, respectively, at 180 min, which 479 suggests that both cases are governed by adsorption. So, the removal of phenol is 480 mainly due to pure adsorption. Similar conclusion was obtained in a previous work for 481 some dyes [48]. According to Dominguez et al [49], as the initial phenol concentration 482 and the phenol/catalyst ratio used in the present work are relatively low, a large fraction 483 of active sites on the carbon surface is available for hydrogen peroxide decomposition 484 into hydroxyl radicals, which may be consumed by non-effective reactions. This 485 probably explains the poor performance of activated carbon in CWPO compared to 486 adsorption.

487

488 *3.2.2.2. Iron supported catalysts*

489 In order to study the influence of surface chemical characteristics of the support 490 on the activity of iron catalysts in phenol oxidation, a set of runs were performed in 491 CWPO for different ZVI supported catalysts. Fig. 7 shows the evolution of phenol 492 removal in the presence of the iron catalyst supported on different supports. 493 ZVI/AC2_Melamine and ZVI/AC2_EDA show a good and similar performance, 494 reaching above 80% phenol conversion after 60 min. On the contrary, the phenol 495 removal by pure adsorption and CWPO using the same material without the presence of 496 ZVI (i.e AC2_Melamine and AC2_EDA) shows low removal efficiency below 15 % 497 after 60 min.

498 It must be strongly noted that the presence of ZVI on N-containing catalysts 499 yields a better phenol removal, reaching values over 85% after 3 h. Here, ZVI acts as a 500 heterogeneous catalyst for the degradation of phenol in aqueous solution by effectively 501 decomposing hydrogen peroxide and generating hydroxyl radical, which reacts at high 502 rate with phenol and its intermediates [50]. These results revealed that there is a direct 503 and positive relation between the catalytic activity and the nitrogen content of the 504 materials. Therefore, the activity of iron supported catalysts based on N-containing AC 505 in CWPO of phenol is enhanced when compared to that of the untreated ones. It is thus 506 observed that the presence of nitrogen groups on the surface of activated carbon, in 507 addition to iron, clearly increases the phenol removal. This synergy can be probably 508 assigned to the ability of retaining the iron ions close to the carbon surface due to the 509 complexing properties of these N-containing groups.

510 Similar trends were observed for TOC removal at the end of adsorption and 511 oxidation tests. Fig. 8 shows that the EDA and melamine treated AC present lower TOC 512 removal performance in the adsorption or oxidation tests when these materials were 513 used as adsorbents or catalysts, respectively. However, the ZVI supported catalysts on 514 the EDA and melamine treated supports also promote the TOC removal. The main 515 reason for this can be explained by the synergetic effect of the presence of nitrogen 516 group and iron on the surface of activated carbon, promoting oxidation instead of 517 adsorption. In this case, the access of H_2O_2 to the iron on the carbon surface could be 518 easier than for the other samples and accelerates the hydroxyl radical generation for 519 deep oxidation of phenol, as observed in Fig. 7. However, additional studies are still 520 required to reveal the exact origin of these effects.

521 Leaching tests were performed in order to evaluate the stability and contribution 522 of homogeneous reaction for each catalyst after 120 and 180 min of reaction. Leaching

523 of iron corresponds to values below 0.01% of the metal initially present in almost all 524 catalysts; also limited values of 0.03 to 0.04% were measured in the case of AC1 and 525 AC2 supported catalysts. This fact can be related with the pH_{PZC} of the support. In fact, 526 AC1 and AC2 present the lowest pH_{PZC} and a relatively high iron leaching. In 527 conclusion, the contribution of homogeneous system is quite negligible, as the amount 528 of leached iron is quite low.

529 Since the stability of a catalyst is one important aspect in the general evaluation 530 of its performance, the most efficient catalysts of this study (ZVI/AC2 Melamine and 531 ZVI/AC2_EDA) were reused three times in consecutive CWPO reactions (Fig. 9). A slight loss of catalyst activity is observed with both catalysts from the 1st to the 2nd run, 532 but the activity being practically maintained from the 2nd to the 3rd run, and the actual 533 removal of phenol is around 80%. The decrease in the removal of phenol observed from 534 535 the first to the second run may be due to the lower adsorption capacity of the used 536 catalyst and/or the slight loss of the catalyst during the recovery process. In fact, at the 537 end of the experiment, a fraction of organic compounds will remain adsorbed onto the 538 catalyst under the employed operating conditions. These compounds can be either 539 phenol or intermediates formed during the oxidation reaction, as discussed elsewhere 540 [51]. In any case, both catalysts (ZVI/AC2 Melamine and ZVI/AC2 EDA) still present 541 a good efficiency under consecutive runs.

542

543 **4.** Conclusions

544

545 Activated carbons with different N-containing precursors were modified, 546 characterized and tested as adsorbents or catalysts for adsorption and peroxide oxidation of phenol. Besides, these modified carbons were also impregnated with iron and used ascatalysts for CWPO.

549 Treatments with EDA and melamine lead to a drastic surface area reduction. 550 This may be due to the presence of numerous groups on the activated carbon surface, 551 which may partially block the access of N₂ molecules to the micropores.

In the urea and EDA treated samples similar total contents of nitrogen are found regardless of the pretreatment applied. However, in the case of melamine treated samples, the nitrogen content is significantly higher when the oxidized activated carbon treated with thionyl chloride is the starting material.

556 Phenol removal by CWPO in the presence of carbon materials was slightly 557 higher than those obtained by adsorption. However, the results lead to the conclusion 558 that these materials are not particularly active for the reaction, the removal of phenol 559 being mainly due to adsorption.

The iron supported catalysts based on N-containing AC show the highest phenol removal efficiency, reaching values over 85% conversion after 3 h. This result revealed that there is a positive relation between the catalytic activity and the nitrogen content of the materials.

564

565 Acknowledgements

566

567 Financial support for this research was provided by the Spanish Ministerio de 568 Educación y Ciencia and FEDER, projects CTM2008-03338 and CTM2011-23069. The 569 Spanish Ministerio de Ciencia e Innovación is also thanked for providing a doctoral 570 scholarship (Programme FPI, BES-2009-017016) to carry out this research work. The 571 author's research group is recognised by the Comissionat per a Universitats i Recerca

572	del	DIUE de la Generalitat de Catalunya (2009SGR865) and supported by the
573	Univ	ersitat Rovira i Virgili (2010PFR-URV-B2-41). Funding from FCT and FEDER in
574	the fi	ramework of Program COMPETE (FCT grant SFRH/BPD/80435/2011 and Project
575	Pest-	C/EQB/ LA0020/2011) is acknowledged by the LCM group of authors.
576		
577	Refe	rences
578		
579	[1]	G. Annadurai, S.R. Babu, K.P.O. Mahesh, T. Murugesan, Bioprocess Eng., 2
580		(2000) 493–501.
581	[2]	P.S. Nayak, B.K. Singh, Desalination 207 (2007) 71–79.
582	[3]	C.E. Paisio, E. Agostini, P.S. González, M.L. Bertuzzi, J. Hazard. Mater. 167
583		(2009) 64–68.
584	[4]	A. Fortuny, C. Bengoa, J. Font, F. Castells, A. Fabregat, Catal. Today 53 (1999)
585		107–114.
586	[5]	A. Santos, P. Yustos, A. Quintanilla, S. Rodríguez, F. García-Ochoa, Appl. Catal.
587		B 39 (2002) 97–113.
588	[6]	C. Comninellis, A. Kapalka, S. Malato, S.A. Parsons, I. Poulios, D. Mantzavinos,
589		J Chem Technol Biotechnol, 83 (2008) 769–776.
590	[7]	D. Mantzavinos, E. Psillakis, J Chem Technol Biotechnol, 79 (2004) 431–454.
591	[8]	V. García-Molina, M. López-Arias, M. Florczyk, E. Chamarro, S. Esplugas, Water
592		Res. 39 (2005) 795–802.
593	[9]	F. Luck, Catal. Today 53 (1999) 81–91.
594	[10]	H. Debellefontaine, S. Crispel, P. Reilhac, F. Périé, J. Foussard, Chem. Eng. Sci.
595		54 (1999) 4953-4959.

- 596 [11] M. Klavarioti, D. Mantzavinos, D. Kassinos, Environ. Int. 35 (2009) 402–417.
- 597 [12] J.J. Pignatello, E.E. Oliveros, A. Mackay, Environ. Sci. Technol. 36 (2006) 1–84.
- 598 [13] E. Ahumada, H. Lizama, F. Orellana, C. Suárez, A. Huidobro, A. Sepúlveda599 Escribano, F. Rodríguez-Reinoso, Carbon 40 (2002) 2827–2834.
- 600 [14] F. Rodríguez-Reinoso, Carbon, 36 (1998) 159–175.
- 601 [15] M. Eugeni, F. Stüber, A. Fortuny, A. Fabregat, J. Carrera, J. Font, Appl. Catal. B:
 602 Environ. 58 (2005) 105–114.
- 603 [16] F. Stüber, J. Font, A. Fortuny, C. Bengoa, A. Eftaxias, and A. Fabregat, Top
 604 Catal., 33 (2005) 3–50.
- 605 [17] A. Quintanilla, J.A.Casas., J.J. Rodriguez, M.T. Kreutzer, F. Kapteijn, J.A.
 606 Moulijn, Int. J. Chem. Reactor Eng. 5 (2007) A62.
- 607 [18] J.L. Figueiredo, M.F.R. Pereira, M.M.A. Freitas, J.J.M. Órfão, Carbon 37 (1999)
 608 1379–1389.
- 609 [19] C.M. Domínguez, A. Quintanilla, P. Ocón, J.A. Casas, J.J. Rodriguez, Carbon 60
 610 (2013) 76–83.
- 611 [20] N. Klamerth, S. Malato, A. Agüera, A. Fernández-Alba, Water Res. 47 (2013)
 612 833–840.
- 613 [21] Z. Yuan, J.M. VanBriesen, Environ. Eng. Sci. 23 (2006) 533–544.
- 614 [22] L. Mabena, S. Ray, S. Mhlanga, N. Coville, Applied Nanoscience 1 (2011) 67–77.
- 615 [23] H. Chen, G. Yang, Y. Feng, C. Shi, S. Xu, W. Cao, X. Zhang, Chem. Eng. J, 198–
 616 199 (2012) 45–51.

- 617 [24] A.B. Ayusheev, O.P. Taran, I.A. Seryak, O.Y. Podyacheva, C. Descorme, M.
- Besson, L.S. Kibis, A.I. Boronin, A.I. Romanenko, Z.R. Ismagilov, V. Parmona,
 Appl. Catal. B: Environ. 146 (2014) 177–185.
- 620 [25] A.R. Silva, M. Martins, M.M.A. Freitas, A. Valente, C. Freire, B. De Castro, J.L.
 621 Figueiredo, Micropor Mesopor Mat. 55 (2002) 275–284.
- 622 [26] J.C. Lennox, R.W. Murray, J. Electroanal. Chem. 78 (1977) 395–401.
- 623 [27] W. Yantasee, Y. Lin, G.E. Fryxell, K.L. Alford, B.J. Busche, C.D. Johnson, Ind.
 624 Eng.Chem. Res. 43 (2004) 2759–2764.
- [28] H. Tamai, K. Shiraki, T. Shiono, H. Yasuda, J. Colloid Interface Sci. 295 (2006)
 299–302.
- 627 [29] A. Houshmand, W.M.A. Wan Daud, M. Lee, M.S. Shafeeyan, Water Air Soil
 628 Pollut 223 (2012) 827–835.
- 629 [30] J.P.S. Sousa, M.F.R. Pereira, J.L. Figueiredo, Fuel Process Technol 106 (2013)
 630 727–733.
- 631 [31] A. Bagreev, J.A. Menendez, I. Dukhno, Y. Tarasenko, T.J. Bandosz, Carbon 42
 632 (2004) 469–476.
- [32] J.J.M. Órfão , A.I.M. Silva , J.C.V. Pereira , S.A. Barata , I.M. Fonseca, P.C.C.
 Faria ,M.F.R. Pereira, J. Colloid Interface Sci. 296 (2006) 480–489.
- 635 [33] M.S. Khayoon, B.H. Hameed Bioresour Technol. 102 (2011) 9229-9235.
- 636 [34] O.S.G.P. Soares, J.J.M. Órfão, M.F.R. Pereira, Appl. Catal. B: Environ. 91 (2009)
 637 441–448.
- [35] J.L. Figueiredo, M.F.R. Pereira, M.A. Freitas, J.J.M. Órfão, Ind Eng Chem Res.,
 46 (2007) 4110–4115.

- 640 [36] Y. El-Sayed and T. J. Bandosz, Langmuir 21 (2005) 1282-1289.
- 641 [37] M. Seredych, D. Hulicova-Jurcakova, G.Q. Lu, T.J. Bandosz, Carbon 46 (2008)
 642 1475–1488.
- 643 [38] C. Moreno- Castilla, Carbon 42 (2004) 83–94.
- 644 [39] L.R. Radovic, I.F. Silva, J.I. Ume, J.A. Menéndez, C.A. Leon y Leon, A.W.
 645 Scaroni, Carbon 35 (1997) 1339–1348.
- 646 [40] S. Haydar, M.A. Ferro-Garcia, J. Rivera-Utrilla, J.P. Joly, Carbon 41(2003), 387–
 647 395.
- 648 [41] F. Villacañas, M.F.R. Pereira, J.J.M. Órfão, J.L. Figueiredo, J. Colloid Interface
 649 Sci. 293 (2006) 128–136.
- 650 [42] O. Taran, E. Polyanskaya, O. Ogorodnikova, V. Kuznetsov, V. Parmon, M.
 651 Besson, C. Descorme, Appl. Catal., A, 387 (2010) 55–66.
- 652 [43] F.W. Shaarani, B. Hameed, Chem. Eng. J., 169 (2011) 180–185.
- 653 [44] G. Yang, H. Chen, H. Qin, Y. Feng, Appl. Surf. Sci. 293 (2014) 299–305.
- [45] V. Fierro, V. Torne-Fernandez, D. Montane, A. Celzard, Micropor Mesopor Mat.
 111 (2008) 276–284.
- 656 [46] F.C. Wu, R.L. Tseng, S.C. Huang, R.S. Juang, Chem. Eng. J. 151 (2009) 1–9.
- 657 [47] G.G. Stavropoulos, P. Samaras, G.P. Sakellaropoulos, J. Hazard. Mater. 151
 658 (2008) 414–421.
- [48] V. P. Santos, M. F. R. Pereira, P. C. C. Faria and J. J. M. Órfão, J. Hazard. Mater.
 162 (2009) 736–742.
- 661 [49] C.M. Domínguez, P. Ocón, A. Quintanilla, J.A. Casas, J.J. Rodriguez, Appl. Catal.
 662 B: Environ. 140–141 (2013) 663–670.

- 663 [50] T. Zhou, Y.Z. Li, J. Ji, F S. Wong, X.H. Lu, Sep. Purif. Technol., 62 (2008) 551–
 664 558.
- [51] J.A. Melero, G. Calleja, F. Martínez, R. Molina, M.I. Pariente, Nanocomposite
 Fe₂O₃/SBA-15: Chem. Eng. J. 131 (2007) 245–256.

1	Zero-valent iron supported on nitrogen-containing activated carbon for catalytic
2	wet peroxide oxidation of phenol
3	
4	S.A. Messele ^a , O.S.G.P. Soares ^b , J.J.M. Órfão ^b , F. Stüber ^a , C. Bengoa ^a , A. Fortuny ^c , A.
5	Fabregat ^a , J. Font ^{a*}
6	
7	^a Departament d'Enginyeria Química, Universitat Rovira i Virgili, Av. Països Catalans
8	26, 43007 Tarragona, Catalunya, Spain
9	^b Laboratório de Catálise e Materiais (LCM), Laboratório Associado LSRE/LCM,
10	Departamento de Engenharia Química, Faculdade de Engenharia, Universidade do
11	Porto, Rua Dr. Roberto Frias, 4200-465 Porto, Portugal
12	^c Departament d'Enginyeria Química, Universitat Politècnica de Catalunya, EUPVG,
13	Av. Víctor Balaguer, s/n, 08800 Vilanova i la Geltrú, Catalunya, Spain
14	(*) corresponding author: jose.font@urv.cat, Tel: 34977558567, Fax: 34977559621
15	
16	
17	Abstract
18	Zero-valent iron supported catalysts were prepared through modifying an
19	activated carbon (AC) support with different nitrogen containing precursors;
20	(ethylenediamine, urea and melamine) and impregnating it with 3 wt% of iron. The
21	supports were characterized by N_2 adsorption at -196 °C, elemental analysis (EA), the
22	pH at the point of zero charge (pH_{PZC}) and temperature programmed desorption (TPD).
23	The iron catalysts were also characterized by temperature programmed reduction (TPR).
24	Subsequently, the catalysts were tested in the adsorption and wet peroxidation of
25	phenol. The results from the different characterization techniques demonstrate that the
1 nitrogen-containing groups are successfully introduced into the carbon surface via all 2 the precursors used. The tests of the different modified carbons as adsorbents/catalysts 3 indicated that the adsorption capacity and the efficiency in phenol oxidation are 4 governed by the specific surface area and functional groups present. Both surface 5 chemistry and textural properties of carbons are influenced by the nitrogen source and 6 the type of oxygen functionalities preexisting on the surface. The modified carbons 7 supported iron catalysts revealed significantly enhanced phenol removal efficiency, 8 reaching over 85% conversion after 3 h.

9

10 Keywords: Phenol; Activated carbon; Surface chemistry; Catalytic wet peroxide

11 oxidation

12

13 **1. Introduction**

14

15 The ever increasing demand for water has caused considerable attention to be 16 focused towards recovery and re-use of wastewaters [1]. Besides, due to the increasing 17 complexity and toxicity of organic pollutants in industrial wastewaters and the more 18 increasingly strict environmental regulations, the toxicity nature must be reduced. 19 Phenol is among the most common water pollutants [2, 3] because it is presented in the 20 effluent of numerous industrial processes such as oil refineries, petrochemical and 21 pharmaceutical industries [4, 5]. This draws the attention of policymakers and scientists 22 to take the necessary measures by studying alternative technologies, such as Advanced 23 Oxidation Processes (AOPs).

AOPs are technologies based on the action of hydroxyl and other radicals to oxidize recalcitrant, toxic and non-biodegradable compounds towards by-products and 1 eventually inert end products [6, 7]. Among others, hydrogen peroxide, ozone or 2 oxygen/air can be used as oxidizing agents, the processes being called wet peroxide 3 oxidation (WPO), ozonation and wet air oxidation (WAO), respectively [8-10]. Fenton 4 process (Fe^{2+}/H_2O_2) belongs to catalytic wet peroxide oxidation (CWPO) technologies 5 and operates under relatively mild operating conditions (room temperature and 6 atmospheric pressure) using hydrogen peroxide as oxidant and iron as a catalyst [11].

Fenton process achieves good results so far. However, its application is mainly
dependent on the narrow operational pH range, the loss of reagent activity and the need
for a subsequent separation step to remove the homogeneous catalyst from the effluent,
which significantly increases the cost of the operation [12].

11 Therefore, taking into consideration the aforementioned inconveniences of 12 Fenton system, a strategy is proposed to prepare a heterogeneous catalyst containing 13 both the catalytic metal (iron) and N-functional groups on the surface of activated 14 carbon (AC).

On one hand, it is well known that AC is widely used as a good adsorbent and supporting material due to its excellent properties in mechanical strength and porous structures [13, 14]. On the other hand, the catalytic activity of AC is far from well understood. Besides the direct relationship with the physical properties (surface area, pore volume, etc.) of activated carbon [15], the surface chemistry can play an important role [16-18]. Moreover, it has been reported [19] that the mineral content (in particular Fe content) of AC is a key to display catalytic activities.

In addition, the presence of a chelating agent in homogeneous Fenton has demonstrated to enhance the oxidation capacity [20], either by powering the oxidation potential or keeping the iron in solution at higher pH, although the addition of a new compound in solution is clearly an inconvenience [21]. Ideally, heterogenisation of this

1 originally homogeneous catalytic system should mostly overcome the mentioned 2 inconveniences. Therefore, the attachment of a ligand (N-functional groups) on the AC 3 surface is expected to help the oxidative power of H_2O_2 . However, care should be given 4 to the load of iron or ligand to be incorporated on the AC.

5 Therefore, the main objective of this work is to prepare surface modified 6 activated carbon materials (for utilisation as supports/catalysts) using different N-7 containing groups, i.e. ethylenediamine (EDA), urea or melamine, which are later iron 8 impregnated and then their performances where evaluated on phenol adsorption and wet 9 peroxide oxidation.

10

11 **2. Experimental**

12

```
13 2.1. Materials
```

14

15 Activated carbon Norit ROX 0.8 was used as starting material for further 16 treatments. Phenol (99%) and sodium chloride (99%) were purchased from Panreac. 17 Nitric acid ($\geq 65\%$), hydrochloric acid (>37%), hydrogen peroxide (30% wt), urea (98%), 18 sodium hydroxide (≥97%) and iron nitrate nonahydrate (>98%) were purchased from 19 Sigma-Aldrich. Ethylenediamine (>99%) was obtained from Merck. Thionyl chloride 20 (>99%), melamine (≥99%) and toluene (>99%) were purchased from Fluka. The main 21 physico-chemical properties of these N-containing compounds are presented in Table 1. 22 All chemicals were used as received without further purification. Deionised water was 23 used throughout the work.

24

25 2.2. Modification of activated carbon supports and catalyst preparation

2	Prior to use, the Norit ROX 0.8 activated carbon was crushed and granules of 25
3	-50 mesh size (0.3 -0.7 mm) were separated and used as starting material (Sample
4	AC0). Activated carbons were modified using a combination of surface modification
5	protocols described below, in order to obtain supports having N-containing surface
6	groups. Iron was then supported on these modified activated carbons by incipient
7	wetness impregnation.
8	
9	2.2.1. Oxidation with nitric acid in liquid phase
10	
11	The starting activated carbon (sample AC0) was oxidized with HNO3 using a
12	125 mL Soxhlet extraction apparatus containing 9 g of activated carbon, connected to a
13	boiling flask and to a condenser. A volume of 250 mL of 6 M HNO3 was introduced
14	into a 500 mL Pyrex round-bottom flask and heated to boiling temperature with a
15	heating mantle. The reflux was stopped after 3 h. The oxidized activated carbon was
16	extensively washed with distilled water to neutral pH and then dried in an oven at 110
17	°C for 24 h (sample AC1) [18].
18	
19	2.2.2. Activation with thionyl chloride
20	
21	About 12 g of the oxidized activated carbon (AC1) was subsequently activated
22	with 40 mL of 5% solution of thionyl chloride in toluene for 5 h at 70 °C. The carbon
23	was then rinsed at least two times with toluene, and then purified by Soxhlet extraction

- 24 with toluene for 2 h, and dried in an oven at 110 °C for 24 h (sample AC2) [22-24].
- 25

3 Samples AC1 and AC2 were the starting materials for functionalization with N-4 containing compounds. The functionalization of activated carbons with EDA, urea or 5 melamine can be conducted directly (using sample AC1) or indirectly (using sample 6 AC2). In the direct method, amine groups of these compounds condense with carboxyl 7 groups on AC to generate surface amide groups. In the indirect method, they are 8 anchored on the surface via a linkage agent thionyl chloride (SOCl₂). In this case, first 9 the linking agent reacts with a carboxylic group of the surface to convert the surface 10 carboxylic groups into acyl chloride functionalities [25]; and in the second stage, the 11 amine groups of EDA/urea/melamine condense with anchored chlorine atoms [26]. 12

Ethylenediamine: 2 g of AC1 (or AC2) were refluxed with 100 mL solution of EDA in toluene (1 M) for 24 h. The amine grafted samples were washed in toluene under sonication for 10 min and then were purified in a Soxhlet extraction unit for 2 h to remove unattached or free EDA from activated carbon. The resulting activated carbons were dried in an oven at 110 °C for 24 h (samples AC1_EDA and AC2_EDA).

18

19 *Urea*: 2 g of AC1 (or AC2) were added into 100 mL of an aqueous urea solution (1 M), 20 and stirred at room temperature for 24 h. Then, the material was filtered and dried in the 21 oven. The sample treated with urea was carbonized under N_2 flow (100 cm³ min⁻¹) at 10 22 °C min⁻¹ up to 450 °C and held at this temperature for 50 min (samples AC1_Urea and 23 AC2_Urea)

1	Melamine: 2 g of AC1 (or AC2) were mixed with a melamine suspension (1.3 g of
2	melamine in 100 mL of 80% ethanol) and stirred at 70 °C for 5 h. Then the mixture was
3	boiled to evaporate the solvent and the slurry was dried at 110 °C for 24 h. The sample
4	impregnated with melamine was carbonized under N_2 flow (100 cm ³ min ⁻¹) at 10 °C
5	min ⁻¹ up to 450 °C and held at this temperature for 50 min [27, 28]. After this treatment
6	the samples are labeled as AC1_Melamine and AC2_Melamine, respectively.

8 2.2.4. Preparation of catalysts

9

10 The active metal was supported on the original (AC0) and in all the modified 11 activated carbons by incipient wetness impregnation of an aqueous solution of 12 iron(III) nitrate nonahydrate. The impregnation was always conducted under vacuum 13 and ultrasonic mixing. The precursor solutions with calculated concentration were 14 added drop wise using a peristaltic pump and the slurry was left at room temperature 15 under ultrasonic mixing for 90 min. After impregnation, the samples were dried at 110 16 °C for 24 h and calcined under a nitrogen flow at 400 °C for 1 h, then finally reduced at 17 400 °C in hydrogen flow for 3 h (sample ZVI/support). The content of metal was 18 maintained constant at 3 wt%. The reduction temperature for the iron catalysts was 19 determined by temperature programmed reduction (TPR). The treatment methods for 20 the activated carbon used in this work are summarized in Table 2.

21

22 2.3. Characterization of supports and catalysts

23

24 The supports were characterized by N_2 adsorption at -196 °C, temperature 25 programmed desorption (TPD), elemental analysis and determination of the pH at the point of zero charge (pH_{PZC}). The catalysts were also characterized by temperature
 programmed reduction (TPR).

3

4 2.3.1. Textural characterization

5

6 The textural characterization of all the materials was checked in order to 7 evaluate if there had been significant textural changes after the modification of surface 8 chemistry. This characterization was based on the N₂ adsorption isotherms, determined 9 at -196 °C with a Quanthachrome NOVA 4200e multi-station instrument. Prior to the 10 measurements, the samples were outgassed at 120 °C for 5 h under vacuum. The 11 specific surface area of the mesopores (S_{meso}) and the micropore volume (V_{micro}) were 12 calculated by the t-method. Moreover, the surface area (S_{BET}) of the samples was 13 calculated by the B.E.T. method.

14

15 2.3.2. Temperature programmed reduction (TPR)

16

17 Temperature programmed reduction (TPR) analysis allows finding the most 18 appropriate reduction temperature of the metal and to evaluate the effect of modified 19 activated carbons on the reducibility. TPR profiles were obtained with a fully automated 20 AMI-200 (Altamira Instruments). The catalyst (0.15 g) was placed in a U-shaped quartz tube inside an electrical furnace and heated at 5 °C min⁻¹ up to 600 °C under a flow of 21 5% (v/v) H₂ in Ar at 30 cm³ min⁻¹. The H₂ consumption was monitored by a thermal 22 23 conductivity detector (TCD). The temperature range where reduction occurs could be 24 indicated directly from the H₂ consumption peaks.

3 The surface chemistry of the starting and modified activated carbons was 4 characterized by temperature-programmed desorption (TPD) [18]. The TPD spectra of 5 CO and CO₂ were obtained with a fully automated AMI-300 (Altamira instruments). 6 The carbon sample (0.10 g) was placed in a U-shaped quartz tube inside an electrical furnace and heated at 5 °C min⁻¹ up to 1100 °C under a constant flow rate of He at 25 7 8 $cm^3 min^{-1}$ (STP). The amounts of CO (m/z = 28) and CO₂ (m/z = 44) released from the 9 carbon samples were monitored with a mass spectrometer (Dymaxion, Ametek). CO 10 and CO_2 were calibrated at the end of each analysis.

11 The pH at the point of zero charge (pH_{PZC}) was determined by mixing 0.05 g of 12 each sample with 20 mL of 0.01 M NaCl solution with pH values adjusted between 2 13 and 11, by adding 0.1 M HCl or 0.1 M NaOH solutions. The final pH was measured 14 after 48 h of shaking at room temperature. Blank experiments (without addition of 15 carbon) were also performed for each pH and the values measured after 48 h are 16 considered as the initial pH, in order to avoid the variation of pH caused by the effect of 17 CO₂ present in head space. The pH_{PZC} value of each carbon sample was determined by 18 intercepting the obtained final pH vs. initial pH curve with the straight line final pH =19 initial pH [29].

20 Carbon, hydrogen, nitrogen and oxygen (by difference) contents were 21 determined using a Carlo Erba EA 1108 Elemental Analyzer. Prior to the analysis, 22 water adsorbed on the surface of the supports was removed by drying at 110 °C in the 23 oven overnight.

Both adsorption and CWPO experiments were carried out in a magnetically
stirred tank (batch reactor). The reactor was filled with 100 mL of a phenol aqueous
solution (150 mg/L) and heated by immersion in a water bath at controlled temperature.

5 The solution pH was adjusted to a value of 3.0 using H₂SO₄ solution (1 M). 6 When the desired temperature was reached (30 °C), a calculated volume of H_2O_2 (0.25 mL) was added into the system after catalyst addition (100 mg), in order to reach a 7 8 H₂O₂ concentration of 750 mg/L in the CWPO runs (the theoretical stoichiometric 9 amount needed to completely mineralize phenol); at this moment the reaction was 10 assumed to start. The CWPO oxidation runs were carried out both using the activated 11 carbons (without iron) and iron supported carbon as catalysts, in order to evaluate the ability of the supports alone to decompose H_2O_2 , and subsequently to promote phenol 12 13 oxidation. This may give an idea to discriminate between a potential catalytic surface 14 chemistry and the catalytic activity owing to the presence of iron.

In addition, pure adsorption runs were performed in order to assess the contribution of adsorption to the removal of phenol. The experiments were carried out in the same conditions, but without H_2O_2 addition.

18 Selected experiments were performed in triplicate and the relative error of the
19 experimental results was below ±4%.

20

21 2.5. Analytical methods

22

Both in pure adsorption and in CWPO, liquid samples were periodically
withdrawn from the reactor. Then, each sample was filtered with a syringe filter of 0.20

μm nylon (Teknokroma, ref._TR-200101) and placed in a glass vial (Agilent) for
 immediate analysis.

3 The determination of phenol concentration was performed by High Performance 4 Liquid Chromatograph (HPLC, model 1220 Infinity LC, Agilent Technologies) 5 equipped with UV detector and C18 reverse phase column (Hypersil ODS, 5 µm, 25 x 6 0.4 cm from Agilent Technologies). The analyses were carried out with a mobile phase 7 (flow rate 1 mL/min) of a 40/60% mixture of methanol and ultrapure water (Milli-Q 8 water). The pH of the water was adjusted at 1.41 with sulphuric acid (H_2SO_4). The 9 detection was performed by UV absorbance at a wavelength of 254 nm. The automatic 10 injection volume was 20 µL.

Total Organic Carbon (TOC) was also determined at the end of the experiments
in a TC Multi Analyzer 2100 N/C equipment from Analytic Jena with a non-diffractive
IR detector.

Fe leached to the reaction media was determined by using Atomic Absorption
Spectroscopy (AAS) at 249 nm.

16 The main parameters used to compare the results in the discussion section are 17 the removal of phenol (X_{PhOH}) and TOC abatement (X_{TOC}), which are respectively 18 defined as:

19
$$X_{PhOH}(\%) = \frac{[PhOH]_0 - [PhOH]_t}{[PhOH]_0} \times 100$$
 (2)

20 where $[PhOH]_0$ is the initial phenol concentration and $[PhOH]_t$ is the concentration at 21 time t, and

$$X_{TOC}(\%) = \frac{TOC_0 - TOC_t}{TOC_0} \times 100$$
(3)

23 where TOC_0 is the initial TOC concentration and TOC_t is the concentration at time t.

24

3 3.1. Characterization of activated carbons and catalysts

4

5 Activated carbons were characterized by N_2 adsorption at -196 °C, temperature 6 programmed desorption (TPD), elemental analysis (EA), and by determination of the 7 pH at the point of zero charge (pH_{PZC}). The catalysts were also characterized by 8 temperature programmed reduction (TPR).

9

10 3.1.1. Textural properties

11 BET surface area (S_{BET}), mesopore surface area (S_{meso}) and micropore volume 12 (V_{micro}) of the original AC, modified AC and one iron supported catalyst are 13 summarized in Table 3. The results revealed that the original AC (AC0) is highly microporous and has high BET surface area, 984 m^2/g . In the case of acid-treated (AC1) 14 15 and HNO₃ plus SOCl₂-treated (AC2) activated carbons, the BET surface area decreased 16 following the treatments by 13% and 35%, respectively. The slight change in the 17 surface area of AC1 can be explained by the abundant presence of oxygen-containing 18 groups on the surface of the activated carbon, which are introduced by the HNO_3 19 treatment. These oxygen containing groups possibly block the entry of N₂ inside the 20 small pores [30] or the carbon structure of activated carbon is partially destroyed and 21 some micropores are vanished by treatment with HNO₃ [24]. Nevertheless, samples 22 AC0 and AC1 have similar micropore volumes. In the case of AC2, which is treated with HNO₃ and SOCl₂, the reduction in the surface area can be explained by the 23 24 introduction of both oxygen and sulfur-containing groups. Such sulfur moieties 25 (sulfones, sulfides, sulfoxides, and sulfur atoms) may lead to active sites occlusion

1 inside the pores and on the surface of AC, as reported by Khayoon et al. [31]. 2 Treatments with EDA and melamine lead to a drastic reduction of the surface area. This 3 may be due to the presence of numerous groups on the activated carbon surface, which 4 may partially block the access of N₂ molecules to the micropores. However, an 5 appreciable increase of surface area and pore volume can be noticed only for urea 6 treated samples compared to AC1 and AC2. This is probably due to significant carbon 7 gasification and creation of additional microporosity [32]. In general, the textural 8 properties of carbons are highly influenced by the nitrogen containing precursor; 9 specially, for the AC2_Melamine sample, for which the micropore volume decreased up 10 to nearly zero. On the other hand, the impregnation of 3 wt% iron did not substantially 11 change the textural properties of the original carbon material.

12

13 *3.1.2. TPR*

14 TPR profiles of iron catalyst on the different supports are shown in Fig. 1. For 15 comparison purpose, the TPR analyses of selected modified activated carbons were also 16 performed and no additional peaks were observed. As it can be seen from Fig. 1, the 17 position, width and intensity of the peaks mainly depend on the nature of the support. 18 The major reduction peak is at a temperature around 400 °C for original (AC0) and 19 oxidized AC (AC1). The second peak observed at temperature above 550 °C in AC0 20 profile could be related to iron oxide which is strongly interacting with the support. In 21 general, the reduction peak for samples treated with N-containing compounds is shifted 22 to the right irrespective of the starting material (AC1 or AC2). The reduction 23 temperature depends on the degree of interaction between the active species and the 24 support. For all samples treated with N-containing compounds using either AC1 or AC2

as a starting material, the iron reduction peaks shift to higher temperatures. Exceptions
are the samples treated with urea, which present a reduction peak at lower temperature.

3

4 *3.1.3. Surface chemistry characterization*

5 3.1.3.1. TPD and pH_{PZC}

6 TPD analyses were carried out to evaluate the surface chemistry of the different 7 AC supports. The nature of the groups can be assessed by the decomposition 8 temperature and the type of gas released. CO and CO₂ are released by thermal 9 decomposition of oxygen containing groups on the surface of carbon materials at 10 different temperatures [18]. The TPD profiles of the original and modified activated 11 carbons are depicted in Fig. 2. From the TPD profiles, it is possible to identify and 12 quantify the amounts of the oxygenated groups of each sample. CO_2 peaks result from 13 carboxylic acids at low temperatures (150 - 450 °C), or lactones at higher temperatures 14 (600 - 800 °C); carboxylic anhydrides originate both CO and CO₂ (400 - 650 °C); 15 groups such as phenols (600 - 800 °C) and carbonyls/quinones (750 - 1000 °C) 16 originate CO peaks [33].

17 It can be seen that the nitric acid treatment (AC1) increase the amount of 18 oxygenated surface groups, which is evidenced by the increase of CO_2 (Fig. 2a) and CO 19 (Fig. 2b) released. The samples treated with EDA originate CO peaks at lower 20 temperatures (200-400 °C) than the others.

21 The total amounts of CO and CO_2 released from the AC materials were 22 calculated from the corresponding TPD spectra. This information as well as the point of 23 zero charge of the samples with different surface chemistries are shown in Table 4.

As expected, AC1 has a low pH_{PZC} and the highest amount of CO and CO₂, due to the introduction of oxygen-containing surface groups having acidic properties,

1 mainly carboxylic acids. After thionyl chloride treatment (sample AC2), the carboxyl 2 groups were mostly converted into acid chloride groups, resulting a decrease of the 3 amount of CO and CO₂, but this material (AC2) still has acidic properties with the 4 corresponding pH_{PZC} value similar to that of sample AC1. With the exception of 5 AC2_EDA, all materials obtained after the treatment with N-containing precursors 6 using sample AC1 or AC2 as starting materials have approximately neutral or slightly 7 basic properties. This can be explained by the presence of nitrogen groups having basic 8 properties.

9 In general, the TPD profile shows that it is important to have high density of 10 oxygenated groups on the surface of carbon before functionalization with the N-11 containing compounds. The N-containing groups (EDA, urea and melamine) strongly 12 interact with carboxylic acids, anhydrides and lactones [24, 34].

13

14 3.1.3.2. Elemental analysis

The carbon, hydrogen, nitrogen and oxygen contents obtained by elemental analysis are summarized in Table 5. Considering the treatment carried out, it was expected that AC2 and AC2-based samples may contain some sulphur, which is summed with the oxygen content as oxygen determined by difference.

19 The original AC contains a small amount (about 1%) of nitrogen. Significant 20 amounts of nitrogen were introduced on the surface after treatment with N-containing 21 compounds. The highest amount of nitrogen incorporated into the structure was 22 obtained when melamine was the source of nitrogen, as its molecule contains up to six 23 nitrogen atoms.

Thus, the differences in the amounts of nitrogen may be related to the content of nitrogen in the precursors (67% in melamine; 47% in urea and EDA) and the promoting

1 effect of surface acidity, enhanced by oxidation/activation of SOCl₂ on the retention of 2 N-containing organic bases [28,35]. On the other hand, the EDA treated samples 3 contain a relatively high content of hydrogen compared to others; this is probably due to 4 the higher content of hydrogen found in the compound. It is interesting to note that in 5 the urea and EDA treated samples similar total contents of nitrogen are found regardless 6 the pretreatment applied, either using oxidized (AC1) or activated with thionyl chloride 7 (AC2). However, in the case of melamine treated samples, the nitrogen content is higher 8 when the AC2 sample is the starting material in comparison to the inactivated (only 9 oxidized) sample AC1.

10

11 *3.2. Catalytic activity*

Although it is difficult to compare the performance of catalysts having different textural properties and surface chemistries, we performed some tests using the activated carbons as adsorbents or catalysts and the iron supported on modified AC as catalysts for the removal of phenol in solution.

16 The phenol removal on the original and modified activated carbons was studied 17 as a function of time under the following conditions: 100 mL of solution having 150 18 mg/L of phenol, adsorbent/catalyst amount = 1 g/L, initial pH of the solution = 3.0, T = 19 $30 \,^{\circ}$ C, atmospheric pressure and stoichiometric amount of H₂O₂.

20

21 3.2.1. Adsorption experiments

Adsorption experiments were carried out to evaluate the effect of textural and surface properties of the original and modified activated carbons on phenol removal by adsorption. The results of phenol removal in adsorption runs performed with the original and modified activated carbons are compiled in Fig. 3.

1 Even though the difference is not big compared to the original AC, it is clear that 2 urea treated samples (AC1 Urea and AC2 Urea) improve the adsorption capacity of the 3 original AC (AC0). These samples (AC0, AC1_Urea and AC2_Urea) present similar 4 surface areas and pore volumes, their difference being mainly related with their surface 5 chemistry. All the other materials showed lower adsorption capacities compared to the 6 original AC. These adsorption results confirm that the textural properties of the carbon 7 materials have a primary role in adsorption of phenol, irrespective of surface chemistry, 8 but closely following the available surface area.

In the adsorption of aromatic compounds in liquid phase on activated carbons, 9 10 there are mainly two types of interactions: (1) electrostatic and (2) dispersive [36]. The 11 first mechanism is involved when the adsorbate is dissociated under the experimental 12 conditions. For example, if solution $pH > pH_{PZC}$, then the carbon surface is negatively charged, the opposite occurs when the solution $pH < pH_{PZC}$ as described elsewhere [37]. 13 14 In the case of the second mechanism, the existence of π - π dispersion interactions is 15 commonly accepted [36, 38]. Taking this in consideration, at pH of 3.0, phenol 16 $(pK_a=9.89)$ is found in solution predominantly in the molecular form and only the 17 dispersive interactions are most probably involved in its adsorption on the carbon 18 surface [39, 40].

Among the materials tested, the best performance was obtained with AC2_Urea, reaching 67% of phenol removal after 3 h. The BET surface area of this adsorbent is slightly lower and the pH_{PZC} is higher compared to the original AC. The EDA treated samples show moderate adsorption performance having moderate surface area and pH_{PZC} . However, the melamine treated samples show a poor adsorption performance due to the drastic decrease of surface area, though having a higher pH_{PZC} compared to AC1 and AC2 (pH_{PZC} of 2.5).

1 Moreover, the analysis of the results shows that there is a correlation between 2 the phenol removal efficiency by adsorption and the S_{BET} of the tested activated 3 carbons. The correlation is shown in Fig. 4. The figure clearly shows that the phenol 4 removal efficiency increases linearly with the S_{BET}, irrespective of the surface 5 chemistry. This indicates that the surface area of activated carbons is the principal 6 responsible for the phenol adsorption capacity. Fierro and co-workers [41] also 7 observed that the amount of adsorbed phenol in microporous activated carbons 8 increases linearly with the increase in micropore volume.

As mentioned above, the textural and chemical properties of the tested samples 9 10 are different. So, an additional study is needed to take conclusions about the effect of 11 textural properties and surface functionalities. For this purpose, the apparent rate 12 constants for the first-order (k_1) and second-order (k_2) adsorption models were 13 determined for all samples tested. The constants of the two models are listed in Table 6. The ratio between k_1 or k_2 and S_{BET} is presented in the same table. Even though, both 14 15 the first-order and the second-order model fit the experimental data quite well with R^2 16 values close to unity, the second-order model was suitable for the adsorption of lower 17 molecular weight adsorbates on smaller adsorbent particles as reported by Wu et al. [42] 18 and therefore the corresponding kinetic constants are used in the following discussion.

It is reported that phenol adsorption is dependent on both the surface area and the presence of surface groups [41, 43]. In order to evaluate the influence of the nitrogen content, the second-order apparent rate constants (k_2) are normalized by the specific surface area (S_{BET}). It can be observed in Fig. 5 that high normalized adsorption rate constants correspond to samples with high amounts of N-containing surface groups (see Table 5).

1 3.2.2. CWPO experiments

2 3.2.2.1. Modified activated carbons as catalysts

In order to evaluate the ability of the modified activated carbons to act as catalysts (without any supported metal) in the CWPO of phenol, runs were performed under the same conditions as the adsorption tests but now adding H_2O_2 . The corresponding phenol removal curves are shown in Fig. 6.

7 Phenol removal by CWPO in the presence of modified activated carbons as 8 catalysts was similar or slightly higher than those obtained by adsorption. However, the 9 results of phenol removal by adsorption (Fig. 3) and CWPO (Fig. 6) lead to the 10 conclusion that these modified carbon materials are not particularly active for CWPO of 11 phenol. For instance, the phenol removal at 60 min for AC1_Urea is 43% without H_2O_2 12 and 42% in the presence of H_2O_2 , or 65% and 64%, respectively, at 180 min, which 13 suggests that both cases are governed by adsorption. So, the removal of phenol is 14 mainly due to pure adsorption. Similar conclusion was obtained in a previous work for 15 some dyes [44].

16

17 *3.2.2.2. Iron supported catalysts*

In order to study the influence of surface chemical characteristics of the support on the activity of iron catalysts in phenol oxidation, a set of runs were performed in CWPO for different ZVI supported catalysts. Fig. 7 shows the evolution of phenol removal in the presence of the iron catalyst supported on different supports. ZVI/AC2_Melamine and ZVI/AC2_EDA show a good and similar performance, reaching above 80% phenol conversion after 60 min. On the contrary, the phenol removal by pure adsorption and CWPO using the same material without the presence of ZVI (i.e AC2_Melamine and AC2_EDA) shows low removal efficiency below 15 %
 after 60 min.

3 It must be strongly noted that the presence of ZVI on N-containing catalysts 4 yields a better phenol removal, reaching values over 85% after 3 h. These results 5 revealed that there is a direct and positive relation between the catalytic activity and the 6 nitrogen content of the materials. Therefore, the activity of iron supported catalysts 7 based on N-containing AC in CWPO of phenol is enhanced when compared to that of 8 the untreated ones. It is thus observed that the presence of nitrogen groups on the 9 surface of activated carbon, in addition to iron, clearly increases the phenol removal. 10 This synergy can be probably assigned to the ability of retaining the iron ions close to 11 the carbon surface due to the complexing properties of these N-containing groups.

12 Similar trends were observed for TOC removal at the end of adsorption and 13 oxidation tests. Fig. 8 shows that the EDA and melamine treated AC present lower TOC 14 removal performance in the adsorption or oxidation tests when these materials were 15 used as adsorbents or catalysts, respectively. However, the ZVI supported catalysts on 16 the EDA and melamine treated supports also promote the TOC removal. The main 17 reason for this can be explained by the synergetic effect of the presence of nitrogen 18 group and iron on the surface of activated carbon, promoting oxidation instead of 19 adsorption. In this case, the access of H_2O_2 to the iron on the carbon surface could be 20 easier than for the other samples and accelerates the hydroxyl radical generation for 21 deep oxidation of phenol, as observed in Fig. 7. However, additional studies are still 22 required to reveal the exact origin of these effects.

Leaching tests were performed in order to evaluate the stability and contribution of homogeneous reaction for each catalyst after 120 and 180 min of reaction. Leaching of iron corresponds to values below 0.01% of the metal initially present in almost all

1 catalysts; also limited values of 0.03 to 0.04% were measured in the case of AC1 and 2 AC2 supported catalysts. This fact can be related with the pH_{PZC} of the support. In fact, 3 AC1 and AC2 present the lowest pH_{PZC} and a relatively high iron leaching. In 4 conclusion, the contribution of homogeneous system is quite negligible, as the amount 5 of leached iron is quite low.

6 Since the stability of a catalyst is one important aspect in the general evaluation 7 of its performance, the most efficient catalysts of this study (ZVI/AC2 Melamine and 8 ZVI/AC2 EDA) were reused three times in consecutive CWPO reactions (Fig. 9). A slight loss of catalyst activity is observed with both catalysts from the 1st to the 2nd run, 9 but the activity being practically maintained from the 2nd to the 3rd run, and the actual 10 11 removal of phenol is around 80%. The decrease in the removal of phenol observed from 12 the first to the second run may be due to the lower adsorption capacity of the used 13 catalyst and/or the slight loss of the catalyst during the recovery process. In fact, at the 14 end of the experiment, a fraction of organic compounds will remain adsorbed onto the 15 catalyst under the employed operating conditions. These compounds can be either 16 phenol or intermediates formed during the oxidation reaction, as discussed elsewhere 17 [45]. In any case, both catalysts (ZVI/AC2_Melamine and ZVI/AC2_EDA) still present 18 a good efficiency under consecutive runs.

19

20 4. Conclusions

21

Activated carbons with different N-containing precursors were modified, characterized and tested as adsorbents or catalysts for adsorption and peroxide oxidation of phenol. Besides, these modified carbons were also impregnated with iron and used as catalysts for CWPO.

Treatments with EDA and melamine lead to a drastic surface area reduction.
 This may be due to the presence of numerous groups on the activated carbon surface,
 which may partially block the access of N₂ molecules to the micropores.

In the urea and EDA treated samples similar total contents of nitrogen are found regardless of the pretreatment applied. However, in the case of melamine treated samples, the nitrogen content is significantly higher when the oxidized activated carbon treated with thionyl chloride is the starting material.

8 Phenol removal by CWPO in the presence of carbon materials was slightly 9 higher than those obtained by adsorption. However, the results lead to the conclusion 10 that these materials are not particularly active for the reaction, the removal of phenol 11 being mainly due to adsorption.

12 The iron supported catalysts based on N-containing AC show the highest phenol 13 removal efficiency, reaching values over 85% conversion after 3 h. This result revealed 14 that there is a positive relation between the catalytic activity and the nitrogen content of 15 the materials.

16

17 Acknowledgements

18

Financial support for this research was provided by the Spanish Ministerio de Educación y Ciencia and FEDER, projects CTM2008-03338 and CTM2011-23069. The Spanish Ministerio de Ciencia e Innovación is also thanked for providing a doctoral scholarship (Programme FPI, BES-2009-017016) to carry out this research work. The author's research group is recognised by the Comissionat per a Universitats i Recerca del DIUE de la Generalitat de Catalunya (2009SGR865) and supported by the Universitat Rovira i Virgili (2010PFR-URV-B2-41). Funding from FCT and FEDER in

1	the framework of Program COMPETE (FCT grant SFRH/BPD/80435/2011 and Project						
2	Pest-	C/EQB/ LA0020/2011) is acknowledged by the LCM group of authors.					
3							
4	Refe	rences					
5							
6	[1]	G. Annadurai, S.R. Babu, K.P.O. Mahesh, T. Murugesan, Bioprocess Eng., 2					
7		(2000) 493–501.					
8	[2]	P.S. Nayak, B.K. Singh, Desalination 207 (2007) 71–79.					
9	[3]	C.E. Paisio, E. Agostini, P.S. González, M.L. Bertuzzi, J. Hazard. Mater. 167					
10		(2009) 64–68.					
11	[4]	A. Fortuny, C. Bengoa, J. Font, F. Castells, A. Fabregat, Catal. Today 53 (1999)					
12		107–114.					
13	[5]	A. Santos, P. Yustos, A. Quintanilla, S. Rodríguez, F. García-Ochoa, Appl. Catal.					
14		B 39 (2002) 97–113.					
15	[6]	C. Comninellis, A. Kapalka, S. Malato, S.A. Parsons, I. Poulios, D. Mantzavinos,					
16		J Chem Technol Biotechnol, 83 (2008) 769–776.					
17	[7]	D. Mantzavinos, E. Psillakis, J Chem Technol Biotechnol, 79 (2004) 431–454.					
18	[8]	V. García-Molina, M. López-Arias, M. Florczyk, E. Chamarro, S. Esplugas, Water					
19		Res. 39 (2005) 795–802.					
20	[9]	F. Luck, Catal. Today 53 (1999) 81–91.					
21	[10]	H. Debellefontaine, S. Crispel, P. Reilhac, F. Périé, J. Foussard, Chem. Eng. Sci.					
22		54 (1999) 4953-4959.					
23	[11]	M. Klavarioti, D. Mantzavinos, D. Kassinos, Environ. Int. 35 (2009) 402-417.					

- [13] E. Ahumada, H. Lizama, F. Orellana, C. Suárez, A. Huidobro, A. SepúlvedaEscribano, F. Rodríguez-Reinoso, Carbon 40 (2002) 2827–2834.
- 4 [14] F. Rodríguez-Reinoso, Carbon, 36 (1998) 159–175.
- 5 [15] M. Eugeni, F. Stüber, A. Fortuny, A. Fabregat, J. Carrera, J. Font, Appl. Catal. B:
 6 Environ. 58 (2005) 105–114.
- 7 [16] F. Stüber, J. Font., A. Fortuny, C. Bengoa, A. Eftaxias, and A. Fabregat, Top
 8 Catal., 33 (2005) 3–50.
- 9 [17] A. Quintanilla, J.A.Casas., J.J. Rodriguez, M.T. Kreutzer, F. Kapteijn, J.A.
 10 Moulijn, Int. J. Chem. Reactor Eng. 5 (2007) A62.
- [18] J.L. Figueiredo, M.F.R. Pereira, M.M.A. Freitas, J.J.M. Órfão, Carbon 37 (1999)
 12 1379–1389.
- [19] M. Baricot, Seyed A. Dastgheib, Fortuny, F. Stüber, C. Bengoa, A. Fabregat, L.
 Le Coq, J. Font, Carbon, 2007 (16-21 July), Seattle USA.
- 15 [20] N. Klamerth, S. Malato, A. Agüera, A. Fernández-Alba, Water Res. 47 (2013)
 16 833–840.
- 17 [21] Z. Yuan, J.M. VanBriesen, Environ. Eng. Sci. 23 (2006) 533–544.
- 18 [22] J.C. Lennox, R.W. Murray, J. Electroanal. Chem. 78 (1977) 395–401.
- 19 [23] W. Yantasee, Y. Lin, G.E. Fryxell, K.L. Alford, B.J. Busche, C.D. Johnson, Ind.
 20 Eng.Chem. Res. 43 (2004) 2759–2764.
- [24] H. Tamai, K. Shiraki, T. Shiono, H. Yasuda, J. Colloid Interface Sci. 295 (2006)
 22 299–302.

1	[25] A.R. Silva, M. Martins, M.M.A. Freitas, A. Valente, C. Freire, B. De Castro, J.L.
2	Figueiredo, Micropor Mesopor Mat. 55 (2002) 275–284.

- 3 [26] A. Houshmand, W.M.A. Wan Daud, M. Lee, M.S. Shafeeyan, Water Air Soil
 4 Pollut 223 (2012) 827–835.
- 5 [27] J.P.S. Sousa, M.F.R. Pereira, J.L. Figueiredo, Fuel Process Technol 106 (2013)
 6 727–733.
- 7 [28] A. Bagreev, J.A. Menendez, I. Dukhno, Y. Tarasenko, T.J. Bandosz, Carbon 42
 8 (2004) 469–476.
- 9 [29] J.J.M. Órfão, A.I.M. Silva, J.C.V. Pereira, S.A. Barata, I.M. Fonseca, P.C.C.
- 10 Faria ,M.F.R. Pereira, J. Colloid Interface Sci. 296 (2006) 480–489.
- 11 [30] C.Y. Yin, M. Aroua, W. Daud, Sep Purif Technol 52 (2007) 403–415.
- 12 [31] M.S. Khayoon, B.H. Hameed Bioresour Technol. 102 (2011) 9229-9235.
- [32] M. Gurrath, T. Kuretzky, H.P. Boehm, L.B. Okhlopkova, A.S. Lisitsyn, V.A.
 Likholobov, Carbon 38 (2000) 1241–1255.
- [33] J.L. Figueiredo, M.F.R. Pereira, M.A. Freitas, J.J.M. Órfão, Ind Eng Chem Res.,
 46 (2007) 4110–4115.
- 17 [34] Y. El-Sayed and T. J. Bandosz, Langmuir 21 (2005) 1282-1289.
- 18 [35] M. Seredych, D. Hulicova-Jurcakova, G.Q. Lu, T.J. Bandosz, Carbon 46 (2008)
 19 1475–1488.
- 20 [36] C. Moreno- Castilla, Carbon 42 (2004) 83–94.
- [37] L.R. Radovic, I.F. Silva, J.I. Ume, J.A. Menéndez, C.A. Leon y Leon, A.W.
 Scaroni, Carbon 35 (1997) 1339–1348.

- [38] S. Haydar, M.A. Ferro-Garcia, J. Rivera-Utrilla, J.P. Joly, Carbon 41(2003), 387–
 395.
- 3 [39] F. Villacañas, M.F.R. Pereira, J.J.M. Órfão, J.L. Figueiredo, J. Colloid Interface
 4 Sci. 293 (2006) 128–136.
- 5 [40] O. Taran, E. Polyanskaya, O. Ogorodnikova, V. Kuznetsov, V. Parmon, M.
 6 Besson, C. Descorme, Appl. Catal., A, 387 (2010) 55–66.
- 7 [41] V. Fierro, V. Torne-Fernandez, D. Montane, A. Celzard, Micropor Mesopor Mat.
 8 111 (2008) 276–284.
- 9 [42] F.C. Wu, R.L. Tseng, S.C. Huang, R.S. Juang, Chem. Eng. J. 151 (2009) 1–9.
- [43] G.G. Stavropoulos, P. Samaras, G.P. Sakellaropoulos, J. Hazard. Mater. 151
 (2008) 414–421.
- [44] V. P. Santos, M. F. R. Pereira, P. C. C. Faria and J. J. M. Órfão, J. Hazard. Mater.
 162 (2009) 736–742.
- [45] J.A. Melero, G. Calleja, F. Martínez, R. Molina, M.I. Pariente, Nanocomposite
 Fe₂O₃/SBA-15: Chem. Eng. J. 131 (2007) 245–256.

N-containing	Molecular	Molar mass	Chemical structure	N content
compounds	formula	$(g mol^{-1})$		(wt. %)
Ethylenediamine	$C_2H_8N_2$	60.10	H ₂ NNH ₂	46.59
Urea	CH ₄ N ₂ O	60.06	H ₂ N ^C NH ₂	46.62
Melamine	C ₃ H ₆ N ₆	126.12	NH2 NNH2 H2NNNH2	66.60

Table 1. Chemical properties of the N-containing compounds used

Abbreviation	Description
AC0	Untreated AC
AC1	HNO ₃ -treated AC0
AC2	SOCl ₂ -treated AC1
AC1_EDA	EDA-treated AC1
AC2_EDA	EDA-treated AC2
AC1_Urea	Urea- treated AC1
AC2_Urea	Urea- treated AC2
AC1_Melamine	Melamine- treated AC1
AC2_Melamine	Melamine- treated AC2
ZVI/ACx_y	Iron impregnated on sample ACx_y

Table 2. Abbreviations and description of original, modified and iron supported

activated carbons

x_y: represents the modification undergone on activated carbon

Sample	$S_{BET}(\pm 10 \text{ m}^2 \text{ g}^{-1})$	S_{meso} (±10 m ² g ⁻¹)	V_{micro} (±0.01 cm ³ g ⁻¹)
AC0	984	207	0.351
AC1	853	141	0.342
AC2	635	178	0.165
AC1_EDA	335	148	0.086
AC2_EDA	242	175	0.030
AC1_Urea	949	180	0.339
AC2_Urea	939	200	0.342
AC1_Melamine	194	111	0.038
AC2_Melamine	47	47	0.000
ZVI/AC0	975	216	0.348

 Table 3. Textural properties of samples

Sample	CO ($\pm 20 \ \mu mol \ g^{-1}$)	$CO_2 (\pm 20 \ \mu mol \ g^{-1})$	pH _{PZC} (±0.1)	
AC0	598	164	6.7	
AC1	2311	754	2.5	
AC2	1660	565	2.4	
AC1_EDA	927	150	6.8	
AC2_EDA	616	102	5.3	
AC1_Urea	1481	193	7.6	
AC2_Urea	1096	118	7.9	
AC1_Melamine	1305	175	6.2	
AC2_Melamine	739	102	6.7	

Table 4. Surface chemistry data of activated carbon supports (total amounts of CO and CO_2 released, and pH_{PZC}).

Sample	C (wt%)	H(wt%)	N (wt%)	O ^a (wt%)
AC0	90.95	0.92	0.92	7.66
AC1	85.79	0.18	1.80	12.24
AC2	84.43	1.36	1.23	12.97
AC1_EDA	83.00	1.88	5.74	9.38
AC2_EDA	81.49	1.99	5.53	10.99
AC1_Urea	85.24	0.30	3.20	11.26
AC2_Urea	88.38	0.57	2.38	8.68
AC1_Melamine	81.69	0.74	15.95	1.62
AC2_Melamine	74.06	0.98	22.07	2.90

Table 5. Elemental analysis of the original and modified AC

^aOxygen determined by difference.

Sample	Pseudo-fi	rst-order	er Pseudo-second-order			er		
	qe	k ₁	\mathbf{R}^2	10^{3}	qe	$10^3 k_2$	R^2	10^{6}
	(mg/L)	(min ⁻¹)		$k_{\rm l}/S_{\rm BET}$	(mg/L)	$(g_{AC}mg^{-1}min^{-1})$		$k_2\!/S_{BET}$
AC0	80.87	0.017	0.988	0.017	104.10	0.268	0.981	0.272
AC1	54.52	0.011	0.891	0.013	73.52	0.336	0.951	0.394
AC2	37.35	0.014	0.915	0.022	52.08	0.706	0.983	1.111
AC1-EDA	25.27	0.016	0.978	0.047	33.89	1.060	0.982	3.164
AC2-EDA	36.29	0.015	0.987	0.062	47.39	0.522	0.972	2.157
AC1-Urea	84.76	0.020	0.992	0.021	108.69	0.228	0.995	0.240
AC2-Urea	83.62	0.021	0.995	0.022	107.52	0.309	0.996	0.329
AC1-Melamine	19.13	0.017	0.985	0.087	25.57	1.410	0.987	7.268
AC2-Melamine	3.39	0.028	0.965	0.596	4.65	1.640	0.995	34.890

 Table 6. Kinetics parameters of the first-order and second-order adsorption models

FIGURE CAPTIONS

Fig. 1. TPR profiles of supported iron catalysts using AC1 (a) and AC2 (b) as starting material.

Fig. 2. TPD spectra of the activated carbon supports (a) and (c) CO; (b) and (d) CO₂.

Fig. 3. Phenol removal obtained in pure adsorption experiments (150 mg/L of phenol, pH 3.0, T = 30 °C and adsorbent load = 1 g/L) with different carbon supports.

Fig. 4. Correlation between specific surface area and phenol removal obtained in pure adsorption for original and treated activated carbons.

Fig. 5. Plot of the normalized rate constants for phenol adsorption (k_2/S_{BET}) versus N-content for original and treated activated carbons.

Fig. 6. Phenol removal obtained in CWPO experiments (150 mg/L of phenol, H_2O_2 concentration of 750 mg/L, pH 3.0, T = 30 °C and catalyst load = 1 g/L) with different activated carbons.

Fig. 7. Phenol removal obtained in CWPO experiments (150 mg/L of phenol, H_2O_2 concentration of 750 mg/L, pH 3.0, T = 30 °C and catalyst load = 1 g/L) with ZVI supported on different activated carbons.

Fig. 8. TOC removal after 180 min obtained in adsorption and CWPO experiments (150 mg/L of phenol, H_2O_2 concentration of 750 mg/L, pH 3.0, T = 30 °C and

adsorbent/catalyst load = 1 g/L) with carbon samples and ZVI supported on activated carbons.

Fig. 9. Phenol removal obtained by CWPO in sequential experiments (150 mg/L of phenol, H_2O_2 concentration of 750 mg/L, pH 3.0, T = 30 °C and catalyst load = 1 g/L) using ZVI/AC_EDA and ZVI/AC2_Melamine catalysts.





Fig. 1.



Fig. 2.



Fig. 3.


Fig. 4.



Fig. 5.



Fig. 6.



Fig. 7.



Fig. 8.



Fig. 9.



Highlights

- Activated carbons were successfully modified with different N-containing precursors.
- Urea and EDA treated samples have similar total contents of nitrogen regardless of the pretreatment applied.
- The nitrogen content of melamine treated samples is significantly higher.
- The iron supported catalysts based on N-containing AC show the highest phenol removal efficiency.