

Combining cationic and anionic mixed-mode sorbents in a single cartridge to extract basic and acidic pharmaceuticals simultaneously from environmental waters

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Abstract The aim of the present study is to broaden the applications of mixed-mode ion-exchange solid-phase extraction sorbents to extract both basic and acidic compounds simultaneously by combining the sorbents in a single cartridge and developing a simplified extraction procedure. Four different cartridges containing negative and positive charges in the same configuration were evaluated and compared to extract a group of basic, neutral, and acidic pharmaceuticals selected as model compounds. After a thoroughly optimization of the extraction conditions, the four different cartridges showed to be capable of retaining basic and acidic pharmaceuticals simultaneously through ionic interactions, allowing the introduction of a washing step with 15 mL methanol to eliminate interferences retained by hydrophobic interactions. Using the best combined cartridge, a method was developed, validated, and further applied to environmental waters to demonstrate that the method is promising for the extraction of basic and acidic compounds from very complex samples.

Keywords Cationic mixed-mode solid-phase extraction · Anionic mixed-mode solid-phase extraction · Sorbent combination · High-resolution mass spectrometry · Environmental waters

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Introduction

Solid-phase extraction (SPE) remains the most commonly used sample preparation technique for liquid samples in several analytical methods. Its wide acceptance is the result of the great advantages that it provides, such as the enrichment of the analytes with high recoveries and the enhanced selectivity thanks to the availability of different types of sorbents [1]. SPE is often combined with chromatographic techniques such as liquid chromatography (LC) and gas chromatography (GC) coupled with mass spectrometry (MS) detection, obtaining methods with high sensitivity and selectivity to determine several target compounds in complex matrices at trace levels [2–5].

Among the different sorbents available for SPE, mixed-mode sorbents combine a polymeric structure with one type of ionic functional groups (cationic or anionic) in each cartridge, giving them the capability of retaining compounds through reversed-phase and ion-exchange interactions [6]. They were developed to promote selectivity for ionic compounds, broadening the groups of analytes that can be retained by a single sorbent [7–9]. The most important feature of mixed-mode sorbents is the possibility of including a washing step with organic solvents in the SPE procedure, which allows the elimination of interferences retained by hydrophobic interactions. In several studies, the potential of these sorbents has been proven to reduce the matrix effect (ME) in LC-MS-based methods, by eliminating interferences during washing steps in the SPE procedure [10–13].

There are four types of mixed-mode sorbents depending on the functional groups attached to the polymer particles: strong or weak cationic-exchangers (SCX or WCX) and strong or weak anionic-exchangers (SAX or WAX). The sorbents with strong properties include functional groups that are charged in the entire pH range (such as sulfonic acid or quaternary

67 ammonium groups), while those with weak properties have
68 groups with a more reduced working pH range depending on
69 their pK_a values (carboxylic acid or tertiary or secondary
70 amine groups). When working with any of these sorbents,
71 the selection of the SPE conditions is very important, as pa-
72 rameters such as pH or type of clean-up and elution solvent
73 have a great impact on the performance of the sorbents [6]. By
74 fine-tuning the SPE protocol, these sorbents can extract an
75 extensive range of compounds with different properties
76 (non-polar, polar, or ionic) or they can be selective towards
77 ionic compounds if appropriate washing steps are applied.

78 Successful applications of mixed-mode sorbents commer-
79 cially available or prepared in-house with strong or weak
80 properties have been reported, describing the extraction of
81 different groups of compounds from complex matrices
82 [14–19]. Analytes such as pharmaceuticals, drugs of abuse,
83 and compounds of biological interest have been extracted
84 from environmental waters, foodstuff, and biologic fluids
85 [20–23]. In most of the studies published to date, the potential
86 of mixed-mode sorbents is limited to basic or acidic analytes
87 depending on the type of ion-exchanger selected. Only a few
88 studies have explored the result of combining both types of
89 sorbents to extract a whole list of target compounds with both
90 basic and acidic properties. Lavén et al. [24] developed an
91 SPE procedure in which a SCX sorbent was placed in tandem
92 with an SAX sorbent to extract 15 basic, neutral, and acidic
93 pharmaceuticals simultaneously. More recently, Deeb et al.
94 [25] proved that a tandem combination of SCX and SAX
95 sorbents gave the highest recoveries in ultrapure water when
96 compared to several SPE sorbents. However, the procedures
97 of these methods are more complicated than those using a
98 single cartridge. Besides, in cartridges coupled in tandem,
99 the ionic interactions of cations and anions are not established
100 simultaneously but firstly in the first cartridge and later in the
101 second cartridge during the wash. Therefore, pH control of
102 both processes is more difficult. The aim of the present study
103 is to broaden the use of mixed-mode sorbents by combining
104 for the first time mixed-mode sorbents with strong or weak
105 cationic and anionic properties to obtain single SPE cartridges
106 with the purpose of extracting selectively and simultaneously
107 basic and acidic compounds. Cation- and anion-exchangers
108 have been combined in the same cartridge before [26] but with
109 the sole purpose of cleaning interferences from the sample and
110 not ionically retaining basic and acidic analytes at the same
111 time. When combining the sorbents in a single cartridge, the
112 SPE procedure is simplified as long as the SPE conditions are
113 correctly selected. With this in mind, a systematic evaluation
114 was performed of combinations of the four types of mixed-
115 mode sorbents (SCX, WCX, SAX, and WAX) by pairs with
116 opposite charges for the selective extraction of ionizable phar-
117 maceuticals. The influence of strong or weak functional
118 groups in the sorbents was observed, as well as the effect of
119 changing the pH and other parameters on the performance of

the extraction. This is the first time that the four possible
combinations of available mixed-mode ion-exchangers are
studied in detail for different parameters in order to design a
simple load-wash-elution SPE protocol for cartridges with
zwitterionic character. Subsequently, the best combination of
sorbents was evaluated in environmental waters in terms of
recoveries and ME using LC-high-resolution mass spectrom-
etry (HRMS).

Experimental

Reagents and standards

The pharmaceuticals or metabolites with basic properties, atenolol (ATE), ranitidine (RAN), trimethoprim (TRI), metoprolol (MET) and propranolol (PROP); as well as the neutral pharmaceuticals, caffeine (CAFF), antipyrine (ANTI) and carbamazepine (CBZ); and those with acidic properties, salicylic (SAL AC) and clofibric acid (CLO AC), fenoprofen (FEN), diclofenac (DICLO) and ibuprofen (IBP), were purchased as pure standards from Sigma-Aldrich (St. Louis, MO, USA). The structures, pK_a values, and exact masses of all of the analytes are shown in Table S1 in the Electronic Supplementary Material (ESM). Solid standards were dissolved in methanol (MeOH) to prepare stock solutions of 1 mg/mL which were stored at $-20\text{ }^\circ\text{C}$. Working solutions with the mixture of all the pharmaceuticals were prepared in ultrapure water every week and stored at $4\text{ }^\circ\text{C}$.

Ultra-gradient HPLC-grade MeOH and acetonitrile (ACN) were obtained from J.T. Baker (Deventer, the Netherlands), while ultrapure water was obtained from a water purification system (Veolia, Sant Cugat del Vallès, Spain). Acetic acid (CH_3COOH) and formic acid (HCOOH) were purchased from SDS (Peypin, France) and Sigma-Aldrich, respectively. Ammonium hydroxide (NH_4OH) was obtained from Panreac (Barcelona, Spain).

Sampling

River water samples were collected from the River Ebre in Catalonia, while influent and effluent wastewater samples were collected from sewage treatment plants located in Tarragona and Reus. Both treatment plants include primary and secondary treatments in their processes. Once the samples were collected in pre-cleaned bottles, they were stored at $-20\text{ }^\circ\text{C}$ until the day of analysis. Before any SPE procedure, the samples were filtered through a $1.2\text{-}\mu\text{m}$ glass-fiber membrane filter (Fisherbrand, Loughborough, UK) and then through a $0.22\text{-}\mu\text{m}$ nylon membrane filter (Scharlab, Barcelona, Spain).

165	Solid-phase extraction procedure	
166	Four different 100 mg cartridges containing combinations of	215
167	individual sorbents with different functional groups (Table 1)	216
168	were evaluated to extract basic and acidic pharmaceuticals	217
169	simultaneously. The sorbents combined in the cartridges were	218
170	the commercially available Oasis MCX, Oasis MAX, Oasis	219
171	WCX, and Oasis WAX from Waters Corporation (Milford,	220
172	MA, USA). For simplicity, they will be referred to as SCX,	221
173	SAX, WCX, and WAX from now on. The amount in grams of	222
174	each individual sorbent used for each cartridge was	223
175	established to obtain balanced cationic and anionic moieties.	224
176	The configuration that gave the best results was the strong	225
177	cationic/strong anionic one (SCX/SAX). The protocol was	226
178	then transferred to 500 mg cartridges (110 mg of SCX and	227
179	390 mg of SAX) of this combination to extract the environ-	228
180	mental waters. These cartridges were conditioned with 10 mL	229
181	of MeOH, followed by 10 mL of ultrapure water adjusted to	230
182	pH 7. The selected loading volumes were 100 mL for river	231
183	water and effluent wastewater samples and 50 mL for influent	232
184	wastewater samples. All samples were adjusted to pH 7 using	233
185	either NH ₄ OH or HCOOH depending on the original pH of	234
186	the raw water. After loading the samples, a washing step was	235
187	introduced consisting of 15 mL of MeOH. The elution was	236
188	performed in two subsequent steps: (1) 5 mL of a 10%	237
189	HCOOH in MeOH solution; and (2) 5 mL of a 5% NH ₄ OH	238
190	in MeOH solution. Both fractions were collected in the same	239
191	vial and the extract was evaporated to dryness using a centrif-	240
192	ugal evaporator miVac Duo (Genevac, Ipswich, UK) and later	241
193	reconstituted with 1 mL of ultrapure water for river water sam-	242
194	ples and 2 mL for effluent and influent wastewater samples.	
195	LC-HRMS	
196	A Thermo Scientific Accela 1250 UHPLC system (Bremen,	243
197	Germany) equipped with an Accela Autosampler automatic	244
198	injector and an Accela 1250 pump was coupled with a	245
199	Thermo Scientific Exactive Orbitrap™ mass spectrometer	246
200	for the chromatographic analysis. The mass spectrometer	247
201	worked with a heated electrospray ionization (HESI) source	248
202	and a higher energy collisional dissociation (HCD) cell. The	249
203	chromatographic column used was the Ascentis Express C ₁₈	250
204	(100 mm × 2.1 mm i.d., 2.7 μm particle size) supplied by	251
205	Supelco (Sigma-Aldrich), and the mobile phase was a mixture	252
206	of solvent A (0.5% CH ₃ COOH in H ₂ O) and solvent B	253
207	(MeOH). The optimal pH for the separation of the analytes	254
208	was 2.8. The column was kept at 25 °C and the mobile phase	255
209	was pumped at 400 μL/min. The injection volume used was	256
210	25 μL and the tray of vials inside the automatic injector was	257
211	kept at 10 °C. The optimal gradient profile started with 2% of	258
212	solvent B which was increased to 30% within 6 min and then	259
213	increased again to 80% within 6 min and held for a further	260
214	1 min. After this, solvent B was increased to 100% within	261
	2 min and left isocratic for 3 min and later returned to the	262
	initial conditions within 2 min.	263
	In the ion source, basic pharmaceuticals (ESM Table S1)	
	were ionized in the positive mode using the following optimal	
	parameters: spray voltage, 2 kV; skimmer voltage, 25 V; cap-	
	illary voltage, 40 V; and tube lens voltage, 80 V. In the case of	
	acidic pharmaceuticals, the negative ionization mode was	
	used and the optimal parameters were as follows: spray volt-	
	age, 3.5 kV; skimmer voltage, - 15 V; capillary voltage,	
	- 15 V; and tube lens voltage, - 80 V. Gas flow rates and	
	temperatures were the same for both ionization modes: sheath	
	gas, 40 AU (adimensional units); auxiliary gas, 5 AU; heater	
	and capillary temperature, 350 °C; and probe position adjust-	
	ment: side to side, 0, vertical C and micrometer, 0.5.	
	Four time windows were used to acquire the data: the first	
	and third (0 to 7.5 min and 9.5 to 11 min) were set in positive	
	mode, the second (7.5 to 9.5 min) in both modes, and the last	
	(11 to 20 min) in negative mode. In all of the windows, two	
	scan events were used for each ionization mode, correspond-	
	ing to a full scan (at 50,000 FWHM with 250 ms of injection	
	time), which was alternated with a fragmentation scan (at	
	10,000 FWHM with 50 ms of injection time). Because the	
	second window operated in both positive and negative modes,	
	four scan events were used. The optimal voltage in the HCD	
	cell selected in all the fragmentation scans was 25 eV. For	
	quantification, the response of the molecular ions was used	
	and, for confirmation, the presence of the most abundant frag-	
	ment ions and the corresponding ion ratios were considered.	
	Results and discussion	
	Optimization of LC-HRMS conditions	
	There are several studies in the literature describing the chro-	
	matographic separation of the group of pharmaceuticals se-	
	lected in the present study, where the use of a C ₁₈ stationary	
	phase is quite common [27–29]. For this study, the Ascentis	
	Express C ₁₈ (100 mm × 2.1 mm i.d., 2.7 μm particle size) was	
	compared to the Ascentis RP-Amide (100 mm × 2.1 mm i.d.,	
	2.7 μm particle size) which has proven to offer better retention	
	for polar compounds [30]. The mobile phase was optimized	
	with respect to the organic solvent (ACN or MeOH) in each	
	stationary phase and the type of acid added to the aqueous	
	phase (HCOOH or CH ₃ COOH). As relevant observations, it	
	can be said that CH ₃ COOH was chosen over HCOOH be-	
	cause it offered better ionization for FEN and IBP in the ion	
	source. The best results were obtained using ACN in the RP-	
	Amide phase and MeOH in the C ₁₈ phase. However, in the	
	RP-Amide phase, the first eluting compounds eluted near the	
	void volume, for which the C ₁₈ column was selected, using	
	MeOH as the organic solvent of the mobile phase (“LC-	
	HRMS” section).	

Table 1 Configurations of the four cartridges prepared from the individual mixed-mode sorbents

Individual sorbent	Ionic functional group	meq/g	Configuration	mg	meq
SCX (Oasis MCX)	Sulfonic acid	1	(1) SCX/SAX	22 SCX 78 SAX	0.0220 SCX 0.0195 SAX
SAX (Oasis MAX)	Dimethylbutylamine	0.25	(2) SAX/WCX	76 SAX 24 WCX	0.0190 SAX 0.0180 WCX
WCX (Oasis WCX)	Carboxylic acid	0.75	(3) SCX/WAX	38 SCX 62 WAX	0.0380 SCX 0.0372 WAX
WAX (Oasis WAX)	Piperazine	0.6	(4) WCX/WAX	44 WCX 56 WAX	0.0330 WCX 0.0336 WAX

264 For optimization of the HRMS conditions, the voltages, gas
 265 flow rates, temperatures, and the position of the ionization
 266 probe were varied until obtaining a compromise of the highest
 267 response for all the compounds. As expected, basic and neutral
 268 pharmaceuticals (ATE, RAN, TRI, CAFF, MET, ANTI, PROP,
 269 and CBZ) showed the highest response in the positive mode,
 270 while acidic pharmaceuticals (SAL AC, CLO AC, FEN,
 271 DICLO, and IBP) were better ionized in the negative mode.

272 Optimum fragmentation energy was selected as a compro-
 273 mise of the voltage at which the highest response was obtained
 274 for all fragments. The exact mass of the molecular ions and the
 275 selected fragments for each analyte are shown in Table S1 (see
 276 ESM) and they are according to the literature [31–34]. The
 277 fragments that displayed the highest response were considered
 278 to determine the instrumental limits of detection (ILODs),
 279 which were the concentrations at which the peak correspond-
 280 ing to the fragment showed a signal to noise ratio (S/N) of 3,
 281 or the signal was higher than 1×10^3 for the analytes with no
 282 noise. The instrumental limit of quantification (ILOQ) was the
 283 concentration corresponding to the first point of the calibra-
 284 tion curve. The observed ILODs ranged between 0.05 and
 285 2.5 $\mu\text{g/L}$, while the ILOQs ranged between 0.2 and
 286 2.5 $\mu\text{g/L}$, with FEN, DICLO, and IBP being the analytes with
 287 higher limits, due to their lower response. Linearity was eval-
 288 uated from 0.2 to 1000 $\mu\text{g/L}$, with lack of linearity being
 289 observed for the entire range. Therefore, low level and high
 290 level calibration curves were constructed for most of the
 291 analytes. Each analyte showed different ranges of linearity
 292 but, in general, low level calibration curves were between
 293 0.2 and 100 $\mu\text{g/L}$, while high level calibration curves were
 294 between 25 and 1000 $\mu\text{g/L}$.

295 **Solid-phase extraction**

296 *Behavior of the four cationic/anionic combinations*

297 The individual commercial sorbents Oasis MCX, Oasis
 298 MAX, Oasis WCX, and Oasis WAX were combined to obtain
 299 four configurations such that the ratio of cationic-exchanger to
 300 anionic-exchanger was 1:1 in terms of their ion-exchange

301 capacity. As can be seen in Table 1, the result was four differ-
 302 ent types of cartridges with cationic/anionic functionalities
 303 and strong/strong, strong/weak, or weak/weak properties
 304 (SCX/SAX, SCX/WAX, WCX/SAX, and WCX/WAX).

305 Careful attention was paid to the pH values of the loading
 306 and the elution step, taking into account the pK_a of both the
 307 functional groups of the sorbents and the pK_a of the analytes.
 308 In this sense, when loading the sample, both the functional
 309 groups of the sorbents and the analytes must be in their ionic
 310 form in order to establish ionic interactions [6, 35, 36]. In
 311 contrast, in the elution step, either the functional groups of
 312 the sorbents and/or the analytes must be in their neutral form
 313 to disrupt retention and favor the elution of the analytes.

314 For instance, a sample can be loaded in the SAX/SCX
 315 configuration using a pH value between 5 and 8 (Fig. 1) be-
 316 cause basic analytes are charged up to $\sim \text{pH } 8$ and acidic
 317 analytes are charged from $\sim \text{pH } 3$ or 5 (see pK_a values in
 318 ESM Table S1). The pK_a values of the moieties of the sorbents
 319 were not taken into account in this case because strong func-
 320 tionalities are always charged throughout the pH range.
 321 Therefore, the pH value for the elution was selected such that
 322 the analytes are converted into their neutral form. Acidic phar-
 323 maceuticals were eluted using an acidic solution while basic
 324 pharmaceuticals were collected using a basic solution.

325 When weak moieties are included in the configuration of
 326 the cartridges, the pK_a of the functional groups attached to the
 327 sorbents must be considered. In the case of the SCX/WAX
 328 configuration, the pK_a of the piperazine group attached to
 329 the polymer (WAX) is ~ 6 , as stated by the manufacturer.
 330 Therefore, the optimal loading pH range is between 5 and 6
 331 (Fig. 1), considering the pK_a values of the analytes and the
 332 sorbents. Furthermore, all of the analytes should elute in a
 333 single fraction when using a basic solution because basic phar-
 334 maceuticals would convert into their neutral form, as well as
 335 the piperazine groups in the polymer, disrupting their interac-
 336 tions with the acidic pharmaceuticals.

337 As a result of the previous considerations, the SPE proce-
 338 dure initially used for the four configurations is described
 339 below. A volume of 50 mL of ultrapure water adjusted to
 340 pH 5 using HCOOH spiked with a mixture of the analytes

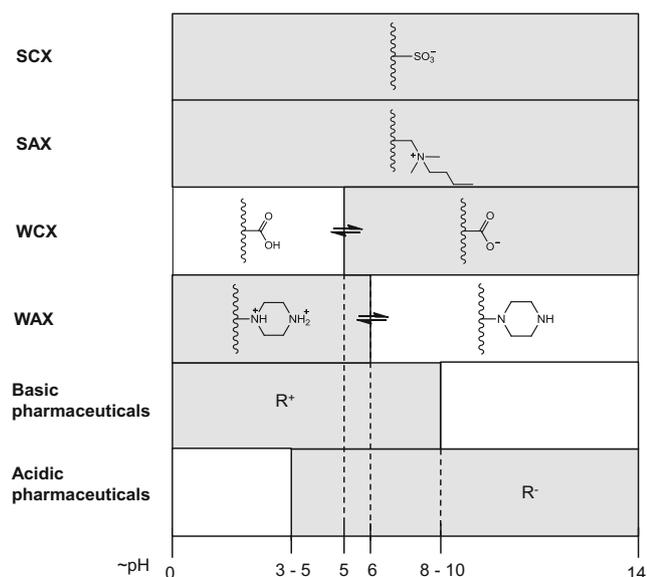


Fig. 1 Charge state of functional groups of the sorbents and the analytes along the pH range (indicated in gray when the charge state is ionic)

341 (75 µg/L) was percolated through the cartridges after condi-
 342 tioning. The washing step introduced consisted of two frac-
 343 tions of 1 mL of MeOH each. Finally, the elution step was
 344 performed in two steps: (1) 5 mL of a 5% HCOOH solution
 345 and, (2) 5 mL of a 5% NH₄OH solution for all the cartridges,
 346 except for the SCX/WAX configuration, for which the elution
 347 steps were interchanged (the basic elution was performed be-
 348 fore the acidic elution). In this case, the elution steps were
 349 interchanged to promote the elution of all the charged analytes
 350 in the first fraction. During the present evaluation of the car-
 351 tridges, all SPE fractions were diluted with ultrapure water (to
 352 5 mL the washing fraction and to 25 mL the elution one) and
 353 analyzed separately in order to evaluate possible losses of the
 354 analytes.

355 From the results detailed in Table 2, it can be seen that all of
 356 the analytes were retained by the cartridges during the loading
 357 step, which was expected, as the combinations of the sorbents
 358 should have the capability of retaining basic, acidic, and neu-
 359 tral compounds. During the washing step, only CAFF, ANTI,
 360 and CBZ were completely rinsed, which is in line with their
 361 pK_a values, as they are neutral compounds. Their retention in
 362 cation-exchangers has been proven [37], behavior that could
 363 be explained by their ability to develop partial charges through
 364 electron delocalization. However, in a cartridge in which neg-
 365 ative charges are coexisting with positive charges, this delo-
 366 calization might be compromised and so the analytes end up
 367 behaving as neutral compounds rather than weak bases.

368 The rest of the pharmaceuticals were isolated during the
 369 elution step, suggesting that these compounds with basic and
 370 acidic properties establish strong ionic interactions with the
 371 sorbents in the prepared cartridges. Furthermore, the analytes
 372 were isolated in the acidic or the basic elution fractions exactly

as predicted by their pK_a values (Table 2). For instance, in the
 SAX/SCX configuration, the acidic pharmaceuticals eluted in
 the acidic elution, while the basic pharmaceuticals eluted in
 the basic elution. In the case of the SCX/WAX configuration,
 all of the analytes were isolated during the basic elution. No
 differences were observed in the retention of the selected
 analytes with regard to the strong and weak character of the
 sorbents included in the cartridges, showing that all of the
 combinations of the sorbents work well as long as the SPE
 conditions are properly chosen.

It must be highlighted that, when combining sorbents using
 the same weight rather than the same ion exchange capacity
 (meq), the charge in excess had an effect on the analytes with
 the same charge, showing, for instance, how acids were par-
 tially lost in the washing fraction (data not shown) when the
 negative charge was predominant in the cartridge. In summa-
 ry, the potential of applying combinations of mixed-mode
 sorbents to extract basic and acidic analytes simultaneously
 was demonstrated, using a simple SPE procedure with the
 advantage of eliminating neutral interferences in the washing
 step.

Because in the SAX/WCX cartridge sometimes the basic
 pharmaceuticals were partially lost in the wash, and the WAX/
 WCX cartridge has a very narrow range of pH at which both
 the analytes and sorbents are charged, the most promising
 combinations were the SCX/SAX and SCX/WAX cartridges.
 These two configurations were used to evaluate if the prelimi-
 nary conditions predicted as most favorable were in fact opti-
 mal, by studying the effect of changing the SPE parameters
 in the performance of the extraction. In addition, the use of
 conditions that, in theory, should not work properly will fur-
 ther confirm the presence of ionic interactions as the driving
 force of the retention of the charged analytes.

Influence of pH on retention

The effect of changing the pH of the loading was studied in
 both the SCX/SAX and the SCX/WAX cartridges, as these
 parameters are very important in the overall procedure.
 Initially, different pH values (2, 7, 10, and 12) were tested in
 the loading step, following the same SPE procedure described
 in the previous section, for both the SCX/SAX and SCX/
 WAX cartridges. In theory, for the SCX/SAX cartridge, the
 optimal loading pH range is ~ 5 to 8, because acidic pharma-
 ceuticals are in their neutral form at pH 2 and basic analytes
 are uncharged at pH 10 and 12 (Fig. 1). Table 3 shows how
 acidic pharmaceuticals were lost in the washing step when
 loading at pH 2, while basic ones were lost in the washing
 step when loading at pH 12, as expected. In this table, the
 recoveries obtained for pH 2 and 12 are shown, as these were
 the pH conditions under which ionic interactions were weak-
 ened. It was also expected that, when loading at pH 5 and 7, all
 pharmaceuticals (with the exception of the neutral ones) were

Table 2 Recoveries (%) obtained when 50 mL of ultrapure water adjusted at pH 5 was percolated through the four cartridge combinations

	Analyte	SCX/SAX				SAX/WCX				SCX/WAX				WCX/WAX				
		w1	w2	e _a	e _b	w1	w2	e _a	e _b	w1	w2	e _b	e _a	w1	w2	e _a	e _b	
Basic	ATE			76	76			85	85			81	81			82	82	
	RAN			79	79			64	64			80	80			56	56	
	TRI			82	82			85	85			81	81			85	85	
	MET			76	76			85	85			80	80	2		83	85	
	PROP			82	82			92	92			86	86			91	91	
Neutral	CAFF	76	19			94	88	11		100	60	35	5	100	70	30	100	
	ANTI	63	19			82	79	10		89	18	23	50	92	59	28	87	
	CBZ	65	21			86	70	20		90	50	35		85	52	33	2	87
Acidic	SAL AC			35	25	59		4	17	21			90	90			19	19
	CLO AC			88	19	107		91		91			85	85		85	85	
	FEN			91		91		91		91			89	89		91	91	
	DICLO			80		80		80		80			83	83		73	73	
	IBP			92		92		97		97	1	2	83	86	6	93	99	

%RDS lower than 15% in all cases

w1 and w2 wash fraction 1 and 2, e_a acidic elution, e_b basic elution, w1 1 mL MeOH, w2 1 mL MeOH, e_a 5 mL 5% HCOOH in MeOH, e_b 5 mL 5% NH₄OH in MeOH, Σ total sum of the recoveries of the four fractions

424 retained through ionic interactions. At pH 10, basic pharma- 454
 425 ceuticals were partially lost in the washing step (7 to 23%) 455
 426 rather than being completely lost, as expected. This was not 456
 427 surprising, as pH 10 is around the pK_a values of these 457
 428 analytes, so the test was performed at the limit of the conver- 458
 429 sion between the ionic and the neutral form.

430 For the SCX/WAX cartridge, the optimal loading pH range
 431 is ~ 5 to 6, because acidic pharmaceuticals are in their neutral
 432 form at pH 2 and the piperazine group of the WAX sorbent is
 433 in its neutral form from pH ~ 6 upwards, as stated by the
 434 manufacturer of the sorbents (Fig. 1). In Table 3, it can be
 435 seen that both acidic and basic pharmaceuticals were lost in
 436 the washing step when loading the sample at pH 12, while, at
 437 pH 2, it was the acidic analytes that were isolated in this step,
 438 just as expected. As anticipated at pH 7 (data not shown), the
 439 acidic pharmaceuticals were strongly retained during the
 440 MeOH wash, rather than losses being observed. Actually,
 441 acidic pharmaceuticals are partially lost starting from pH 10,
 442 suggesting that the piperazine group might have a higher pK_a
 443 within the polymeric network. At this pH, basic pharmaceu-
 444 ticals were also ionically retained, which is explained by the
 445 fact that these conditions are at the limit of the pK_a values of
 446 these analytes, just as observed before for the SCX/SAX car-
 447 tridge. Low recoveries observed for SAL AC during these
 448 tests was explained by excessive retention on the cartridges,
 449 which was resolved by raising the % HCOOH in MeOH from
 450 5 to 10% for the elution in further tests.

451 In this section, it was confirmed that the strong retention
 452 observed for charged pharmaceuticals was due to the estab-
 453 lishment of ionic interactions between the analytes and the

charged functional groups of the sorbents. These interactions
 are only possible in the range of pH values at which both the
 analytes and sorbents are in their charged form. When these
 conditions are not met, the ionic interactions are weakened
 and retention is driven only by hydrophobic interactions.

Table 3 Recoveries (%) obtained in ultrapure water when the SCX/SAX and the SCX/WAX configurations were loaded at pH 2 and 12

	Analyte	pH 2				pH 12							
		SCX/SAX		SCX/WAX		SCX/SAX		SCX/WAX					
		w	e _a	e _b	w	e _b	w	e _a	e _b	L	w	e _b	
Basic	ATE			79	85	63	12			54	30		
	RAN			84	86	77	2			78	6		
	TRI			82	84	82				83			
	MET			82	83	59	15			53	30		
	PROP	1		89	90	48	1	36		44	46		
Neutral	CAFF	56			67			69					84
	ANTI	15	19	37	3	78	79						87
	CBZ	86			85			87					89
Acidic	SAL AC	9	28	12		93		16	9	69	1		
	CLO AC	44	39		12	71		86		78	11		
	FEN	81	6		75	16		85			91		
	DICLO	55	25		22	58		83			92		
	IBP	84			84			87			81		

%RSD lower than 15% in all cases

L loading, w w1 + w2 (1 mL MeOH + 1 mL MeOH), e_a 5 mL 5% HCOOH in MeOH, e_b 5 mL 5% NH₄OH in MeOH

459	<i>Optimization of other SPE conditions</i>	
460	Using the same cartridges selected previously, SCX/SAX and	
461	SCX/WAX, other parameters of the SPE procedure were eval-	
462	uated, such as volume in the different steps and washing sol-	
463	vent. To optimize the washing volume, 50 mL of ultrapure	
464	water adjusted to pH 5 was loaded in both cartridges, which	
465	were then washed with subsequent fractions of 1 mL of	
466	MeOH up to 15 mL. It was observed that the neutral pharma-	
467	ceuticals were lost completely in the first 1 mL washing frac-	
468	tion in both types of cartridges, while no losses were recorded	
469	for the rest of the analytes in any of the fractions. Only IBP	
470	was partially lost (40%) starting from the seventh fraction.	
471	After ascertaining that this loss was due to the selection of a	
472	pH value too close to the pK _a of IBP (4.85), and confirming	
473	that, at pH 7, the compound was not lost, the loading pH was	
474	set at this value from this point onwards. Thus, the washing	
475	volume was set at 15 mL of MeOH because no losses were	
476	observed for any of the target analytes and a volume higher	
477	than this was considered excessive. Actually, this volume of	
478	15 mL is already high for the amount of sorbent used (100 mg)	
479	and, as such, it was expected to eliminate a large number of	
480	interferences in complex matrices.	
481	In subsequent tests, 50 mL of ultrapure water adjusted to	
482	pH 7 was loaded in both cartridges, which were further	
483	washed with ACN instead of MeOH, to evaluate the influence	
484	of a different organic modifier. No differences were observed	
485	between the two organic solvents in any of the cartridges. As	
486	MeOH is most commonly used for the washing steps in	
487	mixed-mode SPE procedures, it was used for further extrac-	
488	tions. The elution volume was optimized by passing 1 mL	
489	fractions up to 5 mL of the elution solvents as follows, for	
490	SCX/SAX: (1) 10% HCOOH in MeOH solution followed by	
491	(2) 5% NH ₄ OH in MeOH solution; as for SCX/WAX: only	
492	5% NH ₄ OH in MeOH. For both cartridges, the volume ini-	
493	tially selected of 5 mL of each eluting solution proved to be	
494	enough to completely elute all of the analytes, so it was con-	
495	sidered optimal.	
496	For both configurations, SCX/SAX and SCX/WAX, the	
497	SPE protocol established to this point was transferred to	
498	500 mg cartridges, which allowed the loading of up to	
499	500 mL of ultrapure water (adjusted to pH 7) to then be	
500	washed with 15 mL of MeOH. As a result, up to 500 mL could	
501	be passed through the cartridges without observing losses of	
502	any of the analytes during the loading or washing step,	
503	obtaining recoveries that ranged from 84 to 97%. It can be	
504	seen that the retention of the combined sorbents for ionic	
505	pharmaceuticals is good and it is comparable or better to the	
506	results obtained in individual separated cartridges. For exam-	
507	ple, in studies where cation-exchangers were used to extract	
508	the same basic pharmaceuticals, recoveries were between 63	
509	and 114% even when only aqueous washing steps were used	
510	[23, 38]. When acidic pharmaceuticals were extracted in an in-	
	house anion-exchanger, recoveries ranged between 91 and	511
	98% when introducing a washing step with 10 mL of	512
	MeOH [16]. Compared to cartridges combined in tandem	513
	[24], the present results were also similar or better and the	514
	protocol was significantly simpler.	515
	Environmental water samples	516
	The optimized SPE procedures for each type of cartridge,	517
	SCX/SAX and SCX/WAX, were evaluated in river water	518
	and wastewaters to observe the performance of the combina-	519
	tions of the sorbents when dealing with complex matrices. The	520
	neutral analytes (CAFF, ANTI, and CBZ) will not be	521
	discussed below, as it was proven that they are lost during	522
	the washing step fraction. A considerable ME is commonly	523
	observed when determining pharmaceuticals in environmental	524
	waters, which increases when using high volumes of sample	525
	[39]. Thus, the first parameter evaluated when testing the op-	526
	timized method in these types of matrices was the sample	527
	volume. The optimal volumes are indicated in “Solid-phase	528
	extraction procedure” section, selected according to the break-	529
	through volume of the analytes and the ME observed.	530
	At this point, both the SCX/SAX and SCX/WAX car-	531
	tridges were compared with regard to their performance in	532
	environmental waters to select a single configuration for sub-	533
	sequent tests. The optimal procedure described in “Solid-	534
	phase extraction procedure” section was applied to analyze	535
	100 mL of effluent wastewater sample (spiked at 2.5 μg/L)	536
	using the two types of cartridges. The results are shown in	537
	Table 4. The %R _{SPE} was defined as the recovery obtained only	538
	in the SPE procedure and it was calculated as the ratio be-	539
	tween the concentrations obtained from a sample spiked be-	540
	fore and after the SPE procedure. The ME was calculated from	541
	the concentration obtained when the extract of the sample was	542
	spiked just before injection into the LC-HRMS. This concen-	543
	tration (C _{exp}) was introduced in the formula %ME = [(C _{exp} /	544
	C _{theo}) × 100] – 100, where C _{theo} is the theoretical concentra-	545
	tion in the final volume of sample injected in the LC-HRMS	546
	instrument. The %R _{apparent} was defined as the recovery of the	547
	whole method and it was calculated from the concentration	548
	obtained from a sample spiked at the beginning of the com-	549
	plete analysis. All of the experimental concentrations men-	550
	tioned were calculated using a calibration curve prepared in	551
	pure standard.	552
	It can be observed in Table 4 that, for the SCX/SAX con-	553
	figuration, values of %R _{SPE} were between 83 and 104%, with	554
	the exception of RAN (60%) and SAL AC (62%), similar to	555
	the SCX/WAX configuration, which showed values between	556
	79 and 109%, except for RAN (69%) and IBP (68%).	557
	%R _{apparent} values ranged from 45 to 88% and from 47 to	558
	88% for the SCX/SAX and SCX/WAX cartridges, respective-	559
	ly. The ME was below – 28% in both cartridges, except for	560
	ATE and TRI, which showed values around – 48%. Clearly,	561

Table 4 Comparison of the performance between the SCX/SAX and SCX/WAX configurations for extracting the selected basic and acidic pharmaceuticals from 100 mL of effluent wastewater

	Analyte	SCX/SAX			SCX/WAX		
		%R _{apparent}	%R _{SPE}	%ME	%R _{apparent}	%R _{SPE}	%ME
Basic	ATE	53	104	- 49	55	105	- 48
	RAN	51	60	- 15	52	69	- 25
	TRI	50	97	- 48	54	97	- 44
	MET	70	84	- 17	73	91	- 20
	PROP	68	83	- 18	58	79	- 27
Acidic	SAL AC	45	62	- 28	79	109	- 27
	CLO AC	88	97	- 10	88	105	- 17
	FEN	72	95	- 24	74	102	- 28
	DICLO	74	101	- 27	76	104	- 27
	IBP	68	96	- 28	47	68	- 31

%RSD lower than 10% in all cases

no differences were observed between both types of configurations, suggesting that either of them can be equally used for simultaneously extracting the charged pharmaceuticals from the samples. For further tests, the SCX/SAX was selected from the two types of sorbents because it gave a higher %R_{apparent} for IBP, which is one of the analytes with a lower response in the LC-HRMS.

Using the SCX/SAX configuration, the ME was compared when extracting effluent and influent wastewaters (spiked at 5 and 10 µg/L, respectively), and the washing step was applied or omitted, to determine the efficiency of including this cleaning step in the SPE procedure. Figure 2 shows the results of this evaluation, demonstrating a decrease in the ME obtained for several of the analytes, especially basic pharmaceuticals, when the washing step was included. The ME obtained when applying the washing step ranged between - 49 and - 15% for effluent wastewater, and between - 51 and -20% in the case of influent wastewater, which was higher than expected considering that the volume used for the washing step (15 mL MeOH) was quite high. These results can be attributed to the presence of a high content of ionic interferences in the samples that contribute to the ME to a high degree.

The SPE procedure using the SCX/SAX configuration was further evaluated in river and wastewater samples in terms of %R_{SPE} and %R_{apparent}. The results obtained for river waters and influent wastewaters spiked at 2.5 and 10 µg/L, respectively, are summarized in Table 5. Satisfactory %R_{SPE} values were obtained for all the matrices, being higher than 90% in 70% of the cases. Values of %R_{apparent} for river water samples were between 45 and 109%, with the exception of IBP (31%), while, for effluent and influent wastewater samples, values ranged from 52 to 83% (except for SAL AC, which were 30%) and from 34 to 76%, respectively. The ME obtained with the present method can be attributed only to ionic interferences. The method demonstrated the capability of

simultaneously retaining acidic and basic analytes and the advantage of eliminating all neutral interferences, features that could be transferred to other groups of ionizable compounds and highly complex matrices.

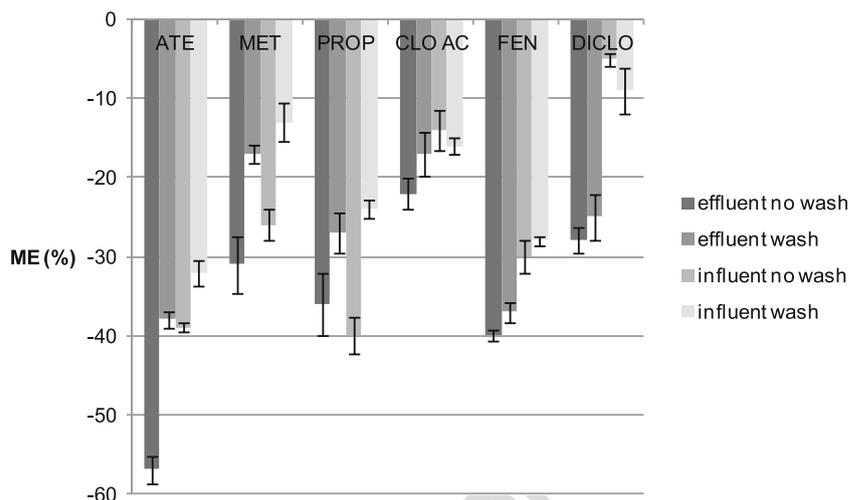
Method validation and application to environmental samples

The optimized method using the SCX/SAX cartridge was validated in river water and effluent and influent wastewater samples to check its repeatability and detection and quantification limits. Repeatability was expressed as the relative standard deviation (%RSD, n = 5) and it was evaluated within the same day (results shown in Table 5) and on consecutive days. The method exhibited satisfactory precision, as %RSD values ranged between 0.4 and 24% for all of the pharmaceuticals in all matrices.

Because several of the target analytes were present in the non-spiked wastewater samples, the use of matrix-matched calibration curves to correct the ME was not possible for these matrices. In these cases, external calibration curves were used for quantification, taking %R_{apparent} values into consideration. In the case of river water, a matrix-matched calibration curve was prepared by spiking at different concentrations 100 mL volumes of river water, which were extracted in the SCX/SAX cartridges and subsequently injected into the LC-HRMS instrument. Linearity was good for all of the compounds (R² ≥ 0.9988) between the MQLs (reported in Table 5) and 500 ng/L, except for RAN, SAL AC, and IBP, which showed poor linearity in the concentration range tested. In these cases, external calibration curves were also used.

For the wastewater samples, method detection (MDLs) and quantification (MQLs) limits were estimated from the instrumental limits (ILODs and ILOQs), taking into account the %R_{apparent} values. For river water samples, MDLs were the spiked concentrations that showed a signal for the more

Fig. 2 ME (%) obtained for 100 mL of effluent (spiked at 5 µg/L) and 50 mL of influent (10 µg/L) wastewater samples when applying or not the washing step (15 mL of MeOH) of the optimized SPE procedure



632 abundant fragment around 1×10^3 , while MQs were the first
 633 points of the matrix-matched calibration curves. MDLs were
 634 between 0.5 and 5 ng/L for river water samples, while for
 635 effluent and influent wastewater samples, values were be-
 636 tween 1 and 75 ng/L and between 3 and 260 ng/L, respecti-
 637 vely. Table 5 shows the MQs values, which ranged from 1 and
 638 25 ng/L for river water samples, while, for effluent and influ-
 639 ent wastewater samples, values ranged from 5 to 80 ng/L and
 640 from 15 to 265 ng/L, respectively.

641 Three different samples of effluent and influent wastewater
 642 and two different samples of river water were analyzed using
 643 the validated method for the SCX/SAX combination. The
 644 ranges of concentrations found are shown in Table S2 in the
 645 ESM. Moreover, Figs. S1–S3 in the ESM show the extract ion
 646 chromatograms of analyzed samples for each type of matrix
 647 studied. In river samples, several compounds were detected,
 648 but only ATE, MET, SAL AC, and CLO AC were quantified
 649 at concentrations between 1 and 50 ng/L. Most of the com-
 650 pounds were quantified in effluent and influent wastewater

651 samples with levels ranging from 20 to 2500 ng/L in the case
 652 of effluent wastewater samples and from 40 to 50,000 ng/L in
 653 the case of influent wastewater. The pharmaceuticals found at
 654 the highest concentrations were ATE, SAL AC, DICLO, and
 655 IBP and levels found for all of the analytes were in line with
 656 those reported in the literature [4, 31, 40].

Conclusions

657
 658 The four combinations tested SCX/SAX, SCX/WAX, WCX/
 659 SAX, and WCX/WAX, simultaneously and strongly retained
 660 basic and acidic pharmaceuticals, as long as the charges are
 661 balanced and the SPE conditions are carefully selected. No
 662 substantial differences were observed between the four com-
 663 binations evaluated in ultrapure water apart from the optimum
 664 loading pH ranges, suggesting that they all could be potential-
 665 ly useful depending on the application. The correct selection

Table 5 %R_{apparent}, %R_{SPE}, ME, %RSD (n = 5), and MQs values obtained when 100 mL of river water or 50 mL of influent wastewater sample was percolated through SCX/SAX cartridge

Analyte	%R _{apparent}	River				%R _{apparent}	Influent			
		%RSD	%R _{SPE}	%ME	MQ (ng/L)		%RSD	%R _{SPE}	%ME	MQ (ng/L)
ATE	109	3	104	- 5	1	61	1	102	- 42	15
RAN	68	11	83	- 25	20	71	1	79	10	110
TRI	98	5	94	- 4	2	67	1	98	- 35	120
MET	102	2	93	5	2	70	8	93	- 24	115
PROP	90	16	90	- 3	2	70	10	97	- 33	115
SAL AC	50	4	92	- 34	20	34	14	57	- 40	120
CLO AC	65	16	90	- 21	2	76	22	101	- 19	35
FEN	45	6	76	- 36	5	49	8	99	- 41	15
DICLO	48	18	74	- 32	2	62	4	107	- 24	15
IBP	31	14	60	- 41	25	38	2	116	- 61	265

666 of the pH value used to load the samples and elute the analytes
 667 was very important in terms of favoring the ionic interactions
 668 between the analytes and the sorbents.

669 The strong retention of the analytes on the sorbents allowed
 670 the introduction of a washing step with a high volume of
 671 MeOH (15 mL), which proved to eliminate neutral interfer-
 672 ences present in the matrices. As indicated by the results ob-
 673 tained when extracting environmental waters using the present
 674 method, ionic interferences contribute to the matrix effect to a
 675 high degree. The performance of the method was comparable
 676 to other studies reported but the protocol was considerably
 677 simpler thanks to the combination of the cartridges. The meth-
 678 od was validated in river and wastewater samples and several
 679 of the selected analytes were successfully quantified in the
 680 samples at levels that were similar to those reported in other
 681 studies.

682 The potential of combining sorbents to obtain positive and
 683 negative charges in the same SPE cartridge was confirmed and
 684 optimal extraction conditions were given to obtain the best
 685 performance. Promising results might be expected for other
 686 basic or acidic compounds and other samples with complex
 687 matrices.

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692 **Compliance with ethical standards**

694 **Conflict of interest** The authors declare that they have no conflicts of
 695 interest.

696 **References**

698 1. Plotka-Wasyłka J, Szczepanska N, de la Guardia M, Namiesnik J.
 699 Modern trends in solid phase extraction: new sorbent media. *Trends*
 700 *Anal Chem.* 2016;77:23–43.
 701 2. Leendert V, Van Langenhove H, Demeestere K. Trends in liquid
 702 chromatography coupled to high-resolution mass spectrometry for
 703 multi-residue analysis of organic micropollutants in aquatic envi-
 704 ronments. *Trends Anal Chem.* 2015;67:192–208.
 705 3. Petrovic M, Farré M, López de Alda M, Pérez S, Postigo C, Koeck
 706 M, et al. Recent trends in the liquid chromatography-mass spec-
 707 trometry analysis of organic contaminants in environmental sam-
 708 ples. *J Chromatogr A.* 2010;1217(25):4004–17.
 709 4. Richardson SD, Kimura SY. Water analysis: emerging contami-
 710 nants and current issues. *Anal Chem.* 2016;88(1):546–82.
 711 5. Hernández F, Ibáñez M, Bade R, Bijlsma L, Sancho JV.
 712 Investigation of pharmaceuticals and illicit drugs in waters by liq-
 713 uid chromatography-high-resolution mass spectrometry. *Trends*
 714 *Anal Chem.* 2014;63:140–57.
 715 6. Fontanals N, Cormack PAG, Marcé RM, Borrull F. Mixed-mode
 716 ion-exchange polymeric sorbents: dual-phase materials that im-
 717 prove selectivity and capacity. *Trends Anal Chem.* 2010;29(7):
 718 765–79.

719 7. González-Marino I, Quintana JB, Rodríguez I, Rodil R, González-
 720 Penas J, Cela R. Comparison of molecularly imprinted, mixed-
 721 mode and hydrophilic balance sorbents performance in the solid-
 722 phase extraction of amphetamine drugs from wastewater samples
 723 for liquid chromatography-tandem mass spectrometry determina-
 724 tion. *J Chromatogr A.* 2009;1216(48):8435–41.
 725 8. Ngoc Han T, Hu J, Ong SL. Simultaneous determination of PPCPs,
 726 EDCs, and artificial sweeteners in environmental water samples
 727 using a single-step SPE coupled with HPLC-MS/MS and isotope
 728 dilution. *Talanta.* 2013;113:82–92.
 729 9. Al-Odaini NA, Zakaria MP, Yaziz MI, Surif S. Multi-residue ana-
 730 lytical method for human pharmaceuticals and synthetic hormones
 731 in river water and sewage effluents by solid-phase extraction and
 732 liquid chromatography-tandem mass spectrometry. *J Chromatogr*
 733 *A.* 2010;1217(44):6791–806.
 734 10. Sousa MA, Goncalves C, Cunha E, Hajslova J, Alpendurada MF.
 735 Cleanup strategies and advantages in the determination of several
 736 therapeutic classes of pharmaceuticals in wastewater samples by
 737 SPE-LC-MS/MS. *Anal Bioanal Chem.* 2011;399(2):807–22.
 738 11. Li Y, Yang J, Huang C, Wang L, Wang J, Chen J. Dendrimer-
 739 functionalized mesoporous silica as a reversed-phase/anion-ex-
 740 change mixed-mode sorbent for solid phase extraction of acid drugs
 741 in human urine. *J Chromatogr A.* 2015;1392:28–36.
 742 12. Pavlovic DM, Babic S, Horvat AJM, Kastelan-Macan M. Sample
 743 preparation in analysis of pharmaceuticals. *Trends Anal Chem.*
 744 2007;26(11):1062–75.
 745 13. Carpinteiro I, Ramil M, Rodríguez I, Cela R. Determination of
 746 fungicides in wine by mixed-mode solid phase extraction and liquid
 747 chromatography coupled to tandem mass spectrometry. *J*
 748 *Chromatogr A.* 2010;1217(48):7484–92.
 749 14. Ansermot N, Brawand-Amey M, Eap CB. Simultaneous quantifi-
 750 cation of selective serotonin reuptake inhibitors and metabolites in
 751 human plasma by liquid chromatography-electrospray mass spec-
 752 trometry for therapeutic drug monitoring. *J Chromatogr B.*
 753 2012;885:117–30.
 754 15. González-Marino I, Quintana JB, Rodríguez I, González-Diez M,
 755 Cela R. Screening and selective quantification of illicit drugs in
 756 wastewater by mixed-mode solid-phase extraction and
 757 quadrupole-time-of-flight liquid chromatography-mass spectrome-
 758 try. *Anal Chem.* 2012;84(3):1708–17.
 759 16. Bratkowska D, Davies A, Fontanals N, Cormack PAG, Borrull F,
 760 Sherrington DC, et al. Hypercrosslinked strong anion-exchange
 761 resin for extraction of acidic pharmaceuticals from environmental
 762 water. *J Sep Sci.* 2012;35(19):2621–8.
 763 17. Triñanes S, Casais MC, Mejuto MC, Cela R. Selective determina-
 764 tion of COXIBs in environmental water samples by mixed-mode
 765 solid phase extraction and liquid chromatography quadrupole time-
 766 of-flight mass spectrometry. *J Chromatogr A.* 2015;1420:35–45.
 767 18. Salas D, Borrull F, Marcé RM, Fontanals N. Study of the retention
 768 of benzotriazoles, benzothiazoles and benzenesulfonamides in
 769 mixed-mode solid-phase extraction in environmental samples. *J*
 770 *Chromatogr A.* 2016;1444:21–31.
 771 19. Buchberger WW. Current approaches to trace analysis of pharma-
 772 ceuticals and personal care products in the environment. *J*
 773 *Chromatogr A.* 2011;1218(4):603–18.
 774 20. Bratkowska D, Marcé RM, Cormack PAG, Sherrington DC,
 775 Borrull F, Fontanals N. Synthesis and application of
 776 hypercrosslinked polymers with weak cation-exchange character
 777 for the selective extraction of basic pharmaceuticals from complex
 778 environmental water samples. *J Chromatogr A.* 2010;1217(10):
 779 1575–82.
 780 21. Wang J, Gardinali PR. Analysis of selected pharmaceuticals in fish
 781 and the fresh water bodies directly affected by reclaimed water
 782 using liquid chromatography-tandem mass spectrometry. *Anal*
 783 *Bioanal Chem.* 2012;404(9):2711–20.

- 784 22. David A, Abdul-Sada A, Lange A, Tyler CR, Hill EM. A new
785 approach for plasma (xeno)metabolomics based on solid-phase ex-
786 traction and nanoflow liquid chromatography-nanoelectrospray
787 ionisation mass spectrometry. *J Chromatogr A*. 2014;1365:72–85.
788 23. Kasprzyk-Hordern B, Dinsdale RM, Guwy AJ. Multi-residue meth-
789 od for the determination of basic/neutral pharmaceuticals and illicit
790 drugs in surface water by solid-phase extraction and ultra perfor-
791 mance liquid chromatography-positive electrospray ionisation tan-
792 dem mass spectrometry. *J Chromatogr A*. 2007;1161(1–2):132–45.
793 24. Lavén M, Alsberg T, Yu Y, Adolfsson-Erici M, Sun H. Serial
794 mixed-mode cation- and anion-exchange solid-phase extraction
795 for separation of basic, neutral and acidic pharmaceuticals in waste-
796 water and analysis by high-performance liquid chromatography-
797 quadrupole time-of-flight mass spectrometry. *J Chromatogr A*.
798 2009;1216(1):49–62.
799 25. Deeb AA, Schmidt TC. Tandem anion and cation exchange solid
800 phase extraction for the enrichment of micropollutants and their
801 transformation products from ozonation in a wastewater treatment
802 plant. *Anal Bioanal Chem*. 2016;408(16):4219–32.
803 26. Bortolomeazzi R, Munari M, Anese M, Verardo G. Rapid mixed
804 mode solid phase extraction method for the determination of acryl-
805 amide in roasted coffee by HPLC-MS/MS. *Food Chem*.
806 2012;135(4):2687–93.
807 27. Grujic S, Vasiljevic T, Lausevic M. Determination of multiple phar-
808 maceutical classes in surface and ground waters by liquid
809 chromatography-ion trap-tandem mass spectrometry. *J*
810 *Chromatogr A*. 2009;1216(25):4989–5000.
811 28. Kleywegt S, Pileggi V, Yang P, Hao C, Zhao X, Rocks C, et al.
812 Pharmaceuticals, hormones and bisphenol A in untreated source
813 and finished drinking water in Ontario, Canada—and treatment
814 efficiency. *Sci Total Environ*. 2011;409(8):1481–8.
815 29. Dasenaki ME, Thomaidis NS. Multianalyte method for the deter-
816 mination of pharmaceuticals in wastewater samples using solid-
817 phase extraction and liquid chromatography-tandem mass spec-
818 trometry. *Anal Bioanal Chem*. 2015;407(15):4229–45.
819 30. Tarcomnicu I, van Nuijs ALN, Simons W, Bervoets L, Blust R,
820 Jorens PG, et al. Simultaneous determination of 15 top-prescribed
821 pharmaceuticals and their metabolites in influent wastewater by
822 reversed-phase liquid chromatography coupled to tandem mass
823 spectrometry. *Talanta*. 2011;83(3):795–803.
824 31. Gómez MJ, Gómez-Ramos MM, Malato O, Mezcuá M, Fernández-
825 Alba AR. Rapid automated screening, identification and quantifi-
826 cation of organic micro-contaminants and their main transformation
827 products in wastewater and river waters using liquid
828 chromatography-quadrupole-time-of-flight mass spectrometry with
874 an accurate-mass database. *J Chromatogr A*. 2010;1217(45):7038–
54. 830
32. Medana C, Calza P, Carbone F, Pelizzetti E, Hidaka H, Baiocchi C. 831
Characterization of atenolol transformation products on light- 832
activated TiO₂ surface by high-performance liquid 833
chromatography/high-resolution mass spectrometry. *Rapid* 834
Commun Mass Spectrom. 2008;22(3):301–13. 835
33. Martínez Bueno MJ, Aguera A, Gómez MJ, Dolores Hernando M, 836
García-Reyes JF, Fernández-Alba AR. Application of liquid 837
chromatography/quadrupole-linear ion trap mass spectrometry and 838
time-of-flight mass spectrometry to the determination of pharma- 839
ceuticals and related contaminants in wastewater. *Anal Chem*. 840
2007;79(24):9372–84. 841
34. Jakimska A, Sliwka-Kaszynska M, Reszczynska J, Namiesnik J, 842
Kot-Wasik A. Elucidation of transformation pathway of ketoprofen, 843
ibuprofen, and furosemide in surface water and their occurrence in 844
the aqueous environment using UHPLC-QTOF-MS. *Anal Bioanal* 845
Chem. 2014;406(15):3667–80. 846
35. How ZT, Busetti F, Linge KL, Kristiana I, Joll CA, Charrois JWA. 847
Analysis of free amino acids in natural waters by liquid 848
chromatography-tandem mass spectrometry. *J Chromatogr A*. 849
2014;1370:135–46. 850
36. González-Mariño I, Quintana JB, Rodríguez I, González-Diez M, 851
Cela R. Screening and selective quantification of illicit drugs in 852
wastewater by mixed-mode solid-phase extraction and 853
quadrupole-time-of-flight liquid chromatography-mass spectrome- 854
try. *Anal Chem*. 2012;84(3):1708–17. 855
37. Fontanals N, Miralles N, Abdullah N, Davies A, Gilart N, Cormack 856
PAG. Evaluation of strong cation-exchange polymers for the deter- 857
mination of drugs by solid-phase extraction-liquid 858
chromatography-tandem mass spectrometry. *J Chromatogr A*. 859
2014;1343:55–62. 860
38. Batt AL, Kostich MS, Lazorchak JM. Analysis of ecologically rel- 861
evant pharmaceuticals in wastewater and surface water using selec- 862
tive solid-phase extraction and UPLC-MS/MS. *Anal Chem*. 863
2008;80(13):5021–30. 864
39. Gilart N, Cormack PAG, Marcé RM, Fontanals N, Borrull F. 865
Selective determination of pharmaceuticals and illicit drugs in 866
wastewaters using a novel strong cation-exchange solid-phase ex- 867
traction combined with liquid chromatography-tandem mass spec- 868
trometry. *J Chromatogr A*. 2014;1325:137–46. 869
40. López-Serna R, Petrovic M, Barceló D. Occurrence and distribution 870
of multi-class pharmaceuticals and their active metabolites and 871
transformation products in the Ebro River basin (NE Spain). *Sci* 872
Total Environ. 2012;440:280–9. 873