

1 **Direct liquid-liquid extraction of lipid from municipal sewage sludge for biodiesel**  
2 **production**

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15  
16 **Abstract**

17 Municipal sludge from wastewater treatment plants is a promising lipid feedstock for  
18 biodiesel production as it contains a significant amount of lipids. However, the energy  
19 necessary to remove its high water content is a major inconvenience for scaling up  
20 because of the high associated cost. In addition, the expensive conventional sludge  
21 drying methods are not effective enough for for lipid recovery, thus reducing the  
22 potential biodiesel production. This study explores an alternative method, the direct  
23 sequential liquid-liquid extraction, which was performed in a batch mixer-settler reactor  
24 at room temperature, using hexane as a solvent, after previous sludge acidification  
25 showed significant increase in the lipid efficiency. The optimisation study demonstrated  
26 that, after three stages, 91% of lipid from primary sludge was recovered. The optimised

27 extraction gave slightly higher lipid (27%, dry sludge) than the standard method (25%,  
28 dry sludge), supporting the suitability of the proposed process. Finally, this work  
29 demonstrates that the residual lipid-extracted sludge is still a good feedstock for energy  
30 production via anaerobic digestion. Anyway, the economic and environmental aspects  
31 of biodiesel production from sewage sludge could be improved.

32

### 33 **Key-words**

34 Sewage sludge; liquid-liquid extraction; acidification; lipid; biodiesel.

35

## 36 **1. INTRODUCTION**

37

38 The global continuous growth of energy demand poses urgent problem due to the  
39 fossil fuels depletion, as they currently represent about 75% of all energy consumed  
40 worldwide [1]. One of the most promising renewable fuels proposed as an alternative is  
41 biodiesel that can be directly used with current engine and refuelling technology [1-3].  
42 However, the competitive potential of biodiesel is currently limited by the price of the  
43 common lipid feedstocks, which constitutes 70-85% of the overall biodiesel production  
44 cost, thus strongly influencing the final price of this biofuel and raising the concerns of  
45 food shortage versus fuel crisis [1].

46 In turn, municipal sewage sludge from wastewater treatment plants (WWTPs) is  
47 gaining more attention nowadays as a lipid feedstock for the production of biodiesel as  
48 the dry sludge can contain up to 30 wt% of lipids [1-6]. In fact, sewage sludge is a  
49 waste that needs specific treatment before disposal and represents a major cost in the  
50 WWTP operation. In addition, the WWTPs annually produce higher amounts of sludge  
51 due to the expansion of urbanised and industrialised areas. Therefore, the sewage  
52 sludge can be envisaged as a relatively cheap, readily available, and in abundance

53 feedstock, which can make the biodiesel production profitable. Furthermore, it is one  
54 possible alternative to take advantage of the excess sludge, reusing it as a source of  
55 lipid for the production of biodiesel, consequently lowering the WWTP operation cost.  
56 Nevertheless, the production of biodiesel from sludge poses great challenges for a fast  
57 commercialisation. The main challenge to be faced by biodiesel production from waste  
58 sludge is an efficient lipid extraction from water, as water can account for up to 95-98  
59 wt %, so dewatering and drying constitute more than 50% of total biodiesel production  
60 cost [4]. This makes the production very expensive and difficult the scale-up due to the  
61 cost of the energy necessary for water removal step.

62 Most of the literature reports only the utilisation of dry sludge in the extraction of  
63 lipid by an organic solvent [3-4, 6]. Recently, some works have used dewatered primary  
64 [5] and secondary sludge [7] by centrifugation, but the energy of dewatering still  
65 constitutes 14% of total biodiesel production cost [4]. On the other hand, the direct  
66 transesterification of sewage sludge into biodiesel has been also reported “in situ” on  
67 dry [2, 4] and dewatered sludge [7]. Interestingly, the biodiesel yield obtained from  
68 dewatered sludge was about 20% lower than from dried sludge [7]. The “in situ”  
69 process can reduce the time and amount of solvent, however, after transesterification, a  
70 solvent recovery step is then needed, adversely affecting the overall cost of biodiesel.

71 Moreover, water elimination from biomass by conventional thermal drying or freeze-  
72 drying results in the loss of valuable organic compounds [8-9]. This fact can also  
73 provoke the loss of lipids in sewage sludge hence decreasing biodiesel production yield.  
74 Nevertheless, the influence of sludge drying on the lipid extraction efficiency has not  
75 been yet evaluated. Therefore, the effect of common sludge drying methods on the lipid  
76 extraction efficiency as well as the fatty acids composition still needs to be examined.

77 Surprisingly, the direct liquid-liquid extraction has neither been reported, so the  
78 sludge drying and dewatering would thus become unnecessary. Thus, the main

79 objective of this study was to explore this alternative and to demonstrate its feasibility.  
80 Three types of sludge generated in WWTPs were tested. Optimisation of liquid-liquid  
81 extraction was studied varying the ratio sludge/hexane, time of contact, and number of  
82 consecutive batch extraction steps in order to get the most favourable process. In  
83 addition, as the residual sludge after lipid extraction is still a potential biomass for  
84 energy recovery, the residual sludge can be used as feed for anaerobic digestion, which  
85 is widely implemented in municipal WWTPs. Therefore, the lipid-extracted sludge was  
86 subjected to anaerobic digestion to check out its potential for biogas generation. Finally,  
87 a simplified energy consumption estimation of the biodiesel production via liquid-  
88 liquid extraction was conducted.

89

## 90 **2. MATERIALS AND METHODS**

91

### 92 **2.1. Reagents**

93

94 The transesterification/esterification experiments were carried out using anhydrous  
95 methanol and sulfuric acid from Sigma-Aldrich at the highest purity available. Standard  
96 used for identification and quantification of fatty acid methyl esters (FAMES) was  
97 supplied by Supelco (37 component FAMES mix, ref: 47885-U). For the free fatty acids  
98 FFAs analysis, 0.5 M potassium hydroxide volumetric solution was purchased from  
99 Fluka. All other solvents and reagents were high performance liquid chromatography  
100 grade and analytical reagent grade provided by Sigma-Aldrich.

101

### 102 **2.2. Sludge collection and handling**

103

104 Primary, secondary and blended sludge were collected from the municipal WWTP in  
105 Reus (Tarragona, Spain) with a capacity to daily process 25000 m<sup>3</sup> of wastewater. Fig.  
106 1a shows a schematic diagram of the WWTP, illustrating the step where these different  
107 types of sludge are generated. Primary sludge was collected after partial gravity  
108 thickening. Secondary sludge, produced by an activated sludge process, was collected  
109 after partial thickening by flotation. Blended sludge was collected after the combination  
110 of primary and secondary at a ratio of 65:35, v/v. The collected sludges were  
111 immediately stored at 4°C prior to use. Because the sludge properties could be changed  
112 during long storage time, fresh sludge was always used for each experiment.

113 The inoculum used in anaerobic digestion tests was sludge collected from a  
114 mesophilic anaerobic digester in the same facility.

115

## 116 **2.3. Sludge drying**

117

### 118 **2.3.1 Primary sludge – evaluation of drying methods**

119

120 According to standard method 5520E [10], sludge was dried using magnesium  
121 sulfate monohydrate but without previous acidification. Using the referenced method,  
122 the sludge sample was considered as completely dried.

123 Oven drying method was conducted using an universal oven ULE400 (Memmert  
124 GmbH, Germany) at two different temperatures, 105°C for 2 days based on standard  
125 method 2540G [10] or 70°C for 3 days [3].

126 Freeze-drying method was conducted by using the method presented elsewhere [2].  
127 At first, sludge was centrifuged and then allowed to freeze for 2 days at -20°C.  
128 Afterwards, the frozen sludge was freeze-dried in an automatic vacuum freeze dryer,  
129 model FT33-A (Armfield Limited) for 2 days.

130 In the sun drying method, the sludge sample was left outside for 10 days, where the  
131 temperature was in the range of 25-35°C.

132 Drying under fume hood was performed based on the method presented elsewhere  
133 [6]. The sludge was first centrifuged and then put in a fume hood for 4 days at ambient  
134 temperature.

135 Approximately 500 mL of sludge was used for the drying procedures, except for the  
136 standard method. After all drying methods, sludge was crushed to fine particles. To  
137 determine the final moisture content, 1 g of crushed sludge was placed in the oven at  
138 105°C and dried until reaching constant weight.

139

### 140 **2.3.2 Drying of primary, secondary and blended sludge by standard method**

141

142 The sludges were dried using magnesium sulfate monohydrate according to standard  
143 method 5520E [10] with previous acidification. The reference method was used for a  
144 comparison study to a novel liquid-liquid extraction from acidified liquid sludge.

145

## 146 **2.4. Lipid extraction**

147

### 148 **2.4.1 Extraction of lipids from dried sludge**

149

150 The extraction after drying was carried out in a Soxhlet apparatus using hexane as a  
151 solvent according to standard method 5520E [10]. After extraction, the hexane was  
152 removed using a rotary evaporator at 40°C under vacuum at 50 mbar. Then, the remnant  
153 lipid fraction was stored in a desiccator overnight and weighed the next day to  
154 determine the extraction yield.

155

156 **2.4.2 Liquid-liquid extraction of lipids from primary, secondary and blended**  
157 **sludges**

158

159 Sequential liquid-liquid extraction of lipids was performed in a batch mixer-settler  
160 reactor with mechanical agitation (330 rpm), at ambient temperature, using hexane as  
161 solvent and 200 mL of sludge. The effect of previous sludge acidification to pH=2 was  
162 evaluated. This pH was attained by addition of approximately 3 mL of concentrated  
163 HCl to the sample of 200 mL of sludge. The experimental setup for the liquid-liquid  
164 extraction is presented in Fig. 1b. Until nine consecutive extraction stages were  
165 conducted, in which the sludge, after settling, was extracted again with additional fresh  
166 solvent. The mechanical settling was performed at 60 rpm for 15 min for primary and  
167 blended sludge and for 30 min for secondary sludge. After each extraction stage, the  
168 hexane phase was filtered using a 2-4  $\mu\text{m}$  filter paper in order to eliminate the  
169 remaining solid particles and then dried over anhydrous sodium sulfate. Later, hexane  
170 was removed using a rotary evaporator at 40°C under vacuum at 50 mbar and reused for  
171 the consecutive stage. Lipids were stored in a desiccator overnight and weighed the  
172 next day to determine the extraction yield.

173

174 **2.5. Anaerobic digestion of lipid-extracted sludge**

175

176 As depicted in Fig. 1, primary sludge after lipid liquid-liquid extraction was  
177 subjected to anaerobic digestion directly and after residual solvent evaporation. The  
178 residual hexane was removed using a rotary evaporator at 40°C under vacuum at 50  
179 mbar. The sludge was anaerobically digested at 33°C under mesophilic conditions [11].  
180 Lipid-extracted sludge (LES) and evaporated lipid-extracted sludge (ELES) were  
181 digested in order to evaluate the impact of the remaining solvent on biogas production.

182 Anaerobic digestion test was conducted in 120 mL serum bottles in triplicate. Digested  
183 sludge was used as inoculum and, although acclimation is not strictly required, an  
184 anaerobic semi-continuous plant was set to adapt inoculum to a more stable  
185 temperature, 33°C. The optimal digestion conditions were assured with anaerobic basic  
186 medium addition [11].

187 Then, ELES and LES were used as substrates. Substrate to inoculum ratio was fixed  
188 to 0.5:1 in a VS base. Deionised water was added to reach a final volume of 80 mL and  
189 the reactors were closed with a septum and an aluminium crimp. Finally, the reactors  
190 were purged with nitrogen to assure anaerobic conditions and placed into an oven at 33  
191 °C. Blank assays were prepared without substrate addition, and its biogas production  
192 was subtracted from the reactors fed with the substrates. Biogas production was  
193 volumetrically measured by liquid displacement. The experiment was considered  
194 completed after 25 days, when biogas production was negligible.

195 Biogas composition was analysed using an Agilent gas chromatograph (6890GC)  
196 equipped with a thermal conductivity detector and a Porapak Q 50/80 packed column  
197 (CP99960C). Both methane and carbon dioxide were quantified, and the results  
198 expressed as methane percentage in a two component mixture.

199 Volatile fatty acids (VFA) were analysed in the soluble phase by gas  
200 chromatography using a flame-ionization detector (GC-FID). The method was  
201 performed according to Agilent Application Note 228-398 using a HP-INNOWax  
202 column (19091N-133).

203

## 204 **2.6. Lipid and biodiesel analysis**

205

206 The content of free fatty acids (FFAs) was analysed according to section 9.1 of  
207 European standard method EN ISO 660 (2009). Due to the predominance of palmitic



208 acid in the sludge lipids, the results of FFA content were expressed as equivalent to  
209 palmitic acid.

210 The lipids were converted into FAMES (biodiesel) through acid catalysed  
211 transesterification/esterification using a modified version of Christi's method [4], i.e.,  
212 with hexane instead of toluene. This method was chosen because of the high amount of  
213 FFAs in the sludge lipid fraction. The FAMES were analysed by GC-FID according to  
214 Agilent Application Note 228-398 using a HP-INNOWax column (19091N-133). For  
215 the calibration of the method, a 37 component FAMES standard mixture was used  
216 (Supelco: 47885-U). The samples were also subjected to GS-MS analysis  
217 (G1099A/MSD5973) using a HP-FFA column (19091F-433). The results of the GC-  
218 FID were used to estimate the amount of saponifiable (esterifiable) material in the lipid  
219 fraction and hence the maximum mass of biodiesel (FAMES) that could yield. The  
220 compounds which could not be identified by GC-FID are presented as others. The other  
221 compounds identified by GC-MS are described in section 3.4.

222

### 223 **3. RESULTS AND DISCUSSION**

224

#### 225 **3.1. Sludge characterisation**

226

227 Each sample of received sludge was analysed in triplicate in order to determine the  
228 total solids (TS) and volatile solids (VS) contents according to standard method 2540G,  
229 and lipid content according to standard method 5520E [10]. The results in Table 1 show  
230 that TS and VS contents were very similar for all types of sludge tested. On the other  
231 hand, the lipid contents showed clear differences between the sludges. Thus, primary  
232 sludges achieved the greatest lipid fraction, followed by blended and secondary.  
233 Primary sludge mainly consists of organic matter from non-treated raw wastewater, so

234 it is a combination of floating grease and solids; instead, the secondary sludge is mainly  
235 composed of microbial cells and suspended solids produced during the aerobic  
236 biological treatment of the primary treated wastewater. Thus, it is expected that primary  
237 sludge gives the highest lipid fraction as most of this fraction is originally formed by  
238 fats whereas lipids from secondary sludge come from the cells after breaking their  
239 structure. As blended sludge is a mixture of primary and secondary, with a higher  
240 fraction of the first one, it results in the intermediate lipid content.

241 Comparing both primary sludges, some differences in the TS, VS and lipids can be  
242 observed (Table 1). As the primary sludges were collected in different days, this indeed  
243 implies variations in their composition. The fluctuations may be the result of climate  
244 changes or by deviations in the amount and quality of the wastewater received in the  
245 WWTP.

246

### 247 **3.2. Effect of sludge drying methods**

248

249 The influence of the conventional sludge drying methods on the moisture, lipid and  
250 biodiesel yields is illustrated in Table 2. Comparing with  $\text{MgSO}_4 \cdot \text{H}_2\text{O}$  drying, thereafter  
251 the standard method, the other drying methods showed a negative effect on both lipids  
252 extracted and saponifiable matter recovery, thereby decreasing the potential biodiesel  
253 yield.

254 The content of final moisture in the sludge is an important factor that explains the  
255 adverse effect on the lipid extracted as well as biodiesel produced. Water contained  
256 within the biomass has a tendency to shield lipids from the extracting solvent. As seen  
257 in Table 2, the final moisture content in the sludge depends on the temperature of the  
258 drying method. At high temperature (70°C, 105°C), the content of moisture is low, but  
259 always higher than that of the  $\text{MgSO}_4 \cdot \text{H}_2\text{O}$  method. Furthermore, it was observed for

260 experiments with oven at 70°C, freeze-dried, fume hood, and sun drying that the solid  
261 particles were more compacted, even after grinding. The water content could surround  
262 the sludge particles and thus inhibit the good penetration of hexane inside the solid  
263 particle. As Table 2 shows, in general, the greater the amount of moisture contained in  
264 the sludge, the lower the amount of extracted and esterified lipids.

265 Additionally, it can also be noted in Table 2 that the drying methods at high  
266 temperature (70°C, 105°C) had a negative impact on lipid composition giving lower  
267 saponifiable matter, thereby decreasing the potential biodiesel yield when compared to  
268 standard method. Despite the lower temperature used for fume hood and sun drying, the  
269 lipid content extracted from these dried sludge decreased to 12.3% and 11.4%,  
270 respectively, again compared to standard  $\text{MgSO}_4 \cdot \text{H}_2\text{O}$  method, 26.3%. Finally, freeze-  
271 drying also showed a significant loss of extracted lipids (11.2%) but, in contrast, the  
272 rate of saponifiable matter was higher (57.3%) than those from oven at 70°C (53.9%),  
273 sun (45.5%) and fume hood (44.8%) drying methods. The low lipid content extracted  
274 from dried sludge at low temperature and freeze-drying reported here are in agreement  
275 with Cordero Esquivel et al. (1993). They reported that biomass drying by both freeze-  
276 drying and oven drying at low temperature (30 °C) caused an approximately 70% loss  
277 of total lipid content [8].

278 The biodiesel yield regarding  $\text{MgSO}_4 \cdot \text{H}_2\text{O}$  method showed a decrease in all  
279 methods, -17% for oven at 105°C, -53% for oven at 70°C, -65% for freeze-dryer, -71%  
280 for fume hood and -73% for sun drying. On the other hand, the values of biodiesel  
281 obtained from primary sludge dried by oven at 70°C (8.8%) and fume hood (5.5%)  
282 compares well to those reported elsewhere [3, 6].

283 The influence of sludge drying methods on the fatty acid composition was also  
284 studied and the results are collected in Table 3. The same fatty acids were found for all  
285 methods showing a significant amount of palmitic (31.1 to 49.4%), oleic (18.3 to

286 32.6%) and stearic (8.3 to 15.8%) acids in the sludge biodiesel. The most important  
287 differences in the composition were observed for palmitic, palmitoleic, stearic, oleic  
288 and linoleic acids. In detail, oven at 105°C for two days gave the fatty acid composition  
289 almost identical with respect to  $\text{MgSO}_4 \cdot \text{H}_2\text{O}$  method. The other methods showed an  
290 increase in the fraction of oleic, linoleic and palmitoleic acids, counterbalanced by a  
291 decrease of palmitic and stearic acids. This trend, where the fraction of saturated fatty  
292 acids decreased while the fraction of unsaturated fatty acids increased, is particular for  
293 the sludge. Usually, unsaturated fatty acids are more unstable and readily oxidized than  
294 saturated ones.

295 Definitely, usual sludge drying methods adversely affect the yield of extracted lipids  
296 as well as the lipid saponifiable matter, consequently reducing the potential for  
297 biodiesel production. FAMES composition of biodiesel is also modified, too. Among  
298 the methods tested, oven drying at 105°C could be the best option for subsequent  
299 biodiesel production, giving 15.7% of biodiesel produced from a dried sludge basis.  
300 Unfortunately, thermal drying is not cost effective for a large scale application.

301

### 302 **3.3. Sequential liquid-liquid extraction**

303

304 As above commented, it is surprising that liquid-liquid extraction has not yet been  
305 applied to lipid fraction recovery from sludge as this is a fair alternative allowing the  
306 scale-up into a continuous process. In first place, sequential batch liquid-liquid  
307 extraction was performed to examine its feasibility and evaluate the effect of operation  
308 variables. The determination of the partition coefficient for some chosen experiments  
309 will allow the scale-up of the extraction step in a continuous process.

310

#### 311 **3.3.1. Effect of sludge acidification**

312

313 Despite the fact that primary sludge was found to contain the highest lipid content as  
314 compared to secondary and blended (Table 1), the liquid-liquid extraction was also  
315 studied for these sludges, in order to evaluate the suitability of the novel extraction  
316 method for all type of sludge generated in WWTPs.

317 The first step to be evaluated was the effect of previous sludge acidification with  
318 concentrated hydrochloric acid. It is expected that the acidification will facilitate the  
319 extraction of lipids from processed samples, so as well the amount of saponifiable lipid,  
320 thereby the biodiesel yield.

321 Fig. 2 shows the results of the sequential liquid-liquid extraction for acidified and  
322 non-acidified primary, secondary and blended sludges. The accumulated lipid yield was  
323 continually increasing in each extraction stage in all cases. As it was expected, primary  
324 sludge achieved the highest lipid yield followed by blended and secondary, irrespective  
325 of sludge acidification. Sludge acidification highly increased the lipid yield in each  
326 extraction stage. Nonetheless, this trend is more evident in the case of primary and  
327 blended sludge. In the last extraction stage, the lipid yield obtained from primary sludge  
328 was 26.6% and 13.0% for acidified and non-acidified samples, respectively, whereas  
329 blended sludge gave 19.1% and 10.7 for acidified and non-acidified samples,  
330 respectively. Secondary sludge achieved the lowest lipid yield, a meagre 6.3% and 5.1%  
331 for acidified and non-acidified sludge, respectively. The high difference between the  
332 values obtained for primary and blended sludge, with and without acidification owns to  
333 the fact that municipal wastewater contains fatty acids from commercial soaps,  
334 potassium and sodium from household cleaning products, cosmetics, lubricant and  
335 coatings. During primary treatment, the physico-chemical process leads to a rapid  
336 formation of relatively insoluble calcium and magnesium salts precipitating during the  
337 primary wastewater treatment, which remain in the primary sludge [5]. For this reason,

338 the acidification was responsible of the conversion of insoluble soaps into FFAs that are  
339 soluble in the extract solvent, increasing the lipid yield and the saponifiable matter [5].  
340 Since the secondary sludge does not contain insoluble soaps that could be converted  
341 into FFAs, which significantly raises the lipid content, the lipid fraction in secondary  
342 sludge mainly comes from microorganism cells. Thus, the acidification can only release  
343 by acid hydrolysis some additional lipids bonded to the cells, slightly increasing the  
344 lipid yield.

345 The results of FFAs analyses in primary sludge showed that, after sludge  
346 acidification, the FFAs content increased from 39.2% to 68.7% (on the basis of lipids)  
347 showing a good agreement with previous literature data [5]. Moreover, the increase of  
348 FFAs content resulted on significant increase of saponifiable (esterifiable) lipids (from  
349 45.3% to 70.0%), which accounts for the rise of biodiesel production from 5.9% to  
350 18.6% (on the basis of dry sludge).

351 It should be noted that the final lipid yield obtained by liquid-liquid extraction from  
352 acidified primary sludge (26.6%, Fig. 2) was higher than the yield obtained by standard  
353  $\text{MgSO}_4 \cdot \text{H}_2\text{O}$  method (25.2%, Table 1). Therefore, for the first time a process that can  
354 be easily scaled-up, i.e. liquid-liquid extraction, is able to extract all lipid contained in  
355 the primary sludge as the standard method does.

356 On the other hand, the acidified blended and secondary sludge gave lower lipid yield  
357 than that attained by standard  $\text{MgSO}_4 \cdot \text{H}_2\text{O}$  method. The absolute lipid yields obtained  
358 by the liquid-liquid extraction from acidified blended and secondary sludges were  
359 19.1% and 6.3%, respectively (Fig. 2). These values are 10% and 20% less,  
360 respectively, than those achieved by the standard  $\text{MgSO}_4 \cdot \text{H}_2\text{O}$  method (Table 1).  
361 Hence, the liquid-liquid extraction from acidified blended and secondary sludge is not  
362 so effective to extract lipids present in these sludges. Because liquid-liquid extraction  
363 from acidified primary sludge is more favourable than from acidified blended and

364 secondary, the optimization of liquid-liquid extraction was conducted over the primary  
365 sludge.

366

### 367 **3.3.2. Optimisation of liquid-liquid extraction from primary sludge**

368

369 The extraction optimisation from acidified primary sludge was carried out using a  
370 combination of different time of contact in each stage (20, 40 and 60 min) and different  
371 sludge to hexane volume ratio (4:1, 2:1, 1:1 and 1:2). The other operative conditions  
372 were maintained constant, i.e. 200 mL of sludge, 9 consecutive extractions, 330 rpm  
373 agitation speed and ambient temperature.

374 Fig. 3 shows the results of the optimisation of the lipid extraction. In all cases, the  
375 accumulated yields of lipids increased with consecutive extraction stages, reaching a  
376 constant value at the last stages of the extraction. The best value of the accumulated  
377 yield of lipids at the last stage of extraction was 29.6% (based on dry sludge), attained  
378 for the experiment with a sludge to hexane volume ratio 1:2. The 1:1 volume ratio was  
379 able to achieve 29.5%, the 2:1 gave 28.8% and the 4:1 only 28.1%. As expected, the  
380 lower the amount of solvent, the lower the extraction efficiency achieved.

381 The contact time is also of great importance. For the sludge to hexane volume ratios  
382 4:1, 2:1 and 1:1, the lipid yield grew as the contact time increased in each extraction  
383 stage, always reaching the best results for 60 min, 28.1%, 28.8%, and 29.5%,  
384 respectively. In turn, 40 min of extraction time allowed attaining 26.3%, 27.8% and  
385 28.4% of lipids for sludge to hexane volume ratios 4:1, 2:1 and 1:1, respectively. The  
386 lowest lipid yield was obtained for 20 min of extraction time giving 22.9%, 26.7% and  
387 26.6% of lipids for sludge to hexane volume ratio 4:1, 2:1, 1:1, respectively. On the  
388 contrary, the results using sludge to hexane volume ratio 1:2 did not show any influence  
389 of the extraction time. Beyond the third stage, the accumulated lipid yields remained

390 practically unaltered, reaching 27% of lipids based on dry sludge. This value represents  
391 91% of the attainable lipid recovery.

392 Independently of the extraction time, as it was expected, the yield of lipids increased  
393 after each extraction stage when increasing the amount of solvent. However, for volume  
394 ratios of 2:1, 1:1 and 1:2, 60 min of extraction time did not show significant difference  
395 between the lipid yields, after the third stage of extraction. This suggests that the  
396 quantity of hexane used for a ratio 2:1 for 60 min was enough to achieve 27% of lipids  
397 based on dry sludge (91% of the attainable lipid value).

398 Overall, in order to reach at least 91% of lipids present in the primary sludge, only  
399 three consecutive extraction stages were needed. This extraction efficiency was  
400 achieved for 60 min (2:1, 1:1, 1:2, sludge:hexane) as well as for 20 and 40 min (1:2,  
401 sludge:hexane). Taking into account the minimization of solvent used, the best  
402 operation conditions are 60 min using a 2:1 sludge:hexane ratio. On the other hand,  
403 minimizing the extraction time, the best operation conditions are a 1:2 sludge:hexane  
404 ratio for 20 min of extraction time in each stage.

405

### 406 **3.3.3. Scale-up of liquid-liquid extraction process**

407

408 Cost-effective production of biodiesel requires continuous operation plants.  
409 Therefore, design and scale-up of continuous processes must be done from batch data  
410 and operation. Lipid recovery data, starting from acidified primary sludge, obtained  
411 through batch liquid-liquid extraction experiments allow determining partition  
412 coefficients in a wide range of process conditions. Fig. 4 presents an example of the  
413 equilibrium curve obtained in the experiment with these operative conditions: 200 mL  
414 of sludge, 400 mL of hexane, 20 minutes for each extraction, 9 consecutive extractions  
415 with fresh hexane, 330 rpm of agitation speed and ambient temperature. As the liquid-



416 liquid equilibrium thermodynamic diagrams are then available, application of design  
417 methods for typical extraction equipment gives optimised solvent to feed flowrate ratio  
418 and number of stages in continuous operation.

419

#### 420 **3.4. Biodiesel produced from primary sludge by liquid-liquid extraction**

421

422 The results of lipid transesterification from liquid-liquid extraction were compared  
423 with the results from drying by standard  $\text{MgSO}_4 \cdot \text{H}_2\text{O}$  method in order to verify that the  
424 process did not affect the yield of biodiesel and the composition of FAMES. The  
425 optimised liquid-liquid extraction gave  $26.7 \pm 0.1\%$  of lipid (on the basis of dry  
426 sludge), the saponifiable obtained after transesterification was  $72.0 \pm 3.0\%$  (on the basis  
427 of lipid) and the value of biodiesel was  $19.2 \pm 0.1\%$  (on the basis of dry sludge). These  
428 values are higher than those obtained by standard method ( $25.2 \pm 0.2\%$  of lipid,  $69.7 \pm$   
429  $0.7\%$  of saponifiable and  $17.6 \pm 0.2\%$  of biodiesel).

430 Based on the present research, i.e. experimental biodiesel yield of 19.2% from dry  
431 primary sludge basis, the annual biodiesel potential, theoretically produced from  
432 primary sludge generated at WWTP of Reus (1922 ton/year of dry primary sludge  
433 generation), was estimated to be 369 ton. Speculating the biodiesel production from  
434 wastewater produced from all Spanish population,  $6 \text{ hm}^3/\text{day}$ , the annual biodiesel  
435 potential was estimated at 88664 ton. This value may replace approximately 15% of  
436 current biodiesel production in Spain [12].

437 The FAMES composition of biodiesel produced from standard and liquid-liquid  
438 extraction methods is presented in Table 3. At least 12 fatty acids were identified for  
439 both methods, ranging from C12 to C22 with a predominance of palmitic, oleic and  
440 stearic acids. As it can be observed in Table 3, the composition of the two biodiesel is  
441 the same, the differences observed being not essential. This fact is critical as the

442 acidification and liquid-liquid extraction did not affect the characteristics of the  
443 biodiesel produced, making the liquid-liquid extraction technology viable.

444 In addition, the properties of biodiesel strongly depend on the fatty acid composition.  
445 The fact that the amount of polyunsaturated fatty acids found in the sludge lipids is  
446 really low, around 2% of linoleic acid (C18:2), is an advantage in comparison to the  
447 common vegetable oil feedstocks, which contain a large amount of polyunsaturated  
448 fatty acids. The polyunsaturated fatty acids are very susceptible to auto-oxidation,  
449 resulting in a poor oxidation stability of the biodiesel. On the other hand, the high level  
450 of saturated fatty acids found in the sludge, more than 60%, could represent a problem  
451 for the cold flow properties of biodiesel, when it becomes cloudy due to formation of  
452 crystals and solidification of saturated compounds. This could be solved by the  
453 presence of branched-chain and hydroxy fatty acid monoalkyl esters [13-14]. Actually,  
454 these compounds exist and were included as “Others” in Table 3. This fraction was  
455 identified by GC-MS (data not shown) and mainly consists of hydroxy and oxy fatty  
456 acids and branched-chain fatty acid methyl esters. This suggests that, despite the high  
457 amount of saturated fatty acids, the cold flow properties of biodiesel produced from  
458 primary sludge could be even better because of the presence of other fatty acids methyl  
459 ester.

460

### 461 **3.5. Economic estimation of biodiesel production from primary sludge by liquid-** 462 **liquid extraction (laboratory case).**

463

464 The economic evaluation of biodiesel production cost from municipal sewage sludge  
465 has been already carried out elsewhere [4]. This study estimated the price of biodiesel  
466 about 0.83\$/L, taking into account the cost of sludge dewatering and subsequent drying,  
467 which represent about 42-53% of the overall biodiesel production cost. However, in

468 order to avoid the influence of currency, the energy required for the production of 1kg  
469 of FAMES is a better parameter to estimate the final cost [5]. In this study, the  
470 minimum specific energy demand was estimated to be 17 MJ/kg<sub>FAMES</sub> but the result was  
471 given without considering the energy needed for sludge dewatering, which should have  
472 been added to this value.

473 In the present study, in order to perform the economic evaluation of biodiesel  
474 production from primary sludge by liquid-liquid extraction, all different process  
475 operations involving energy demand were included: agitation during extraction and  
476 settling, evaporation of the extract solvent, heating of the esterification mixture,  
477 evaporation of the product mixture, and separation of FAMES by solvent extraction.  
478 Table 4 shows the values used to calculate the specific energy demand and the results of  
479 the economic estimation of biodiesel production based on the experimental results for  
480 the following extraction conditions: 60 min, 2:1 sludge to hexane volume ratio. As  
481 shown in Table 4, the energy demand and the price per litre of FAMES depend on the  
482 number of extraction stages, varying between 60.95 MJ/kg<sub>FAMES</sub>, 1.88 €/L<sub>FAMES</sub> (1  
483 stage) to 290.15 MJ/kg<sub>FAMES</sub>, 8.94 €/L<sub>FAMES</sub> (9 stages). In a continuous process, no more  
484 than three extraction stages should be used to gain 99% of lipids. In addition, scaling-up  
485 of the process from lab-scale to industrial plant should reduce the price.

486 It should be also stated that the values calculated in the present study are final,  
487 including all different operation steps in the production of biodiesel from primary liquid  
488 sludge, and any additional cost of drying or dewatering are not necessary to include it in  
489 the final cost. On the other hand, costs of methanol, hexane and HCl used in the overall  
490 process were not accounted as it was an energetic balance calculation.

491 Finally, it must be noted that the production of biodiesel from primary sewage  
492 sludge reduces the amount of sludge generated at the WWTP facility, which should  
493 subsequently be managed and disposed as a waste. As the above is a major cost in the

494 WWTP operation, this saving should be taken into consideration when calculating the  
495 final price of biodiesel.

496

### 497 **3.6. Anaerobic digestion of lipid-extracted primary sludge**

498

499 As residual sludge after lipid extraction still contains a large amount of organic  
500 matter, this lipid exhausted sludge was subjected to anaerobic digestion in order to  
501 evaluate the remnant potential for biogas generation. Fig. 5 shows the biogas  
502 production during anaerobic digestion of evaporated lipid-extracted sludge (ELES), i.e.  
503 hexane free, and lipid-extracted sludge (LES). The biogas measure was converted at  
504 standard conditions (0°C and 1 atm) and is given as the volume of biogas per gram of  
505 VS fed ( $\text{mL}_{\text{Biogas}}/\text{g}_{\text{VS}}$ ). Biogas production from ELES reached  $365 \pm 10 \text{ mL}_{\text{Biogas}}/\text{g}_{\text{VS}}$ ,  
506 whereas LES only reached  $31 \pm 4 \text{ mL}_{\text{Biogas}}/\text{g}_{\text{VS}}$ . This huge difference, over tenfold, can  
507 be attributed to the presence of hexane in LES. In a mass balance, it was calculated that  
508 solvent still represented approximately 9% of the volume in LES. Furthermore, a VFA  
509 analysis revealed a concentration of  $12.0 \pm 0.1 \text{ mol}/\text{m}^3$  in the reactor with LES, while  
510 no VFA were detected in reactors with ELES. A value over the range 6.7-9.0  $\text{mol}/\text{m}^3$   
511 has been reported to be toxic for methanogenic microorganisms, stopping the biogas  
512 production [15].

513 Methane content in biogas from ELES was 62%, whereas in LES was barely a 31%.  
514 The theoretical methane production based on sludge composition was estimated,  
515 following Buswell's equation [11], in  $486 \text{ mL}_{\text{CH}_4}/\text{g}_{\text{VS}}$  for sludge after lipid extraction.  
516 Based on the experimental methane production, biodegradability (expressed as the ratio  
517 measured to theoretical methane production) resulted 47% and 4% for ELES and LES,

518 respectively. The 47% of biodegradability is in line with the conversion that can be  
519 expected from highly particulated and structured organic matter [11].

520 Hence, it can be concluded that the lipid-extracted sludge can be easily anaerobically  
521 digested with good biogas production, although the elimination of residual hexane is  
522 required. In the proper conditions, this solvent could be recovered and reused for the  
523 extraction step. As the anaerobic digestion is widely installed in WWTPs, the hexane  
524 free residual sludge after lipid extraction could be returned to the WWTP to be  
525 anaerobically stabilised giving additional energy in form of biogas.

526

#### 527 **4. CONCLUSIONS**

528

529 Common sludge drying methods decrease the yield of lipids as well as the  
530 saponifiable fraction, thus reducing the biodiesel production. In addition, they require  
531 high energy input. The proposed alternative, liquid-liquid extraction using hexane, is  
532 feasible and compares well with those classical methods. Previous sludge acidification  
533 improves lipid and subsequently biodiesel yields. The FAMEs obtained from liquid  
534 extracted lipids are similar to those obtained by standard method.

535 The cost of the proposed liquid-liquid extraction process and the lipid yield depend  
536 on the number of extraction stages. The scale-up of the process should allow reducing  
537 the final biodiesel price, as the cost of drying is eliminated. Finally, the lipid extracted  
538 sludge can be used to produce biogas by anaerobic digestion, avoiding the generation of  
539 a new sludge. The biogas obtained maintains a similar composition, i.e. quality, than  
540 that coming from raw excess sludge.

541

#### 542 **References**

543

- 544 [1] E.E. Kwon, S. Kim, Y.J. Jeon, H. Yi, Biodiesel production from sewage sludge:  
545 new paradigm for mining energy from municipal hazardous material. *Environ. Sci.*  
546 *Technol.* 46 (2012) 10222-10228.
- 547 [2] E. Revellame, R. Hernandez, W. French, W. Holmes, E. Alley, Biodiesel from  
548 activated sludge through in situ transesterification. *J. Chem. Technol. Biotechnol.*  
549 85 (2010) 614-620.
- 550 [3] R.M. Willson, Z. Wiesman, A. Brenner, Analyzing alternative bio-waste feedstocks  
551 for potential biodiesel production using time domain (TD)-NMR. *Waste Manage.*  
552 30 (2010) 1881-1888.
- 553 [4] S. Dufreche, R. Hernandez, T. French, D. Sparks, M. Zappi, E. Alley, Extraction of  
554 lipid from municipal wastewater plant microorganisms for production of biodiesel.  
555 *J. Am. Oil Chem. Soc.* 84 (2007) 181-187.
- 556 [5] C. Pastore, A. Lopez, V. Lotito, G. Mascolo, Biodiesel from dewatered wastewater  
557 sludge: A two-step process for a more advantageous production. *Chemosphere* 92  
558 (2013) 667-673.
- 559 [6] M.N. Siddiquee, S. Rohani, Experimental analysis of lipid extraction and biodiesel  
560 production from wastewater sludge. *Fuel Process. Technol.* 92 (2011) 2241-2251.
- 561 [7] E. Revellame, R. Hernandez, W. French, W. Holmes, E. Alley, R. Callahan,  
562 Production of biodiesel from wet activated sludge. *J. Chem. Technol. Biotechnol.*  
563 86 (2011) 61-68.
- 564 [8] B. Cordero Esquivel, D. Voltolina Lobina, F. Correa Sandoval, The biochemical  
565 composition of two diatoms after different preservation techniques. *Comp.*  
566 *Biochem. Phys. B: Comp. Biochem.* 105 (1993) 369-373.
- 567 [9] J.H. Mussnug, V. Klassen, A. Schlüter, O. Kruse, Microalgae as substrates for  
568 fermentative biogas production in a combined biorefinery concept. *J. Biotechnol.*  
569 150 (2010) 51-56.

- 570 [10] E.W. Rice, R.B. Baird, A.D. Eaton, L.S. Clesceri, Standard Methods for the  
571 Examination of Water and Wastewater 22<sup>nd</sup> (2012) ed. APHA AWWA WEF,  
572 Washington.
- 573 [11] I. Angelidaki, W. Sanders, Assessment of the anaerobic biodegradability of  
574 macropollutants. Rev. Environ. Sci. Biotechnol. 3 (2004) 117-129.
- 575 [12] European Biodiesel Board. <http://www.ebb-eu.org/stats.php>. Accessed 26 April  
576 2014.
- 577 [13] O.D. Dailey Jr, N.T. Prevost, G.D. Strahan, Synthesis and structural analysis of  
578 branched-chain derivatives of methyl oleate. J. Am. Oil Chem. Soc. 85 (2008) 647-  
579 653.
- 580 [14] O.D. Dailey Jr, N.T. Prevost, G.D. Strahan, Conversion of methyl oleate to  
581 branched-chain hydroxy fatty acid derivatives. J. Am. Oil Chem. Soc. 86 (2009)  
582 1101-1114.
- 583 [15] L. Appels, J. Baeyens, J. Degrève, R. Dewil, Principles and potential of the  
584 anaerobic digestion of waste-activated sludge Review Article. Prog. Energ.  
585 Combust. 34 (2008) 755-781.

586

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588

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595

596 **Figure captions**

597

598 **Figure 1.** Diagram of the wastewater treatment plant (a) and schematic diagram of the  
599 experimental liquid-liquid extraction setup carried out in the present study (b).

600

601 **Figure2.** Effect of sludge acidification on the lipid yield. Conditions: 1:1 sludge to  
602 hexane volume ratio, each stage extraction time - 20 min.

603

604 **Figure 3.** Effect of extraction time on the lipid yields with different sludge to hexane  
605 volume ratio.

606

607 **Figure 4.** Equilibrium curve lipid in hexane – lipid in sludge. 200 mL of acidified  
608 primary sludge, 400 mL of hexane, 20 minutes for each extraction, 9 consecutive  
609 extractions, 330 rpm of agitation speed and ambient temperature.

610

611 **Figure 5.** Biogas production from lipid-extracted sludge with and without evaporation  
612 process. Batch reactors, 33°C and 25 days.

613

614

615

616



1 **Table 1.** Characteristics of sludge used for different experiments in this work.

Sludge type	Experiment type	TS (%)	VS (%)	Lipid <sup>(a)</sup> (%)
Primary <sup>(b)</sup>	Sludge drying	3.9 ± 0.1	2.9 ± 0.1	26.3 ± 0.5
Primary <sup>(c)</sup>	Liquid-liquid extraction	3.4 ± 0.1	2.7 ± 0.1	25.2 ± 0.2
Secondary <sup>(c)</sup>	Liquid-liquid extraction	3.8 ± 0.1	3.2 ± 0.1	7.7 ± 0.1
Blended <sup>(c)</sup>	Liquid-liquid extraction	3.5 ± 0.1	2.7 ± 0.1	21.1 ± 0.2

<sup>(a)</sup> Extraction according to standard MgSO<sub>4</sub>·H<sub>2</sub>O method, lipid yield on the basis of dry sludge

<sup>(b)</sup> Lipid extracted from not acidified sludge

<sup>(c)</sup> Lipid extracted from acidified sludge

Values are means ± SD, n = 3

2

3

4 **Table 2.** Effect of drying method on the moisture content, lipid and transesterification yields.

Sludge drying method	Moisture (%)	Lipid <sup>(a)</sup> (%)	Saponifiable <sup>(b)</sup> (%)	Biodiesel <sup>(a)</sup> (%)
MgSO <sub>4</sub> ·H <sub>2</sub> O	0.0 ± 0.0	26.3 ± 0.5	71.8 ± 2.4	18.9 ± 0.6
Oven at 105°C	3.4 ± 0.0	26.5 ± 0.2	59.1 ± 0.6	15.7 ± 0.2
Oven at 70°C	6.0 ± 0.4	16.4 ± 0.1	53.9 ± 0.8	8.8 ± 0.1
Freeze-dryer	6.6 ± 0.4	11.2 ± 0.3	57.3 ± 0.8	6.4 ± 0.1
Fume hood	7.6 ± 0.2	12.3 ± 0.1	44.8 ± 1.4	5.5 ± 0.2
Sun	10.8 ± 1.2	11.4 ± 0.2	45.5 ± 0.1	5.2 ± 0.1

<sup>(a)</sup> Lipid and biodiesel yield on the basis of dry sludge

<sup>(b)</sup> Transesterification yield on the basis of lipid

Values are means ± SD, n = 3

5

6

7 **Table 3.** Fatty acids composition of biodiesel produced from primary sludge (average of 3  
 8 experiments).

FAME from fatty acid	(% weight/weight <sub>sample</sub> (SD < 0.1))							
	Primary sludge 1 <sup>(a)</sup>					Primary sludge 2 <sup>(b)</sup>		
	MgSO <sub>4</sub>	105 °C	70 °C	Fre.-dryer	F. hood	Sun	MgSO <sub>4</sub>	Liq-liq
Lauric (C12:0)	0.8	0.8	1.0	1.0	1.0	1.0	1.0	1.1
Myristic (C14:0)	4.6	4.8	4.5	3.4	4.1	4.1	4.1	4.3
Pentadecanoic (C15:0)	0.6	0.7	0.6	0.4	0.5	0.5	0.5	0.5
Palmitic (C16:0)	48.5	49.4	38.1	27.4	31.6	31.1	42.8	41.0
Palmitoleic (C16:1)	1.2	1.3	1.9	3.2	2.2	2.6	2.5	2.5
Heptadecanoic (C17:0)	0.4	0.4	0.3	0.3	0.3	0.3	0.4	0.4
Stearic (C18:0)	15.6	15.8	12.1	8.3	9.6	9.6	13.4	12.6
Oleic (C18:1)	18.3	18.3	28.8	39.6	32.8	32.6	23.3	25.7
Linoleic (C18:2)	2.1	0.6	3.4	7.2	5.3	5.2	1.9	2.0
Arachidic (C20:0)	0.4	0.2	0.4	0.3	0.3	0.3	0.4	0.4
Eicosenoic (C20:1)	-	-	0.5	0.6	0.7	0.7	0.3	0.3
Behenic (C22:0)	0.5	0.6	0.5	0.4	0.4	0.4	0.5	0.4
Others	7.0	7.1	7.9	7.9	11.2	11.7	9.0	8.8

<sup>(a)</sup> Primary sludge used for drying experiment

<sup>(b)</sup> Primary sludge used for liquid-liquid extraction experiment

9

10

11 **Table 4.** Energy and economic evaluation of biodiesel production from primary municipal sludges through liquid-liquid extraction of lipids.

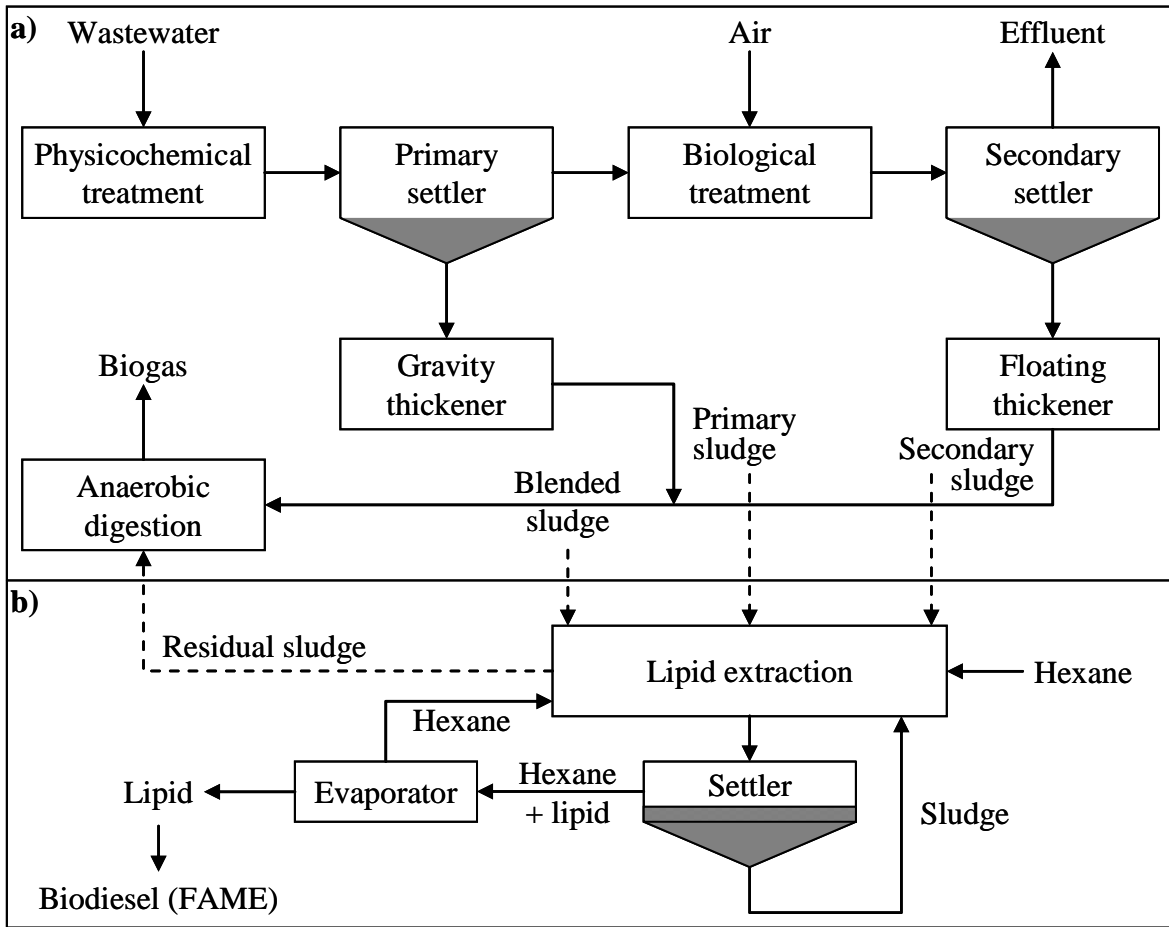
Process	Basis for energy calculation									Energy values
Extraction: Mixing	200 mL sludge, 100 mL hexane, 330/2000 rpm, 50 W, 60 min/stage									$n^{\circ}$ stages $\times$ 29.700 kJ
Extraction: Settling	200 mL sludge, 100 mL hexane, 60/2000 rpm, 50 W, 15 min/stage									$n^{\circ}$ stages $\times$ 1.350 kJ
Extraction: Evaporation of hexane	$\Delta H_{\text{vap}}$ : 0.335 kJ/g, $\rho$ : 0.655 g/mL, 100 mL/stage									$n^{\circ}$ stages $\times$ 21.94 kJ
Reaction: Heating of methanol	$C_p$ : $2.53 \cdot 10^{-3}$ kJ/g·K, $\rho$ : 0.792 g/mL, 2 mL, T: 323.15K									0.120 kJ
Reaction: Evaporation of methanol	$\Delta H_{\text{vap}}$ : 1.099 kJ/g, $\rho$ : 0.792 g/mL, 2 mL									1.741 kJ
Separation FAMES by hexane	$\Delta H_{\text{vap}}$ : 0.335 kJ/g, $\rho$ : 0.655 g/mL, 10 mL									2.194 kJ
Extraction stage	1	2	3	4	5	6	7	8	9	
FAMES recovered (g)	0.94	1.35	1.53	1.60	1.62	1.63	1.64	1.65	1.66	
Total Energy (kJ)	57	110	163	216	269	322	375	428	481	
Specific Energy (MJ/kg <sub>FAME</sub> )	60.95	81.50	106.41	135.15	166.56	197.33	228.68	259.44	290.15	
Price <sup>a,b</sup> (€/L <sub>FAME</sub> )	1.88	2.51	3.28	4.16	5.13	6.08	7.04	7.99	8.94	

<sup>a</sup> Energy price: 0,126 €/kW·h

<sup>b</sup> Density of FAMES (biodiesel): 0.88 kg/L

Figure 1

1

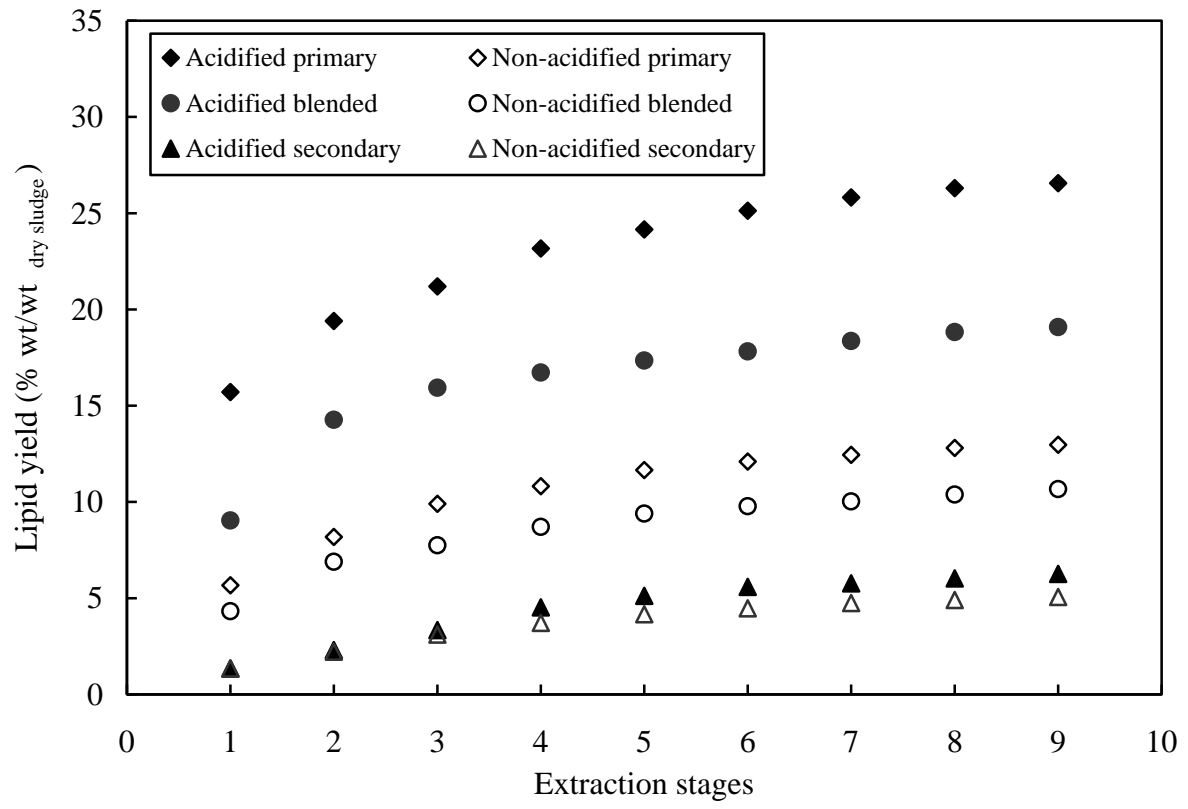


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Figure 2

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Figure 3

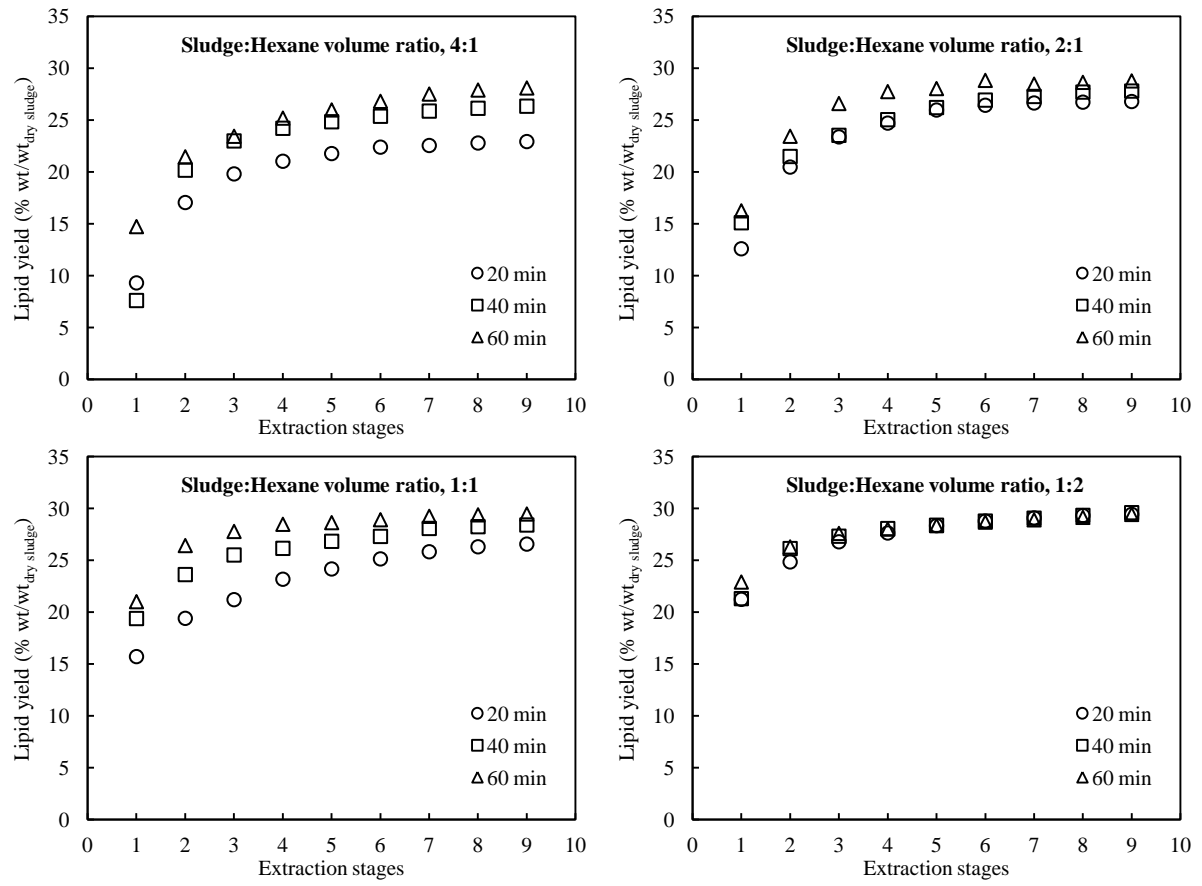


Figure 4

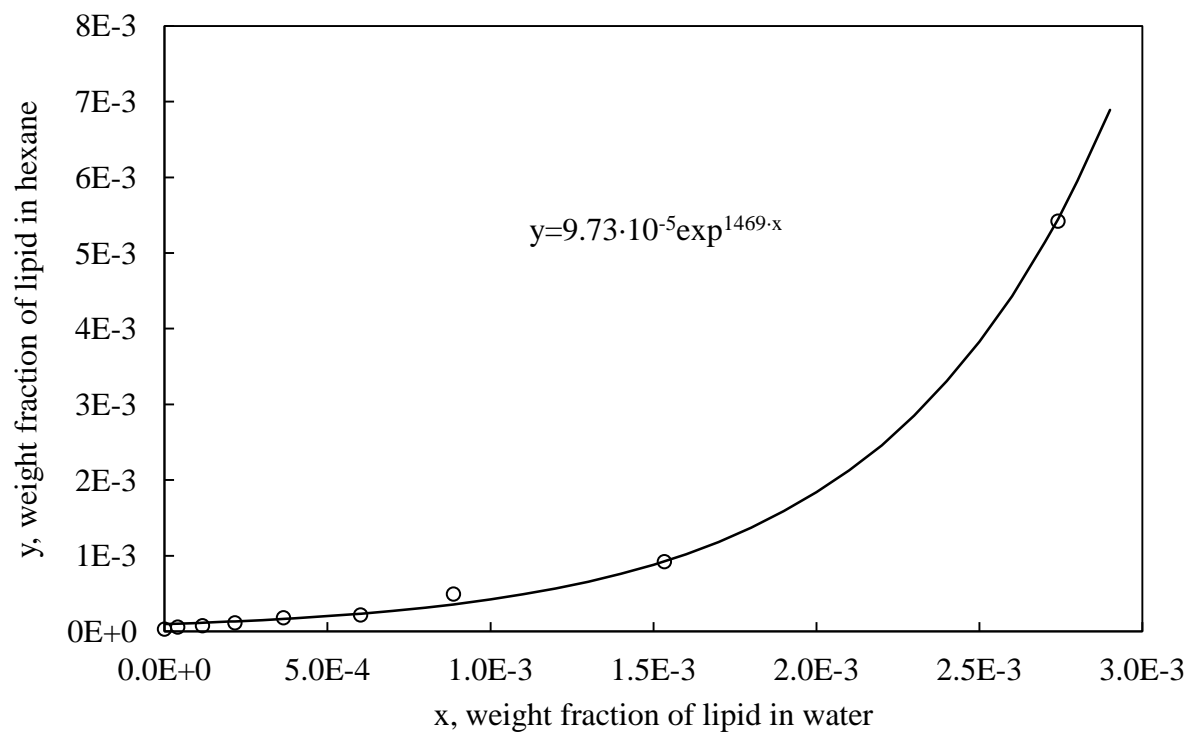
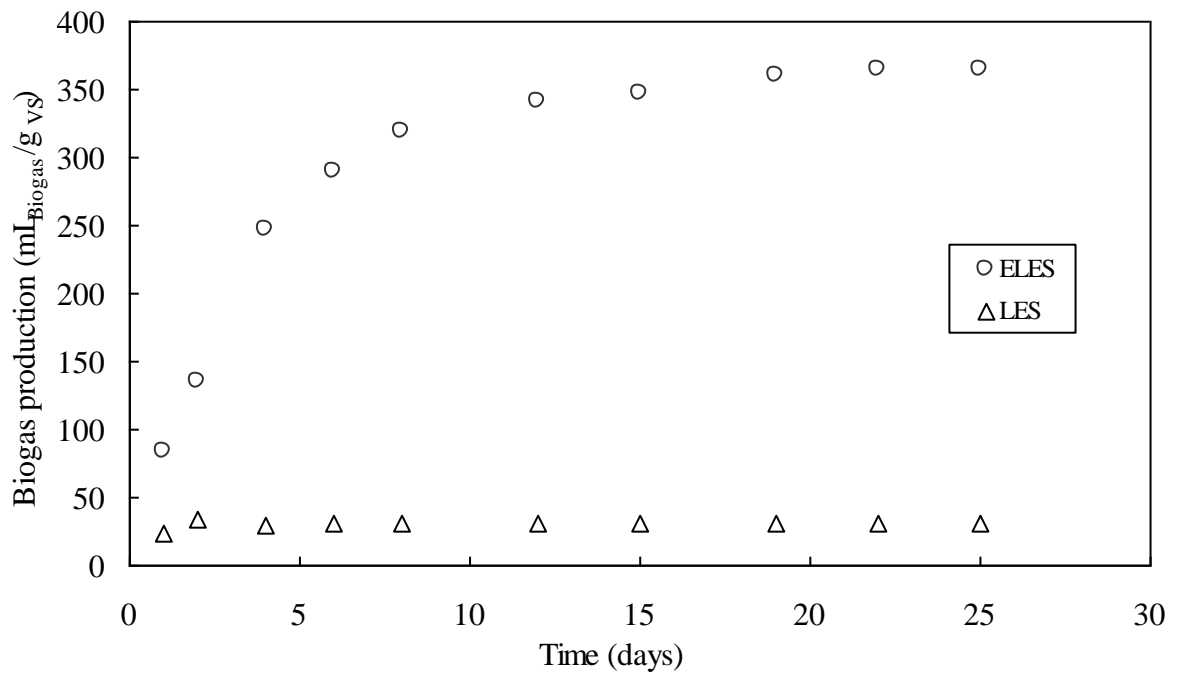




Figure 5

1



2

3

- Sludge drying methods affect the lipid yield, reducing biodiesel production.
- Direct lipid extraction from liquid sludge was successfully developed.
- The method gave high lipid and biodiesel yields after previous sludge acidification.
- Residual lipid-extracted sludge is still a good biomass for biogas production.

Graphical Abstract (for review)

