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1 **Decreasing temporal trends of polychlorinated**  
2 **dibenzo-*p*-dioxins and dibenzofurans in adipose tissue**  
3 **from residents near a hazardous waste incinerator**

4

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22

23 **ABSTRACT**

24

25 Polychlorinated dibenzo-*p*-dioxins and dibenzofurans (PCDD/Fs) are very toxic  
26 chemicals which are emitted in waste incineration and whose exposure has important  
27 adverse effects for the human health. In 2019, adipose tissue samples were collected from  
28 15 individuals with a median age of 61 years, who had been living near a hazardous waste  
29 incinerator (HWI) in Constantí (Catalonia, Spain). The content of PCDD/Fs in each  
30 sample was analyzed. The results were compared with data from previous studies,  
31 conducted before (1998) and after (2002, 2007 and 2013) the facility started to operate,  
32 and based on populations of similar age. In 2019, the mean concentration of PCDD/Fs in  
33 adipose tissue was 6.63 pg WHO-TEQ/g fat, ranging from 0.95 to 12.95 pg WHO-TEQ/g  
34 fat. A significant reduction was observed with respect to the baseline study (1998), when  
35 a mean PCDD/Fs concentration of 40.1 pg WHO-TEQ/g fat was found. Moreover, the  
36 current level was much lower than those observed in the 3 previous studies (9.89, 14.6  
37 and 11.5 pg WHO-TEQ/g fat in 2002, 2007 and 2013, respectively). The body burdens  
38 of PCDD/Fs were strongly correlated with age. The significant reduction of PCDD/Fs  
39 levels in adipose tissue fully agreed with the decreasing trend of the dietary intake of  
40 PCDD/Fs by the population of the zone (from 210.1 pg I-TEQ/day in 2018 to 8.54 pg  
41 WHO-TEQ/day in 2018). Furthermore, a similar decrease has been also observed in other  
42 biological, such as breast milk and plasma. The current data in adipose tissue, as well as  
43 those in other biological monitors, indicate that the population living near the HWI is not  
44 particularly exposed to high levels of PCDD/Fs. However, biomonitoring studies cannot  
45 differentiate the impact of the HWI emissions from food consumption patterns. This  
46 question can be only solved by conducting complementary investigations and contrasting  
47 the results of monitoring and epidemiological studies.

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49 *Keywords:* PCDD/Fs, hazardous waste incinerator, adipose tissue, temporal trend,  
50 biological monitoring.

51

## 52 **1. Introduction**

53

54 Incineration, especially without energy recovery, is not one of the most desirable  
55 methods in the EU priority hierarchy of solid waste treatment. In fact, it is located at the  
56 bottom of the current disposal alternatives, together with landfilling. In turn, prevention,  
57 reuse, recycling and recovery are considered as better options, if feasible, in the “waste  
58 hierarchy” (EC, 2010). According to data from the Eurostat (2016), 100.7 million of tons  
59 of waste were classified as hazardous waste, meaning 4% of the total waste. Hazardous  
60 waste is defined as “a waste with properties that make it dangerous or capable of having  
61 a harmful effect on human health or the environment” (US EPA, 2020). Hazardous waste  
62 is generated in a number of industrial activities, including production/management of raw  
63 chemical materials and chemical products, mining and processing of nonferrous metals  
64 ores, smelting, and petroleum processing and coking, among others (Duan et al., 2008).  
65 In 2016, in the EU-28, 51% of the hazardous waste was still landfilled, while 14% of the  
66 amount was incinerated (6% without energy recovery, and 8% with energy recovery).  
67 Socially, incineration is not well accepted socially, being an issue of great strong public  
68 opposition (Yuan et al., 2019).

69 In 1996-1998, a new HWI was built in Constantí (Tarragona County, Catalonia,  
70 Spain). At that time, it was the first HWI across Spain, and it is still being the only HWI  
71 in the country. Because of the important concern of the residents living nearby, a wide  
72 monitoring program was initiated before the plant started to operate in 1999 (Domingo et

73 al., 1999; Schuhmacher et al., 1999a,b). So far, this surveillance program has been  
74 ongoing for 20 years, with periodical samplings of biological, food and environmental  
75 samples for the analysis of polychlorinated dibenzo-*p*-dioxins and dibenzofurans  
76 (PCDD/Fs) and metals (Batista et al., 1996; Nadal et al., 2008, 2013; Schuhmacher et al.,  
77 2004b, 2009, 2013; Marquès et al., 2018). Since then, biological samples have been  
78 collected every 5 years to determine any temporal trends in the exposure levels of the  
79 local population. Between 2017 and 2019, a new set of biological samples, including  
80 breast milk and plasma were collected for determining the concentrations of PCDD/Fs  
81 (Nadal et al., 2019; Schuhmacher et al., 2019), whereas samples of autopsy tissues, blood  
82 and human hair were collected for metal analysis (Esplugas et al., 2019, 2020; García et  
83 al., 2020).

84 In the context of this biological monitoring survey, this study was aimed at  
85 determining the temporal trends of PCDD/Fs in samples of adipose tissue collected from  
86 residents who had been living near the HWI of Constantí. Data were compared with the  
87 trends of PCDD/Fs burdens in other biological monitors (e.g., breast milk and plasma),  
88 as well as the dietary intake of these chemicals.

89

## 90 **2. Materials and methods**

91

### 92 *2.1. Sample collection*

93

94 In 2019, 15 autopsy samples of adipose tissue were collected from individuals who, at the  
95 time of death, had been living near the HWI of Constantí for, at least, the last 10 years.  
96 As in previous campaigns, the cause of death varied among the 15 study subjects, but it  
97 was unrelated to dioxin exposure, according to the current epidemiologic evidence.

98 Exceptionally, and given its high scientific value, a sample of a child who died at the age  
99 of 5 years was collected in order to confirm the correlation between PCDD/Fs and age.  
100 All samples were stored in polyethylene containers and immediately kept at -20°C until  
101 the time of PCDD/Fs analyses.

102 Samples were collected from 14 men and 1 woman, whose age ranged from 5 to 86  
103 years (mean: 56 years; median: 61 years). The samples were collected in collaboration  
104 with forensic doctors from the Tarragona Division of the Institute of Forensic Medicine  
105 and Forensic Sciences of Catalonia (Spain). None of the subjects had been occupationally  
106 exposed to PCDD/Fs. The study was approved by two ethical committees. On one hand,  
107 the protocol of the biological surveillance program, number 07/2017, was reviewed and  
108 approved by the Ethical Committee for Clinical Research (CEIm) of the Pere Virgili  
109 Health Research Institute (IISPV), Reus/Tarragona, Spain. On the other hand, the specific  
110 protocol for the biomonitoring of autopsy tissues, number PR164/19, was  
111 complementarily evaluated and approved by the Clinical Research Ethics Committee  
112 (CEIC) of the Bellvitge University Hospital, Barcelona, Spain.

113

## 114 *2.2. Analytical procedure*

115

116 The concentrations of PCDD/Fs were analyzed following the US EPA method 8290.  
117 Further details on the analytical treatment have been given elsewhere (Nadal et al., 2009;  
118 Schuhmacher et al., 2014). Briefly, samples were firstly dissolved in hexane (Merck,  
119 Darmstadt, Germany), and a mixture of <sup>13</sup>C<sub>12</sub>-PCDD/Fs standards was added to control  
120 for potential losses during the extraction and cleaning processes. The lipids were  
121 destroyed by the addition of acid silica. To eliminate possible interfering components, a  
122 multi-stage clean-up process was carried out. The first step consisted of cleaning with a

123 multilayer silica column. The extract was then eluted through a basic alumina column by  
124 passing through several solvents, in order to separate the PCDD/Fs from other  
125 compounds. The PCDD/Fs fraction was collected and concentrated to dryness by a stream  
126 of nitrogen. Finally, 25  $\mu$ L of C<sub>13</sub>-labeled injection standards were added.

127 The analytical determination of PCDD/Fs was performed by high-resolution gas  
128 chromatography coupled to high-resolution mass spectrometry (HRGC-HRMS). The  
129 extract was injected into an Agilent 6890 gas chromatograph equipped with a ZB5-MS  
130 capillary column and coupled to a Waters Autospec mass spectrometer. The  
131 chromatographic process separated the 17 2,3,7,8-chlorinated PCDD/Fs congeners from  
132 the remaining compounds. Mass spectrometric parameters allowed the separation of  
133 PCDDs and PCDFs, between different degrees of chlorination, and between C<sub>13</sub>-labeled  
134 and native (C<sub>12</sub>)-labeled congeners. The mass spectrometer measured, at a resolution of  
135 10,000, two ions selected per congener group.

136 The calculation of the concentrations was performed using the corresponding level  
137 of congeners marked as internal standard, applying an automatic correction based on the  
138 potential losses occurred during the analytical process.

139

### 140 *2.3. Data processing*

141

142 Total PCDD/Fs concentrations were calculated using the most recent toxic equivalence  
143 factors (WHO-TEF) (van den Berg et al., 2006). Statistical processing of the results was  
144 performed with the SPSS 25.0 software package. First, the Levene test was applied to  
145 verify the homogeneity of the variances. Statistical significance was determined by  
146 applying the ANOVA or Kruskal-Wallis statistical test based on whether the data  
147 followed a normal probability or not, respectively. A probability less than 0.05 ( $p < 0.05$ )

148 was considered as significant. For the WHO-TEQ calculations, when a congener was  
149 below its detection limit (LOD), its concentration was assumed to be equal to half of its  
150 respective detection limit (ND = 1/2 LOD). LOD values were 0.025 pg/g fat for 2,3,7,8-  
151 TCDD, 1,2,3,7,8-PeCDD, 2,3,7,8-TCDF, 1,2,3,7,8-PeCDF, and 2,3,4,7,8-PeCDF; 0.05  
152 pg/g fat for 1,2,3,4,7,8-HxCDD, 1,2,3,6,7,8-HxCDD, 1,2,3,7,8,9-HxCDD, 1,2,3,4,7,8-  
153 HxCDF, 1,2,3,6,7,8-HxCDF, 1,2,3,7,8,9-HxCDF, and 2,3,4,6,7,8-HxCDF; 0.1 pg/g fat  
154 for 1,2,3,4,6,7,8-HpCDD, 1,2,3,4,6,7,8-HpCDF, and 1,2,3,4,7,8,9-HpCDF; and 0.5 pg/g  
155 fat for OCDD and OCDF. Recovery percentages ranged from 66% to 90%.

156 A multivariate analysis was also conducted to analyze in more detail any temporal  
157 trends on the levels of PCDD/Fs in adipose tissue, as well as potential changes of source  
158 apportionments. More specifically, a Principal Component Analysis (PCA) was executed  
159 using data collected from 1998 to 2019. This is a vastly extended method (Mari et al.,  
160 2017; Škrbić and Marinković, 2019; Lin et al., 2020), which allows to handle a great  
161 amount of data reducing them into a few variables in order to extract as much information  
162 as possible (Nadal et al., 2004). Prior to the execution of the PCA, all data were properly  
163 standardized.

164

### 165 **3. Results and discussion**

166

167 The levels of PCDD/Fs (in pg WHO-TEQ/g fat) in adipose tissue samples, obtained  
168 in autopsies of 15 individuals, are summarized in Table 1. Concentrations ranged between  
169 0.95 and 12.95 pg WHO-TEQ/g fat, with a mean of 6.63 pg WHO-TEQ/g fat and a  
170 median of 5.30 pg WHO-TEQ/g fat. Table 1 also provides additional information on the  
171 age and sex of each subject. Samples were collected from 14 men and 1 woman, aged  
172 between 5 and 86 years at the time of death (mean: 56 years; median: 61 years). The

173 median age was similar to that of the participants in the baseline survey (1998) and the  
174 first study after the HWI started to operate (2003), when the subjects showed a median  
175 age of 66 and 62 years, respectively (Schuhmacher et al., 1999c, 2004a) . In turn, it was  
176 slightly higher to those corresponding to the individuals of the campaigns performed in  
177 2008 and 2013, when the subjects were 53 and 52 years old, of median, at the time of  
178 death (Nadal et al., 2009; Domingo et al., 2017).

179 The temporal evolution of PCDD/Fs concentrations in adipose tissue in the baseline  
180 study (1998), as well as in the sampling campaigns performed in 2002, 2007 and 2013  
181 (Schuhmacher et al. 1999c, 2004a, 2014; Nadal et al., 2009; Domingo et al., 2017), are  
182 shown in Table 2. In the baseline survey (1998), the mean concentration of PCDD/Fs in  
183 adipose tissue was 40.1 pg WHO-TEQ/g fat (range: 24.2-72.1 pg WHO-TEQ/g fat). In  
184 2002, the average level of PCDD/Fs was 9.89 pg WHO-TEQ/g fat (range: 1.4-36.1 pg  
185 WHO-TEQ/g fat), while in 2007, the mean concentration was 14.6 pg WHO-TEQ/g fat  
186 (range: 3.3-55.4 pg WHO-TEQ/g fat). In the previous survey, conducted in 2013, the  
187 mean concentration of PCDD/Fs in adipose tissue was 11.5 pg WHO-TEQ/g fat (range:  
188 2.8-46.3 pg WHO-TEQ/g fat). The current (2019) values mean a non-significant decrease  
189 of 42% ( $p > 0.05$ ) compared to the immediately previous study (2013), and a statistically  
190 significant reduction of 83% ( $p < 0.001$ ) for the global period 1998-2019. The decreasing  
191 trend of PCDD/Fs concentrations in adipose tissue samples is graphically depicted in  
192 Figure 1. The concentration-time curve followed a hyperbolic trend, with data from the  
193 second (2002) to the current (2019) surveys being more than 2-fold lower than the mean  
194 level of PCDD/Fs in the baseline (1998) study. The concentration range varied from 6.63  
195 to 14.6 pg WHO-TEQ/g fat since 2002, while the mean level in 1998 was 40.1 pg WHO-  
196 TEQ/g fat. It could indicate that PCDD/Fs concentrations in these biomonitors might have  
197 reached some kind of stasis, as already found elsewhere (Bichteler et al., 2017).

198 The PCDD/Fs congener profile in adipose tissue, obtained in each sampling  
199 campaign (1998, 2002, 2007, 2013 and 2019), is shown in Figure 2. Despite there were  
200 differences in the respective levels, the profiles were very similar in all the studies. In  
201 2019, OCDD was the predominant congener, followed by 1,2,3,6,7,8-HxCDD and  
202 1,2,3,4,6,7,8-HpCDD. Regarding PCDD/Fs homologues, PCDDs showed much higher  
203 concentrations than PCDD/Fs.

204 Since 14 out of the 15 samples corresponded to men, a statistical analysis of  
205 differences according to the sex could not be carried out. In contrast to previous studies  
206 (Schuhmacher et al., 2004a; Nadal et al., 2009), PCDD/Fs concentration in adipose tissue  
207 of the woman of the present study was lower than the mean levels found in 14 men (1.89  
208 vs 7.44 pg WHO-TEQ/g fat) (Table 3). It is well established that the accumulation of  
209 PCDD/Fs in the human body increases with age (Kiviranta et al., 2005; Uemura et al.,  
210 2008; Alawi et al., 2018). In the current survey, a significant Pearson correlation ( $p$   
211  $<0.001$ ) was observed between the age and the concentrations of PCDD/Fs in adipose  
212 tissue.

213 In order to establish potential differences between samples, a PCA was applied to the  
214 results obtained in all the sampling campaigns (1998, 2002, 2007, 2013 and 2019). A  
215 three-dimensional model was obtained (Figure 3). The first principal component (PC1),  
216 which accounted for 53% of the variance, was related to PCDDs and hexa-PCDFs, while  
217 the second principal component (PC2), with 17% of the variance, was correlated with the  
218 lighter furans (2,3,7,8-Tetra-CDF and 1,2,3,7,8-Penta-CDF). Finally, the third principal  
219 component (PC3), which accounted for 11% of the variance, was related to the heaviest  
220 furans (1,2,3,4,7,8,9-Hepta-CDF and OCDF). Most samples from the baseline (1998)  
221 survey showed a high value of CP1 due to a higher concentration of PCDD/Fs. In contrast,  
222 all samples corresponding to the current (2019) study were grouped in a single "cluster".

223 Together with other biological monitors such as blood/plasma and breast milk,  
224 adipose tissue is one of the main matrices used to determine the accumulation of  
225 polychlorinated compounds in humans (Domingo et al., 2017; Jackson et al., 2017). In  
226 the last decade, only a limited number of articles focused on analyzing the burdens of  
227 PCDD/Fs in adipose tissue has been published. Moon et al. (2011) determined the  
228 concentrations and accumulation features of PCDD/Fs and dioxin-like polychlorinated  
229 biphenyls (DL-PCBs) in samples of adipose tissue collected in 2007-2008 from 53  
230 Korean women aged 40-68. The total mean concentrations of PCDDs and PCDFs were  
231 3.6 and 3.4 pg WHO-TEQ/g fat, being the PCDD/Fs level very similar to that reported  
232 here found for the population living near the HWI of Constantí (7.0 vs. 6.63 pg WHO-  
233 TEQ/g fat).

234 Most scientific investigations have been targeted on the potential association of  
235 PCDD/Fs and other dioxin-like chemicals in adipose tissue with several adverse health  
236 outcomes, including cancer, endometriosis or cardiovascular risks (Cano-Sancho et al.,  
237 2019; Ferro et al., 2019; Koual et al., 2019; Ploteau et al., 2016; Qiu et al., 2020). In  
238 France, Koual et al. (2019) recently measured the occurrence of 49 persistent organic  
239 pollutants (POPs) in adipose tissue from breast cancer patients, with or without lymph  
240 node metastasis. The median concentration of PCDD/Fs in adipose tissue from 38 women  
241 with metastatic breast cancer was 11.8 pg WHO-TEQ/g fat. However, this value was not  
242 significantly different from that observed in 53 women with non-metastatic breast cancer  
243 (11.6 pg WHO-TEQ/g fat). Notwithstanding, when analyzing individually the chemicals,  
244 it was found that 2,3,7,8-TCDD exhibited substantially higher adjusted odd ratios than  
245 the rest of PCDDs or their sum (Koual et al., 2019). The same French research group  
246 conducted a similar study focused on the associations between internal exposure levels  
247 of POPs in adipose tissue and deep infiltrating endometriosis (DIE) with or without

248 concurrent ovarian endometrioma (Ploteau et al., 2017). Adipose tissue and serum  
249 samples were collected from surgically confirmed cases (n=55) and controls (n=44)  
250 enrolled during 2013 and 2015 in Pays de la Loire, France. The levels of POPs were  
251 determined. The median concentration of PCDD/Fs in control samples was 5.53 pg  
252 WHO-TEQ/g fat, while that in samples from women with DIE was 5.80 pg WHO-TEQ/g  
253 fat. Significantly higher concentrations were found in women with DIE and ovarian  
254 endometrioma (8.75 pg WHO-TEQ/g fat;  $p < 0.05$ ). Regarding PCDD/F congeners,  
255 significant associations between DIE and adipose tissue levels of 1,2,3,7,8-PeCDD and  
256 OCDF were noticed. This conclusion agrees with the results published by Martínez-  
257 Zamora et al. (2015), who compared the concentrations of several dioxin-like substances  
258 in adipose tissue of 30 patients with DIE and those from a control group without  
259 endometriosis (n=30). Significantly ( $p < 0.05$ ) higher levels of PCDD/Fs were observed in  
260 the DIE group (6.90 pg WHO-TEQ/g fat) as compared with the control group (6.10 pg  
261 WHO-TEQ/g fat). This difference was basically due to the significant ( $p < 0.01$ )  
262 differences in the levels of 3 congeners: 2,3,7,8-TCDD, 1,2,3,7,8-PeCDD and 2,3,4,7,8-  
263 PeCDF.

264 In Jordan, Alawi et al. (2018) carried out a comparative study of several POPs  
265 concentrations in cancer-affected human organs with those of healthy organs. Thirteen  
266 adipose tissue samples of healthy people were collected from Jordan University Hospital,  
267 while 33 adipose tissue samples of cancer-affected patients were collected from King  
268 Hussein Cancer Center. PCDD/Fs concentrations in adipose tissue samples for cancer-  
269 affected patients were 3 times higher than those in healthy persons, with the congener  
270 2,3,4,7,8-PeCDFs showing the highest value. In turn, Koskenniemi et al. (2015)  
271 conducted a case-control study consisting of 44 cryptorchid cases, and 38 controls  
272 operated for inguinal hernia, umbilical hernia, or hydrocele at the Turku University

273 Hospital (Finland) or Rigshospitalet, Copenhagen (Denmark) in 2002–2006. During the  
274 operation, a subcutaneous adipose tissue biopsy was taken, and the content of PCDD/Fs  
275 was analyzed. In Finland, the median concentration of PCDD/Fs in adipose tissue from  
276 new-born boys affected by cryptorchidism was 7.44 pg TEQ/g fat, while that in healthy  
277 (control) boys was 5.43 pg TEQ/g fat. Substantially higher levels were found for  
278 Denmark, as median levels in cases and controls were 18.5 and 13.0 pg TEQ/g fat,  
279 respectively. In the US, Li et al. (2013) analyzed the burdens of PCDD/Fs in prostate  
280 cancer patients who underwent radical prostatectomy. This was performed in a large study  
281 focused on establishing whether the Agent Orange exposure in Vietnam veterans was  
282 associated to the biochemical recurrence. The median dioxin level in 37 Agent Orange-  
283 exposed patients was significantly higher than that in 56 unexposed patients (22.3 vs  
284 15.0 pg TEQ/g fat;  $p < 0.001$ ). In Japan, Watanabe et al. (2013) analyzed the levels of  
285 PCDD/Fs in the liver and adipose tissue of Japanese cadavers, and their toxicokinetic in  
286 association with hepatic cytochrome P450 (CYP) 1A protein expression levels was  
287 examined. Concentrations of PCDDs and PCDFs in adipose tissue of men from Ehime  
288 Prefecture were found to be 27 and 11 pg TEQ/g fat, respectively. In women, levels were  
289 slightly lower: 21 and 7.7 pg TEQ/g fat for PCDDs and PCDFs, respectively. Summing  
290 up, PCDD/Fs concentrations in Japanese were found to be 38 and 28.7 pg TEQ/g fat in  
291 men and women, respectively. Kim et al. (2011) characterized the POPs total body burden  
292 and their redistribution in obese individuals before and after drastic weight loss. The  
293 results were compared with a variety of molecular, biological, and clinical parameters.  
294 For that purpose, adipose tissue samples were obtained from 71 obese subjects who  
295 underwent bariatric surgery, as well as from 18 lean women. Total POPs body burden  
296 was 2-3 times higher in obese individuals. However, significantly higher concentrations

297 of PCDD/Fs were surprisingly found in adipose tissue from lean subjects compared to  
298 obese individuals (Kim et al., 2011).

299 In comparison to data from the scientific literature, the current mean value of 6.63  
300 pg WHO-TEQ/g fat, found for the population living near the HWI of Constantí, is in the  
301 lower part of the range observed in most industrialized countries. It must be remarked  
302 that values of PCDD/Fs have also decreased considerably compared to those found in  
303 studies conducted in the 90s (Takenaka et al., 2002; Kiviranta et al., 2005), a trend also  
304 recorded in other biomonitoring, such as blood (Consonni et al., 2012). Also in Spain,  
305 Zubero et al. (2017) found a dramatic decrease in the mean values of PCDD/Fs and other  
306 organochlorinated compounds in blood serum of people living near a municipal solid  
307 waste plant in Bilbao (Basque Country) between 2006 and 2013, with a 80% drop of the  
308 body burdens of PCDD/Fs. Campo et al. (2019) recently published a systematic review  
309 of peer-reviewed literature on human biological monitoring to exposure by individuals  
310 living near or working at solid waste incinerators. According to this extensive review,  
311 some earlier studies showed an increase of PCDD/Fs and other chemicals in individuals  
312 (mainly workers) exposed to emissions from old incinerators, while studies from the year  
313 2000 showed no increase of biomarkers. In addition, decreasing trends were observed in  
314 prospective studies. According to Weldon and LaKind (2016), most of the available data  
315 indicate that PCDD/F levels in humans have been declining over the years in the general  
316 population as well as in populations with excessive exposure. It confirms that the  
317 decreasing trends detected in biomonitoring collected from the population living near the  
318 HWI of Constantí has been observed elsewhere.

319 As it has been widely reported in the scientific literature, diet is the main source of  
320 human exposure to PCDD/Fs (Charnley and Doull, 2005; Llobet et al., 2008; Linares et  
321 al., 2010), with percentages of up to more than 95% of total exposure. In fact, the

322 significant reduction in PCDD/Fs levels in adipose tissue observed in the period 1998-  
323 2019 fully agrees with the decreasing trend in the dietary intake of PCDD/Fs. In the  
324 baseline survey (1998), the daily intake of PCDD/Fs by the adult population living near  
325 the HWI of Constantí was estimated in 210.1 pg I-TEQ/day (Domingo et al., 1999). A  
326 gradual reduction was observed in subsequent surveys, being 63.8, 27.8 and 33.1 pg  
327 WHO-TEQ/day in 2002, 2006 i 2012, respectively (Bocio and Domingo, 2005; Martí-  
328 Cid et al., 2008; Domingo et al., 2012). In 2018, the intake of PCDD/Fs through the diet  
329 had decreased considerably, until reaching a value of 8.54 pg WHO-TEQ/day (González  
330 et al., 2018).

331 The significant decrease in the levels of PCDD/Fs in adipose tissue of residents in  
332 the area under evaluation also agrees with the significant reduction in the concentrations  
333 of the same compounds in other biological monitors evaluated in the same surveillance  
334 program: breast milk and plasma. In 2017, the mean concentration of PCDD/Fs in 20  
335 breast milk samples was 2.26 pg I-TEQ/g fat, 81% lower than the value observed in the  
336 baseline study (11.8 pg I-TEQ/g lipid) (Schuhmacher et al., 2019). In plasma, the  
337 percentage of decrease in dioxin levels in the period 1998-2018 was 75%, being reduced  
338 from 27.0 to 6.79 pg I-TEQ/g lipid (Nadal et al., 2019). Likewise, the levels of PCDD/Fs  
339 in breast milk and adipose tissue compared to the 2013 study also decreased: 48% and  
340 58%, respectively. In turn, the concentration of PCDD/Fs in plasma had increased by a  
341 10% in the last period evaluated (2013-2018). The graphical comparison of the temporal  
342 evolution of PCDD/Fs in these three biological monitors, together with that on their  
343 dietary intake, is depicted in Figure 4.

344

#### 345 **4. Conclusions**

346

347           After 20 years of regular operations, the concentrations of PCDD/Fs in samples  
348 of adipose tissue of people who -at the time of death- had spent (at least) the last 10 years  
349 living near the HWI, significantly decreased. This reduction follows the same trend as  
350 that previously observed in other biomonitoring (e.g., breast milk and plasma) collected in  
351 2017 and 2018 in non-occupationally exposed individuals living in the same zone. The  
352 information from this large dataset, retrieved through 20 years, is an indicator that the 3  
353 compartments (adipose tissue, breast milk or plasma) may be good biological monitors  
354 for the follow-up of human exposure to PCDD/Fs. Therefore, and if needed for economic  
355 reasons, any of them could be even excluded from the surveillance program, and the loss  
356 of information would be minimal. The decrease in the body burdens of PCDD/Fs is also  
357 parallel to their dietary intake, which significantly reduced between 1998 and 2018 (from  
358 210.1 pg I-TEQ/day to 8.54 pg WHO-TEQ/day). A global analysis of the data indicates  
359 that the HWI of Constantí should not have an important contribution to exposure to  
360 PCDD/Fs for the population living nearby. However, the single execution of  
361 biomonitoring studies does not allow to determine the contribution of the HWI emission  
362 on the body burdens of PCDD/Fs, as their human exposure occurs through different  
363 pathways, mostly food consumption. Therefore, it cannot be stated that the reduction in  
364 the accumulation of PCDD/Fs would have been even more pronounced if the HWI had  
365 not been built. The only way to solve this question is by conducting an epidemiological  
366 study in the area under potential influence of this HWI, as it has been strongly demanded  
367 to the local and regional authorities (Domingo et al., 2020). Furthermore, since the HWI  
368 emissions may influence locally grown products, future studies should be conducted to  
369 assess the contribution of the emissions on the inhalation and the dietary intake of local  
370 products by the local population. Finally, the next monitoring campaigns of the  
371 surveillance program should try to overcome the limitations of their study, including the

372 use of a larger sample size, being not limited to 15. By doing this, and increasing the  
373 number of samples from women, a stratified analysis of the data will be then possible.

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382

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575

**Table 1. Concentrations of PCDD/Fs in samples of adipose tissue of 15 subjects who had been living near the HWI of Constantí.**

No. of sample	Sex	Age (years)	pg WHO-TEQ/g fat
1	Man	86	9.69
2	Man	65	10.51
3	Man	5	0.95
4	Man	45	3.79
5	Man	42	4.44
6	Man	72	11.32
7	Woman	40	1.89
8	Man	83	9.25
9	Man	59	12.95
10	Man	85	9.96
11	Man	34	3.08
12	Man	72	8.83
13	Man	62	5.30
14	Man	61	4.74
15	Man	35	2.82
Mean		56	6.63
St. Dev.		23	3.87
Median		61	5.30
Min		5	0.95
Max		86	12.95

**Table 2. Mean levels of 17 PCDD/Fs congeners (in pg/g fat) in samples of adipose tissue from 15 subjects who had been living near the HWI of Constantí.**

Congener	1998	2002	2007	2013	2019
2,3,7,8- Tetra-CDD	4.13 ± 3.03 <sup>a</sup>	1.39 ± 1.53 <sup>b</sup>	1.68 ± 1.86 <sup>b</sup>	1.24 ± 1.14 <sup>b</sup>	0.71 ± 0.31 <sup>b</sup>
1,2,3,7,8 Penta-CDD	11.37 ± 4.74 <sup>a</sup>	3.73 ± 3.51 <sup>b,c</sup>	5.28 ± 4.80 <sup>b</sup>	4.11 ± 3.48 <sup>b,c</sup>	2.41 ± 1.35 <sup>c</sup>
1,2,3,4,7,8 Hexa-CDD	5.61 ± 2.86 <sup>a</sup>	2.78 ± 1.73 <sup>a</sup>	3.30 ± 3.61 <sup>a,b</sup>	2.26 ± 2.26 <sup>b,c</sup>	1.14 ± 0.74 <sup>c</sup>
1,2,3,6,7,8 Hexa-CDD	59.4 ± 30.2 <sup>a</sup>	19.2 ± 18.9 <sup>b</sup>	28.1 ± 29.3 <sup>b</sup>	25.3 ± 26.6 <sup>b</sup>	15.3 ± 16.5 <sup>b</sup>
1,2,3,7,8,9 Hexa-CDD	8.12 ± 6.45 <sup>a</sup>	2.08 ± 2.03 <sup>b</sup>	3.55 ± 4.47 <sup>b</sup>	2.63 ± 4.37 <sup>b</sup>	1.22 ± 1.06 <sup>b</sup>
1,2,3,4,6,7,8-Hepta-CDD	84.9 ± 60.9 <sup>a</sup>	10.2 ± 8.0 <sup>b</sup>	20.0 ± 28.9 <sup>b,c</sup>	20.3 ± 52.9 <sup>b,c</sup>	4.83 ± 4.55 <sup>c</sup>
OCDD	477 ± 320 <sup>a</sup>	53.6 ± 51 <sup>b,c</sup>	152 ± 188 <sup>b</sup>	113 ± 196 <sup>b,c</sup>	42.9 ± 41.4 <sup>c</sup>
2,3,7,8-Tetra-CDF	0.94 ± 0.58 <sup>a</sup>	0.34 ± 0.40 <sup>b,c</sup>	0.40 ± 0.40 <sup>b</sup>	0.35 ± 0.27 <sup>b,c</sup>	0.18 ± 0.09 <sup>c</sup>
1,2,3,7,8-Penta-CDF	0.92 ± 0.47 <sup>a</sup>	0.50 ± 0.45 <sup>a</sup>	1.40 ± 1.33 <sup>a,b</sup>	0.31 ± 0.26 <sup>b</sup>	0.12 ± 0.07 <sup>c</sup>
2,3,4,7,8-Penta-CDF	21.1 ± 11.5 <sup>a</sup>	5.71 ± 5.95 <sup>b,c</sup>	9.94 ± 9.43 <sup>b</sup>	7.43 ± 6.93 <sup>b,c</sup>	4.43 ± 2.24 <sup>c</sup>
1,2,3,4,7,8-Hexa-CDF	7.0 ± 3.33 <sup>a</sup>	2.32 ± 1.75 <sup>b,c</sup>	3.29 ± 3.25 <sup>b</sup>	2.83 ± 2.97 <sup>b,c</sup>	1.31 ± 0.67 <sup>c</sup>
1,2,3,6,7,8-Hexa-CDF	8.22 ± 3.99 <sup>a</sup>	2.03 ± 1.86 <sup>b</sup>	3.31 ± 3.47 <sup>b</sup>	2.69 ± 2.86 <sup>b</sup>	1.46 ± 0.95 <sup>b</sup>
1,2,3,7,8,9-Hexa-CDF	0.62 ± 0.35 <sup>a</sup>	0.39 ± 0.41 <sup>b,c</sup>	0.06 ± 0.05 <sup>b</sup>	0.07 ± 0.05 <sup>b</sup>	0.45 ± 0.27 <sup>c</sup>
2,3,4,6,7,8-Hexa-CDF	2.2 ± 1.28 <sup>a</sup>	0.38 ± 0.44 <sup>a</sup>	0.88 ± 0.73 <sup>a</sup>	0.69 ± 1.15 <sup>a</sup>	0.06 ± 0.07 <sup>b</sup>
1,2,3,4,6,7,8-Hepta-CDF	4.81 ± 2.17 <sup>a</sup>	2.06 ± 0.65 <sup>a</sup>	2.99 ± 2.54 <sup>a,b</sup>	2.34 ± 3.13 <sup>b,c</sup>	1.07 ± 0.69 <sup>c</sup>
1,2,3,4,7,8,9-Hepta-CDF	0.39 ± 0.1 <sup>a</sup>	0.31 ± 0.52 <sup>a,b</sup>	0.10 ± 0.06 <sup>b</sup>	0.11 ± 0.05 <sup>b</sup>	0.13 ± 0.13 <sup>b</sup>
OCDF	0.72 ± 0.27 <sup>a</sup>	2.59 ± 1.27 <sup>b</sup>	0.49 ± 0.30 <sup>a,c</sup>	0.31 ± 0.15 <sup>a,c</sup>	0.39 ± 0.36 <sup>c</sup>
WHO-TEQ	40.1 ± 12.7 <sup>a</sup>	9.89 ± 9.27 <sup>b</sup>	14.6 ± 14.2 <sup>b</sup>	11.5 ± 11.1 <sup>b</sup>	6.63 ± 3.86 <sup>b</sup>

*Data given as mean ± standard deviation.*

<sup>a,b</sup>*Different superscripts indicate that the difference between years (columns) is statistically significant (p < 0.05).*

**Table 3. Concentration of PCDD/Fs (in pg WHO-TEQ/g fat) in 15 adipose tissue samples according to the sex of subjects.**

	Number of samples					Mean ± Standard deviation				
	1998	2002	2007	2013	2013	1998	2002	2007	2013	2019
<b>Total</b>	15	15	15	15	15	40.1 ± 12.7	9.9 ± 9.3	14.6 ± 14.2	11.5 ± 11.1	6.63 ± 3.86
<b>Sex</b>										
Men	10	11	11	9	14	35.8 ± 8.0	7.2 ± 3.5	11.2 ± 8.1	9.3 ± 17.9	7.44 ± 3.49
Women	5	4	4	6	1	48.7 ± 16.8	17.4 ± 16.1	23.8 ± 23.9	18.5 ± 5.9	1.89

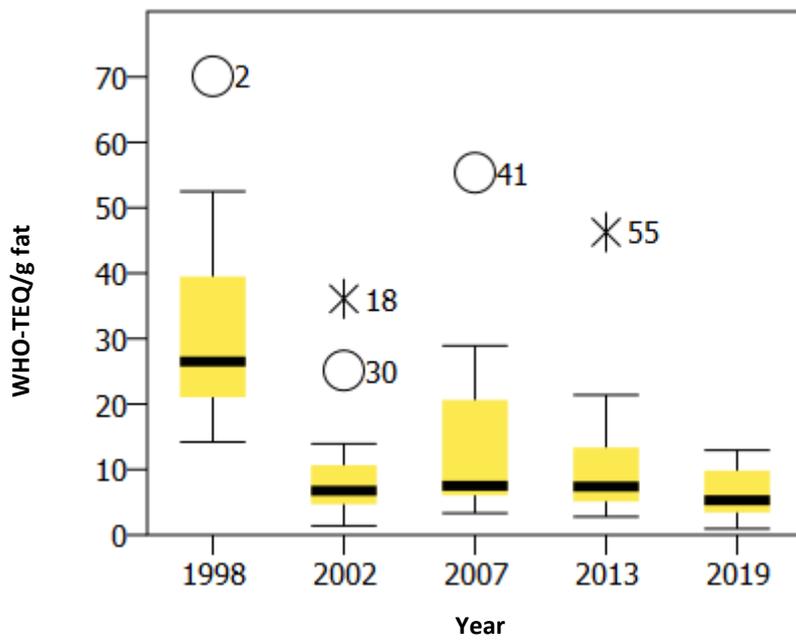


Figure 1. Boxplot of PCDD/Fs concentrations in adipose tissue samples obtained from autopsies of individuals who had been living near the HWI of Constantí.

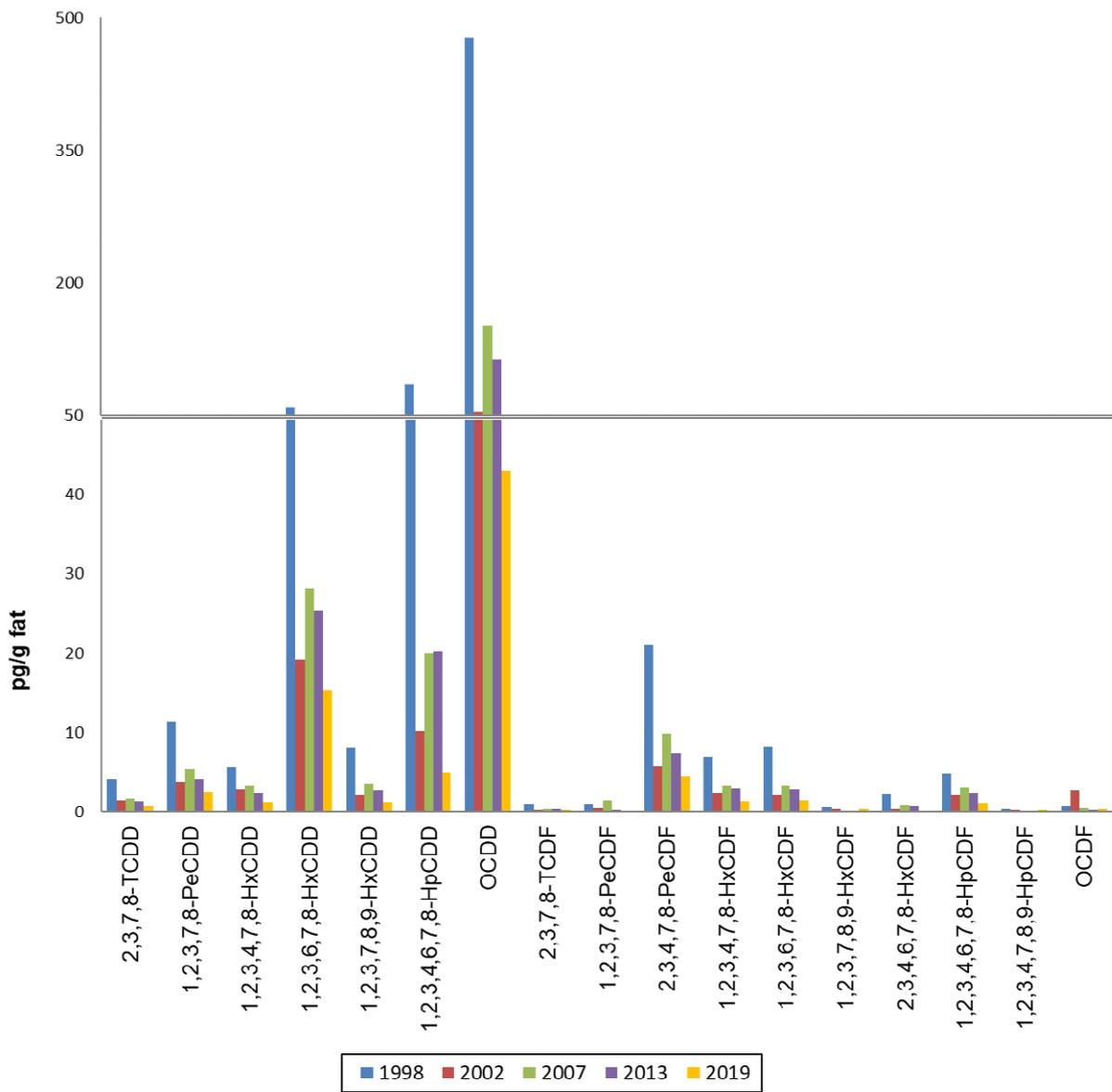


Figure 2. PCDD/Fs congener profiles in samples of adipose tissue from subjects who had been living near the HWI of Constantí.

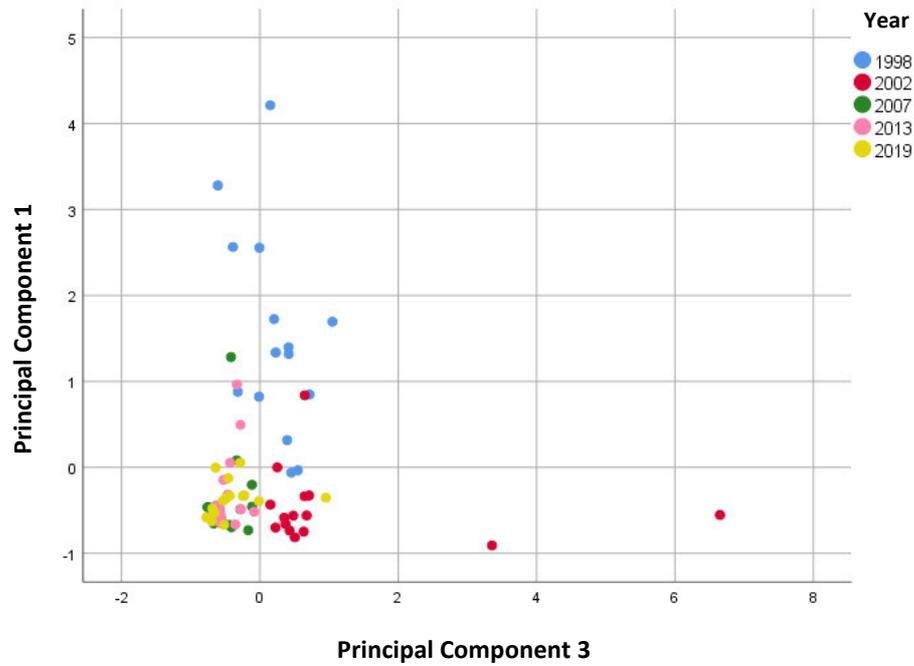
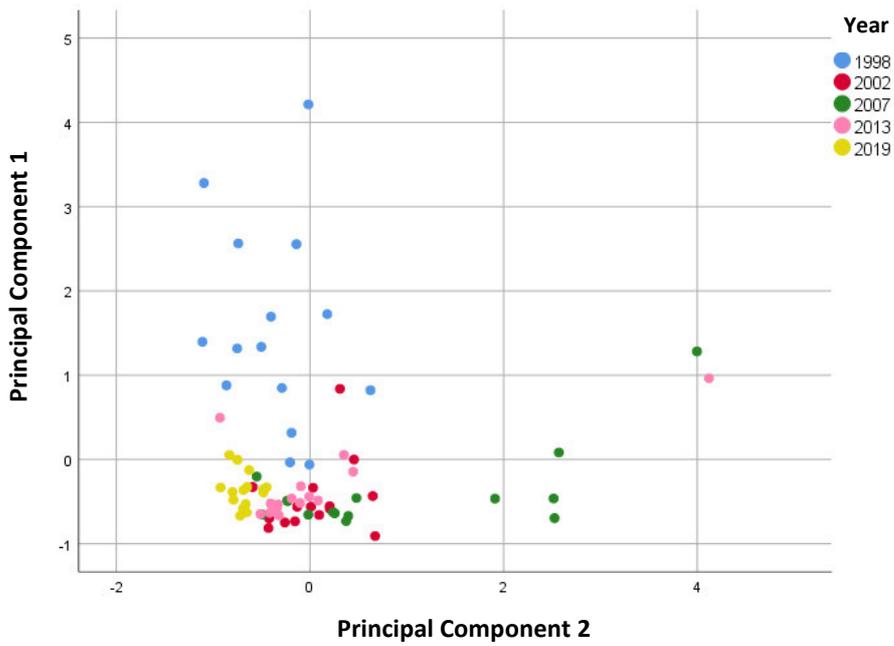


Figure 3. Principal Component Analysis applied to PCDD/Fs concentrations in adipose tissue.

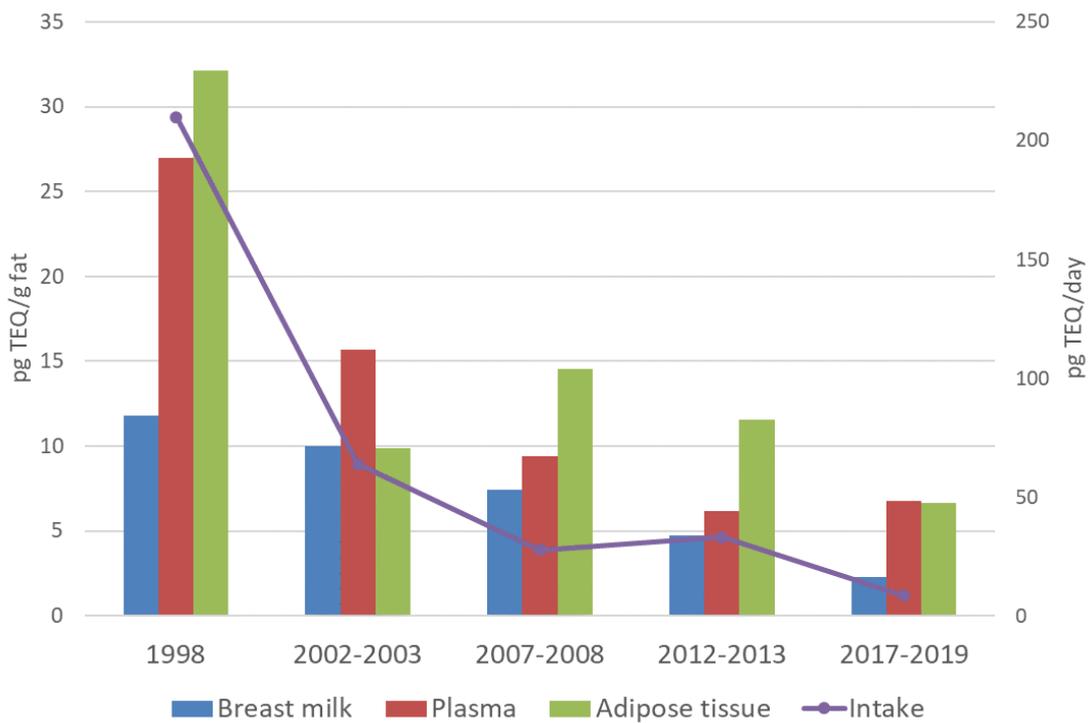


Figure 4. Temporal trend of PCDD/Fs concentrations in 3 biological monitors (in pg TEQ/g fat) and comparison with the dietary intake (in pg TEQ/day) by the adult population living near the HWI of Constantí. Sources: breast milk, Schuhmacher et al. (2019); plasma, Nadal et al. (2019); dietary intake, González et al. (2018).