

1 **Human health risks due to exposure to inorganic and**  
2 **organic chemicals from textiles: A review**

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31 **ABSTRACT**

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33 It is well known that a number of substances used in the textile industry can mean not  
34 only environmental, but also health problems. The scientific literature regarding  
35 potential adverse health effects of chemical substances in that industry is mainly related  
36 with human exposure during textile production. However, information about exposure  
37 of consumers is much more limited. Although most research on the health effects of  
38 chemicals in textiles concern allergic skin reactions, contact allergy is not the only  
39 potential human health problem. In this paper, we have reviewed the current scientific  
40 information regarding human exposure to chemicals through skin-contact clothes. The  
41 review has been focused mainly on those chemicals whose probabilities of being  
42 detected in clothes were rather higher. Thus, we have revised the presence of flame  
43 retardants, trace elements, aromatic amines, quinoline, bisphenols,  
44 benzothiazoles/benzotriazoles, phthalates, formaldehyde, and also metal nanoparticles.  
45 Human dermal exposure to potentially toxic chemicals through skin-contact  
46 textiles/clothes shows a non-negligible presence in some textiles, which might lead to  
47 potential systemic risks. Under specific circumstances of exposure, the presence of  
48 some chemicals might mean non-assumable cancer risks for the consumers.

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51 *Keywords:* Textile industry; clothes; organic substances; trace elements; dermal  
52 exposure; human health risks

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54 **1. Introduction**

55  
56 The textile industry is the responsible for taking raw materials such as cotton or  
57 wool and spinning it into yarn, which is used later to create a fabric. All the processes  
58 involved in converting the raw material into a finished product - developing, producing,  
59 manufacturing, and distributing textiles - are included in the textile industry, which  
60 utilizes a number of types of fabrics, with two major categories, natural and synthetic.  
61 Natural fabrics are those that occur naturally from animals and plants, while synthetic  
62 fabrics are being created in a laboratory and are man-made.

63 China is the largest textile producing and exporting country in the world. In 2016,  
64 China was the top ranked global textile exporter with a value of approximately 106  
65 billion US dollars, followed by the European Union (28 countries), with a value of 65  
66 billion dollars. India, USA and Turkey would be the following with values of 16, 13 and  
67 11 billion dollars, respectively (Statista, 2016). With respect to clothing, China and the  
68 European Union-28 were also the top two exporters in 2016, followed at great distance  
69 by Bangladesh, Vietnam, India, Hong Kong and Turkey. According to the World Trade  
70 Statistical Review (2017), the current dollar values of world textiles (SITC 65) and  
71 apparel (SITC 84) exports were in 2016, \$284 billion and \$443 billion, respectively.

72 As it happens with other industrial activities, textile industry has also environmental  
73 problems, being one of the oldest and most technologically complex of all industries. A  
74 number of substances used in the textile industry can mean not only environmental, but  
75 also health problems. Among the many chemicals whose presence has been detected in  
76 textile wastewater, dyes are among the most important pollutants (Brillas et al., 2015;  
77 Mohamed et al., 2016; Zare et al., 2018). Worldwide environmental problems  
78 associated with the textile industry are mainly associated with water pollution caused by  
79 the discharge of untreated effluent, as well as those due to the use of potentially toxic  
80 substances, especially during processing (Kan and Malik, 2013; Pattnaik et al., 2018).

81 The scientific literature regarding potential adverse health effects of chemical  
82 substances in the textile industry is mainly related with human exposure during textile  
83 production. Thus, examples of physical hazards associated with textile and clothing  
84 manufacture include fire risk, building construction, noise, temperature, humidity,  
85 unsafe machinery, dust and harmful chemicals. In contrast, the information about  
86 exposure of consumers is much more limited (KEM, 2014). There are numerous  
87 activities involved in the textile and clothing industry, going from the treatment of raw

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88 materials to finishing activities such as bleaching, printing, dyeing, impregnating,  
89 coating, plasticizing, etc. As result of these activities, the main chemical pollutants are  
90 dyes, which contain carcinogenic amines, metals, pentachlorophenol, chlorine  
91 bleaching, halogen carriers, free formaldehyde, biocides, fire retardants, and softeners  
92 (Brigden et al., 2012).

93 Most research about health effects of chemicals in textiles concern allergic skin  
94 reactions. Disperse dyes, used for staining synthetic fibers have been reported to be the  
95 most common causes of textile allergy, being contact allergy to disperse dyes a  
96 clinically relevant problem (Ryberg et al., 2006, 2009; Malinauskiene et al., 2013;  
97 Coman et al., 2014). However, contact allergy is not the only human health problem.

98 It is well known that humans are exposed to toxicants mainly through the diet  
99 (food and drinking water) and breathing (air pollution). However, for some chemicals  
100 dermal exposure should not be minimized. In relation to dermal exposure, although  
101 most chemicals added during the processes of manufacturing clothes are rinsed out,  
102 residual concentrations of some substances can remain and can be released during the  
103 use by the consumers (Luongo et al., 2014).

104 Based on the above, the main goal of this paper has been to review the current  
105 information regarding human exposure to chemicals through skin-contact clothes. The  
106 present paper is focused only on chemicals present in textiles in contact with the skin.  
107 For this reason, chemicals such as perfluorinated compounds, which are associated with  
108 specific clothes (i.e., waterproof) were not include in this review.

## 109 110 **2. Flame retardants**

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112 In a review on the presence of additives in fibers and fabrics, which was  
113 published in the seventies of the last century, Barker (1975) assumed that all fibers and  
114 fabrics contained measurable amounts of contaminants and additives. It was concluded  
115 that while the levels of contamination of the fibers was –in general terms- quite low,  
116 significant concentrations of added chemicals were present at that time in fabrics treated  
117 for flame resistance, or for oil and water repellency, for example. Among the flame  
118 retardants traditionally used in the textile industry, polybrominated diphenyl ethers  
119 (PBDEs) have played an important role. PBDEs are a class of brominated flame  
120 retardants (BFRs) that are added to plastics, polyurethane foam, textiles, and electronic  
121 equipments in order to protect people from fires by reducing the flammability of these

122 potentially combustible materials (Lorber, 2008). In recent years, PBDEs have become  
123 widespread environmental pollutants (Malarvannan et al., 2015; Guiguerno and Fernie,  
124 2017), which have been incorporated in tissues of the general population (Schuhmacher  
125 et al., 2009, 2013; Ma et al., 2017; He et al., 2018). A number of studies have shown  
126 that, as for other persistent organic pollutants, dietary intake is the main route of human  
127 exposure to PBDEs (Lorber, 2008; Linares et al., 2015; Domingo, 2012; Domingo et al.,  
128 2008; Perelló et al., 2009). However, information on human exposure to PBDEs  
129 through the skin is very limited. Chen et al. (2009) measured the levels of various BFRs  
130 –including PBDEs- in children’s toys purchased from South China. Higher exposures  
131 predominantly contributed through the mouthing pathway, being inhalation, dermal  
132 contact and oral ingestion less relevant routes of exposure associated with toys.

133         With respect to human exposure to BFRs through clothes, it is important to  
134 remark that clothing covers approximately 85% of the human skin, being able of acting  
135 as a barrier for environmental pollutants. However, clothing can be, in turn, a potential  
136 source of exposure to certain chemicals. Recently, Saini et al. (2016) investigated the  
137 role of clothing as a sorbent of indoor semivolatile organic compounds (SVOCs) and as  
138 source to outdoors through laundering. Phthalates, BFRs and organophosphate esters  
139 (OPEs) were measured. It was demonstrated that clothing acted as an efficient conveyer  
140 of soluble SVOCs from indoors to outdoors through accumulation from air, and then  
141 release during laundering. Clothes drying could also contribute to the release of  
142 chemicals emitted by electric dryers. These findings have implications for potential  
143 dermal exposure. In a subsequent study conducted by the same research group (Saini et  
144 al., 2017), the accumulation of phthalates and BFRs to cotton and rayon was assessed  
145 by deploying these fabrics indoors in 20 homes and 5 offices for 28 days and measuring  
146 uptake over 56 days. The results confirmed the accumulation of both gas-and particle-  
147 phase chemicals to cotton. It was concluded that this large sorptive capacity should have  
148 implications for fabrics as a chemical sink for SVOCs indoors, as well as for human  
149 exposure.

150         In order to better characterize the concentrations of PBDEs in lint, Schechter et al.  
151 (2009) determined the levels of these chemicals in household drier lint in the USA and  
152 Germany. It was found that despite being washed prior to drying, clothes might be a  
153 source of PBDE contamination of dryer lint, and they could serve as an indicator of  
154 indoor exposure to these pollutants. The authors suggested that the source of PBDEs in

155 lint would be mainly derived from dryer electrical components and also dust deposition  
156 onto clothing.

157         Due to the increased regulatory interest in the restrictions of PBDEs and other  
158 flame retardants such as hexabromocyclodecane (HBCD), in recent years the use of  
159 alternative flame retardants has been unavoidable. In Japan, for treatment of textiles,  
160 HBCD and DecaBDE have been the most frequently used BFRs. As they have been  
161 incorporated as additives –without being covalently linked to the polymers to which  
162 these compounds are applied- they tend to migrate, becoming environmental pollutants,  
163 which can be then subjected to photolytic transformations. Based on this, Kajiwara et al.  
164 (2013) investigated the photolytic debromination and isomerization of the major  
165 components of HBCD and DecaBDE in flame-retarded curtains under natural sunlight.  
166 The concentrations of polybrominated dibenzofurans (PBDFs) in the textiles, formed as  
167 products of photodecomposition of DecaBDE, were 4–5 orders of magnitude lower than  
168 the levels of PBDEs. The authors remarked that PBDFs were formed as a result of  
169 sunlight exposure during normal use of textile products treated with DecaBDE. These  
170 results were in the same line than those of a previous study of the same research group,  
171 showing that fibers from the flame-retarded textiles could be an important component of  
172 indoor dust, and consequent of human exposure (Kajiwara et al., 2009). With respect  
173 specifically to the levels of PBDEs in various textiles treated with BFRs, Shin and Baek  
174 (2012) detected BDE-28, -66, -100, -119, -153, -197, -206, and -209 in all textile  
175 samples analyzed by HRGC/HRMS. In contrast, the congeners BDE-3, -7, -47, -49, -71,  
176 -99, -126, and -156 were not detected in any sample. The highest concentration  
177 corresponded to BDE-209. As a consequence of the wide use of PBDEs as flame  
178 retardants in a number of commercial items, including textiles and clothes, and their  
179 transference to the environment, the presence of these compounds in outdoor and indoor  
180 environments is evident (Lim et al., 2014). Recently, Ionas et al. (2015) determined the  
181 concentrations of PBDEs and organophosphate flame retardants in textile home  
182 furnishings, such as carpets and curtains from stores in Belgium. The levels of PBDEs  
183 were typically too low to impart flame retardancy. Only high levels of BDE-209 (11–  
184 18% by weight) were found.

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### 186         **3. Trace Elements**

187 The main toxic pollutants in textiles are dyes, which contain various chemicals  
188 with a well established toxicity. Among these, a number of toxic trace elements can be  
189 found in textile materials because they are frequently present in various textile  
190 processes. In addition, raw textile materials can also contain several trace elements.  
191 Metals -in textile products and clothing- are used for many purposes, such as metal  
192 complex dye (cobalt, copper, chromium, lead), pigments, mordant (chromium), catalyst  
193 in synthetic fabrics manufacture (antimony oxide), synergists of flame retardants  
194 ( $Sb_2O_3$ ), antimicrobials (nanoparticles of silver, titanium oxide and zinc oxide), as well  
195 as like water repellents, and odor-preventive agents (Derden and Huybrechts, 2013;  
196 Muenhor et al., 2010; Simoncic and Tomsic, 2010; Stefaniak et al., 2014; Wöhrle et al.,  
197 2012). The relationship between trace elements and textiles may mean an important  
198 environmental problem for the textile industry, while the presence of certain toxic trace  
199 elements in clothing may also mean health hazards for the consumers. An exhaustive  
200 search of the scientific literature shows that human exposure to metals rarely produce  
201 morbidity, and very rarely mortality. However, a continued exposure to low levels of  
202 toxic elements such as As, Cd, Hg and Pb –among others- is related with a number of  
203 adverse effects (García-Esquinas et al., 2015; Jaishankar et al., 2014; Rodríguez-  
204 Barranco et al., 2014; Roy et al., 2011). In addition, various metals such as Cu, Co, Fe,  
205 Mn, Mo, or Zn, which are essential for humans, can be also dangerous at high exposure  
206 levels (Domingo, 1994, Lucas et al., 2015).

207 Based on the potential health risks derived from exposure to metals, the  
208 concentrations of these elements have been quantified in various textile materials.  
209 Tuzen et al. (2008) determined the concentrations of 6 trace metals (Cu, Cd, Zn, Mn, Fe  
210 and Ni) in various textile samples collected in Turkey. The levels of these metals were  
211 in the range of 0.10-0.25  $\mu\text{g/g}$  for Cd and 3.55-34.3  $\mu\text{g/g}$  for Fe, metals showing the  
212 lowest and highest values, respectively. Copper and Cd concentrations in the analyzed  
213 samples were higher than the limit values given by Oeko-Tex. In turn, Rezic and  
214 Steffan (2007) determined the levels of 17 trace elements in 16 textile samples of  
215 different origin. Results in the sweat extracts (minimum-maximum in  $\mu\text{g/mL}$ ) were the  
216 following: Al 0.11-1.58, Cd 0.02-0.05, Cr 0.01-0.32, Cu 0.05-1.95, Mn 0.01-2.17 and  
217 Ni 0.05-0.10. Concentrations of other elements were below the respective detection  
218 limits. The concentrations of some trace elements were above the limits suggested by  
219 different ecological standards. Thus, Zn and Cd were found in cotton and polyester

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220 samples, Cr was detected in flax, silk and polyester samples, Cu was found in silk  
221 samples, and As was detected in silk and polyester samples. The same research group  
222 (Zeiner et al., 2007) also tested several analytical procedures for determining heavy  
223 metals in the textile industry, concluding that the method for an exact quantification of  
224 these elements should be selected depending on the analytical task. In a subsequent  
225 investigation of the same authors (Rezic et al., 2011), the concentrations of 28 trace  
226 elements (Al, As, Ba, Be, Bi, Ca, Cd, Co, Cr, Cu, Fe, K, Hg, Mg, Mn, Mo, Na, Ni, Pb,  
227 Sc, Si, Se, Sn, Sm, Sr, Tl, V and Zn) were determined in raw textile materials (cotton,  
228 flax, hemp and wool) by inductively coupled plasma optical emission spectrometry  
229 (ICP-OES) after microwave digestion of the samples. The levels of these elements  
230 ranged between <LOD for various elements, and 1170.2 mg/g for K in cotton. In flax,  
231 they ranged between <LOD for various elements and 86.6 mg/g for Mg; in hemp,  
232 between <LOD for various elements and 540 mg/g for Ca, and finally, between <LOD  
233 for various elements and 660 mg/g for Ca in wool. On the other hand, Matoso and  
234 Cadore (2012) measured the concentrations of Sb, As, Pb, Cd, Cr, Co, Cu, Ni and Hg in  
235 polyamide raw materials (pellets) and textiles used in sport T-shirts. The highest levels  
236 of trace elements were found for Cr in black fabrics, but the extractable content –using  
237 acid solution- was lower than the limits suggested by Oeko-Tex Standard 100:2017  
238 Standards (OEKO-Tex, 2018).

239         It is well known that the main routes through which trace elements can reach the  
240 human body are ingestion and inhalation (Giné-Bordonaba et al., 2011). However,  
241 human exposure to metals through skin contact could also represent a non-negligible  
242 pathway for some elements and certain conditions of exposure. Based on this  
243 assumption, we recently conducted in our laboratory various investigations focused on  
244 determining the levels of a number of trace elements in clothes, as well as to assess the  
245 potential health risks for the consumers. In a first study (Rovira et al., 2015), the  
246 concentrations of Al, As, B, Ba, Be, Bi, Cd, Co, Cr, Cu, Fe, Hg, Mg, Mn, Mo, Ni, Pb,  
247 Sb, Sc, Se, Sm, Sn, Sr, Tl, V and Zn were determined in various skin-contact clothes.  
248 The analyzed samples (T-shirts, blouses, underwear, baby pajamas and bodies) were  
249 made of cotton, polyamide, polyester, spandex, and viscose, being classified according  
250 to color, brand, and eco-labeled categories. High levels of Cr in polyamide dark clothes  
251 (605 mg/kg), Sb in polyester clothes (141 mg/kg), as well as Cu in some green cotton  
252 fabrics (around 280 mg/kg) were found. Interestingly, lower concentrations of Al and Sr

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253 were found in “eco” clothes, while no significant differences were noted in branded and  
254 unbranded clothing pieces. Moreover, Al and Sc levels were higher in clothes made in  
255 EU countries than in those made outside the EU. Although the non-carcinogenic and  
256 carcinogenic risks of dermal contact exposure for the consumers were –for most trace  
257 elements- below safe ( $HQ < 1$ ) and acceptable ( $< 10^{-6}$ ) limits, respectively, for Sb, the  
258 non-carcinogenic risk was above 10% of the safety limit ( $HQ > 0.1$ ) for dermal contact  
259 with clothes. In a second study (Rovira et al., 2017a), we assessed the dermal contact  
260 exposure to 28 trace elements and the derived human health risks by analyzing the  
261 levels of these elements in 37 skin-contact clothes (T-shirts, blouses, socks, baby  
262 pajamas and bodies). To establish a more realistic exposure assessment, migration  
263 experiments were also conducted by determining the concentration of the same 28  
264 elements in artificial sweat. Dermal exposure to trace elements for adult males and  
265 females, as well as for <1 year-old children, were calculated and the associated health  
266 risks were assessed. High concentrations of Zn (186–5749 mg/kg) were detected in zinc  
267 pyrithione labeled T-shirts, while high levels of Sb and Cr were found in polyester and  
268 black polyamide fabrics, respectively. In turn, an environmental scanning electron  
269 microscope (ESEM) confirmed the presence of Ag and Ti particles and aggregates in  
270 various clothes. All samples analyzed in that study fulfilled the parameters of the Oeko-  
271 Tex standard. However, four polyester samples exceeded the extractable Sb limit of  
272 TOX-Proof standard, which is set at 1.0 mg/kg. With respect to health risks, for  
273 polyester clothes, the mean HQs for Sb were 0.44, 0.40 and 0.13, for adult males, adult  
274 females, and children <1 year-old, respectively, with one polyester T-shirt reaching a  
275 value of 1.2. For the remaining analyzed trace elements and samples of clothes, non-  
276 carcinogenic and carcinogenic risks were considered as safe ( $HQ < 1$ ) and acceptable  
277 ( $< 10^{-5}$ ), respectively. We also detected ZnPT and Ag nanoparticles in some clothes.  
278 These substances with biocide activity could affect the natural skin microflora and,  
279 consequently, we concluded that they could lead to adverse effects on the human skin.

280 To the best of our knowledge, there was not any available scientific information  
281 on the presence of trace elements in commercially available home textiles, as well as on  
282 the potential adverse health effects of a continued exposure during their use. Therefore,  
283 the main goal of our third study (Rovira et al., 2017b) was focused on these objectives.  
284 For it, the levels of 28 trace elements (Ag, Al, As, B, Ba, Be, Bi, Cd, Co, Cr, Cu, Fe,  
285 Hg, Mg, Mn, Mo, Ni, Pb, Sb, Sc, Se, Sm, Sn, Sr, Ti, Tl, V and Zn) were determined in

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286 78 samples of home textiles (towels, bedclothes and pajamas). Based on the fabric  
287 material, samples were classified into three categories: 100% cotton, cotton + synthetic  
288 and synthetic. Arsenic, Be, Cd, Sc, Se, Sm and Tl showed concentrations below their  
289 respective LODs in all samples. In turn, Hg, Mo and V were detected only in 2–4 items.  
290 The highest mean concentrations corresponded to Mg (142 mg/kg), Cu (32.8 mg/kg), Sb  
291 (26.9 mg/kg), Al (14.7 mg/kg), Fe (12.9 mg/kg) and Ti (10.9 mg/kg). However, after an  
292 individual assessment of the samples, the highest concentrations corresponded to Cu in  
293 a black (100% cotton) sample and to Mn in a brown (50% lyocell–50% cotton) sample,  
294 with concentrations of 1065 and 889 mg/kg, respectively. In agreement with the results  
295 of our previous studies (Rovira et al., 2015, 2017a), polyester items contained high  
296 levels of Sb, while Ti concentrations were also increased in synthetic fiber samples. On  
297 the other hand, textile color was a key issue because of the high levels of Cr, especially  
298 in polyamide black clothes, as well as those of Cu in colored (blue, green, red, and  
299 brown) clothes made of cotton. Regarding human health risks, the maximum HQ for  
300 almost all trace elements was well below 0.01. The only exception was Sb, whose HQs  
301 for dermal exposure due to the use of bedclothes/pajamas and towels were 0.4 and >1,  
302 respectively. Comparing with other daily activities, towels use, by towel to-hand-to-  
303 mouth effect, was the most relevant action leading to dermal exposure for most trace  
304 elements. In general terms, cancer risks did not exceed threshold levels, excepting  
305 Cr(VI), whose risk was above  $10^{-5}$ .

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306 In a similar line to that of our recent studies (Rovira et al., 2015, 2017a,b),  
307 recently Nguyen and Saleh (2017) determined the levels of Ag, Al, As, Ba, Be, Bi, Ca,  
308 Cd, Co, Cr, Cu, Fe, Hg, K, Li, Mg, Mn, Na, Ni, Pb, Sb, Se, Sr, Ti, V and Zn in 120  
309 samples (63 cottons, 44 nylons and 13 polyesters) of different brands and colors of  
310 women undergarments manufactured in 14 countries. It was found that cotton samples  
311 were rich in Al, Fe and Zn, while nylon undergarments contained high levels of Cr, Cu  
312 and Al. In turn, polyester fabrics contained higher concentrations of Ni and Fe  
313 compared to cotton or nylon. China, Egypt and India were the countries for which the  
314 highest levels of trace elements were found in all fabrics. Chromium exceeded the  
315 Oeko-Tex limits in 35% of the samples, while Pb and Ni did it in 14% and 5% of the  
316 analyzed items, respectively. The authors recommended that the consumers should be  
317 aware of the potential health risks of the contents of these metals in their clothing.

#### 4. Aromatic amines

Azo dyes represent the most important class of textile dyes. Although various aromatic amines (AAs) have been used as intermediates in the synthesis of azo dyes (Freeman, 2013), it is well known that AAs may have carcinogenic and genotoxic properties together with other toxicological effects, as well as allergenic potential (Brüschweiler et al., 2014; Brüschweiler and Merlot, 2017; Platzek, 2010). Moreover, dermal, systemic and bacterial biotransformation of azo dyes, can –in turn- release aromatic AAs (Stingley et al., 2010). According to this, the AAs that may release one of the 22 known carcinogenic AAs are currently banned from clothing textiles in the EU. Nevertheless, there are still an important number of gaps on the knowledge of the potential toxicity of non-regulated AAs.

To investigate the occurrence of azo dye cleavage products, Brüschweiler et al. (2014) collected 153 samples of colored or black clothing (T-shirts, underwear, sport clothes, scarves and clothes for children, which were purchased in the Canton of Bern, Switzerland. The most important criteria for sampling were direct contact with the skin, possible contact with sweat, as well as contact with sensitive skin. The 22 high priority non-regulated AAs of toxicological concern were analyzed. Eight of these 22 AAs were found in 17% of the samples, while in 9% of the samples, one or more of the AAs of concern were detected at concentrations >30 mg/kg. Finally, in 8% of the samples these AAs were found between 5 and 30 mg/kg, being the highest measured concentration 622 mg/kg of textile. The authors concluded that there was a major toxicity data gap for many of the AAs that can be cleaved from the 470 textile azo dyes. This means that the regulatory gap must be filled in a systematic and consistent manner. To fill this gap, in a recent study Brüschweiler and Merlot (2017) investigated the mutagenicity of 397 non-regulated AAs, which could be potentially released from the 470 known textile azo dyes. Thirty-six mutagenic AAs, via publicly available databases, were identified. In addition, 40 different AAs, that were found to be mutagenic (primarily in the Ames test), and are potentially released as cleavage products from approximately 180 parent azo dyes used in clothing textiles, were also identified. Based on these results, the authors concluded that not only exposure to single AAs, but also combined exposure to different mutagenic AAs in textiles must be considered for a complete exposure and health risk assessment, taking into account that mutagenic properties of AAs can mean a much higher concern than previously expected. Also in the same line, Nguyen and

353 Saleh (2016) investigated the levels of azo dyes and AAs in women under garment.  
354 Samples (120) of women underwear of different colors, fabric structures, origin,  
355 geographical locations of manufacture and brand names, were evaluated for their  
356 potential release of AAs to the skin. Low level mixtures of AAs were detected in 74  
357 samples, but 18 samples had greater amounts of AAs than that recommended by the EU  
358 and China. The authors remarked the importance of analyzing AAs in garments.

## 360 **5. Quinoline, bisphenols, benzothiazoles and benzotriazoles**

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362 Quinoline and derivatives, bisphenols (BPAs), benzothiazoles (BTHs) and  
363 benzotriazoles (BTRs) are chemicals used in a wide range of applications, including  
364 clothing textile articles. Quinoline, a heterocyclic aromatic organic compound, and its  
365 derivatives are extensively used in the textile industry for the manufacture of dyes.  
366 Some of them are skin irritants and/or also probable human carcinogens. The presence  
367 of these substances has been reported in textile materials (Yang et al., 2013; Luongo et  
368 al., 2014, 2016a). The concentrations of quinoline and 10 quinoline derivatives were  
369 measured in 31 clothing samples of different types: from T-shirts and jeans to dresses.  
370 They represented different colors, materials, brands, countries of manufacture, and  
371 prices (Luongo et al., 2014). Quinolines were detected in 29 of the 31 analyzed samples,  
372 with quinoline contributing up to around 50% of the total amount of quinolines. The  
373 highest levels was 1.9 mg in a single garment, a result that cannot be minimized taking  
374 into account that the skin is exposed to a large surface area of clothing, as well as the  
375 potential health risks of these compounds. In a subsequent study conducted by the same  
376 research group (Luongo et al., 2016a), again 10 quinoline compounds were determined  
377 in textiles made of cotton, polyamide (> 70%) and polyester. Quinoline was detected in  
378 all samples at levels between 0.06 and 6.2 µg/g. As quinoline and isoquinoline are  
379 classified as carcinogens, the authors pointed out the importance of collecting data on  
380 this potential source to daily human exposure through the skin.

381 With respect to BTHs and BTRs, Avagyan and co-workers (2015) analyzed the  
382 concentrations of 11 derivatives of these families of compounds in 26 clothing samples  
383 –including items for babies, toddlers and children- of various textile materials, colors  
384 and manufactured in 14 different countries. Eight of the 11 analyzed compounds could  
385 be detected in the clothing samples, which demonstrated that clothes could be a route of  
386 human exposure to BTHs and BTRs. These substances are genotoxic, can act as

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387 endocrine disruptors, and may be dermal sensitizers -exhibiting allergenic and irritating  
388 properties- among other toxic effects (Ginsberg et al., 2011; Oda et al., 2008). The  
389 sample with the highest concentration of BTH contained 8.3 mg of this chemical, while  
390 it could be also detected (22 µg/g) in a baby body made from "organic cotton" equipped  
391 with the "Nordic Ecolabel" ("Svanenmärkt"). In general terms, the levels of BTHs were  
392 much higher than those for BTRs. Another study performed in the same laboratory  
393 (Luongo et al., 2016b), showed that the average emissions to household wastewater of  
394 benzothiazoles and quinolines during one washing (5 kg of clothes made from polyester  
395 materials) were 0.5 and 0.24 g, respectively. Taking into account that notable amounts  
396 of these compounds remained still in the clothes even after 10 times of washing, these  
397 results corroborated that clothes are a potential source of human exposure to BTHs,  
398 BTRs and quinolines.

399         Recently, Liu et al. (2017) determined the occurrence of benzothiazole  
400 benzotriazole and 7 common derivatives in a total of 79 textile samples, including raw  
401 textiles (fabrics) and infant clothing (blankets, diapers and clothing). The concentrations  
402 of BTHs and BTRs were examined on the basis of fabric type (e.g., cotton, polyester,  
403 and nylon), countries of origin, and colors. BTH was the most frequently detected  
404 compound. However, the concentrations of BTR were elevated in certain textiles, with  
405 the highest concentration of BTR (14,000 ng/g) found in a printed graphic of infant's  
406 bodysuit. The overall mean levels of BTR in the analyzed textiles were higher than  
407 those of BTH. Dermal exposure of these chemicals in infants was also evaluated. It was  
408 found to be high from the use of socks (244 to 395 pg/kg·bw/day), being the exposure  
409 doses of BTHs and BTRs from textiles as high as 3740 pg/kg·bw/day from a themed  
410 graphic imprint on the chest portion of a bodysuit. The same investigators also analyzed  
411 77 textiles and infant clothing pieces to determine the occurrence of bisphenols,  
412 including bisphenol A (BPA) and bisphenol S (BPS), benzophenones, bisphenol A  
413 diglycidyl ethers (BADGEs) and novolac glycidyl ethers (NOGEs) (Xue et al., 2017).  
414 Various samples of fabric types (e.g., cotton, polyester, nylon), colors, and countries of  
415 origin, were collected. They included raw textiles, cloth diapers, blankets, and clothing  
416 marketed for infants aged < 1 year. The results showed that BPA and BPS occurred in  
417 82% and 53% of the textile samples, with mean concentrations of 366 and 15 ng/g,  
418 respectively. In turn, benzophenone-3 (BP3) was found in 70% of the samples at a mean  
419 concentration of 11.3 ng/g. Finally, among the 11 BADGEs and NOGEs analyzed,  
420 BFDGE was the predominant compound, with a mean concentration of 13.6 ng/g.

1 421 Dermal BPA exposure doses from textiles ranged from 201 pg/kg bw/day for 6–12  
2 422 months old infant to 248 pg/kg bw/day for newborns, while BP3 exposure doses ranged  
3 423 between 6.17 and 7.62 pg/kg bw/day. Dermal BPA exposure doses from some textiles  
4 424 were as high as 7280 pg/kg bw/day for newborns. Among the analyzed clothes, most  
5 425 BPA exposure in infants corresponded to socks. The authors of these studies concluded  
6 426 remarking the need for further investigations on the sources and levels of exposure of  
7 427 the chemicals that are present in textiles (Liu et al., 2017; Xue et al., 2017).  
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## 14 429 **6. Phthalates**

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17 431 Phthalates, a wide range of chemical compounds, are mainly used as plasticizers  
18 432 in plastics (especially PVC) to increase softness and flexibility. PVC prints are one of  
19 433 the uses of phthalates in textile industry; plasticized PVC that could lead to long periods  
20 434 of direct skin contact. It should be taken into account that final consumers of textiles  
21 435 with printed PVC are children, who are the most vulnerable group to these endocrine  
22 436 disrupting compounds due to their developmental status (Martinez et al., 2018). A clear  
23 437 example of this is the study conducted by Pedersen and Hartmann (2004), who analyzed  
24 438 phthalate content in children clothes. Phthalates were detected in all the garments tested  
25 439 (19 samples), with levels (sum of all phthalates) between 1.4 mg/kg and 200,000 mg/kg  
26 440 (around 20% of the weight of the sample). The phthalates found at higher amounts were  
27 441 DEHP, DINP, and DHP ((Pedersen and Hartmann, 2004). In turn, Negev et al. (2018)  
28 442 found phthalate (mainly DEHP and DINP) levels in nylon sheets, crib mattress and  
29 443 diaper-changing mats above 0.1% by mass, the standard set by EU Commission  
30 444 Regulation (EC No 552/2009). Similar results (levels >0.1%) were obtained in a study  
31 445 conducted by Li et al. (2015), suggesting that textiles should be monitored for these  
32 446 compounds.  
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## 49 448 **7. Formaldehyde**

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52 450 The International Agency for Research on Cancer (IARC) classifies  
53 451 formaldehyde as human carcinogenic to humans (Group 1) (IARC, 2012). For a long  
54 452 time, textiles and clothes have been treated with formaldehyde releasing compounds and  
55 453 resins in order to improve the anti-creasing properties (Aldag et al., 2017). In 1950s and  
56 454 1960s, the durable press chemical finishing were based on urea–formaldehyde resin and  
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455 melamine/formaldehyde resin, which released considerable amounts of formaldehyde  
456 to clothes (5000 to 12000 ppm: 0.5-1.2%) (de Groot et al., 2010a). These high  
457 formaldehyde concentrations led to the report of a number of cases of contact  
458 dermatitis (de Groot et al., 2010a,b). Nowadays, durable press finishes are based on  
459 modified dimethylol dihydroxyethyleneurea, which releases less formaldehyde (de  
460 Groot et al., 2010b). Various studies have determined formaldehyde levels in clothes.  
461 Novick et al. (2013) analyzed 20 cloth items detecting formaldehyde only in 3 of them.  
462 However, levels of 2 out these 3 detected items (3172 and 1391 ppm) were 40 fold  
463 higher than the concentrations established by international textile regulations.  
464 According to Novick et al. (2013), washing and drying procedures reduced  
465 formaldehyde concentrations between 26% and 72%. The US Government  
466 Accountability Office (USGAO, 2010) performed a study with 180 textile items. Thirty-  
467 five out 180 free formaldehyde samples were detected, while 10 of them exceeded the  
468 regulatory standard (75 ppm in non-baby direct contact with skin clothes), with values  
469 between 75.4 and 206.1 ppm. In turn, Piccini et al. (2007) analyzed 221 samples of  
470 clothes and linen. It was reported that in 89% of them formaldehyde levels were below  
471 30 ppm, while in 97% of them the concentrations were below 75 ppm. Only three items  
472 exceeded the level of 100 ppm, with a maximum value of 163 ppm. Differences  
473 according to the manufacture country and the shops where the samples had been  
474 purchased were found (Piccini et al., 2007).

## 475 476 **8. Nanoparticles**

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478 In recent years, the use of metal nanoparticles has meant a tremendous boom for  
479 many industrial sectors. The ability of manufacturing novel nanomaterials have led to  
480 an increased production and use of engineered nanoparticle (ENPs) and engineered  
481 nanomaterials (ENMs). Among these, ENPs of metals are increasingly used in the  
482 textile industry (Som et al., 2011; Montazer et al., 2014; Yetisen et al., 2016) The  
483 antimicrobial activity of silver-ENP and the UV-absorption of titania-ENP are good  
484 examples of this use (Windler et al., 2013; von Goetz et al., 2013; Lombi et al., 2014).  
485 However, mobilization and migration of ENPs from the textile into human sweat can  
486 result in dermal exposure to these chemicals and their aggregates and agglomerates  
487 (NOAA) (von Goetz et al., 2013). Silver is one of the elements mainly used in metal  
488 nanoparticles. The use of nanoscale Ag in textiles is one the most often mentioned uses

1 489 of nano-Ag. However, the form of Ag present in the textiles remains largely unknown  
2 490 as product labeling is insufficient (Lombi et al., 2014). In relation to this, it has been  
3 491 reported that for Ag-NOAA, its potential dermal exposure from textiles is comparable  
4 492 in magnitude to the major source of Ag-ENP, which are dietary supplements (von Goetz  
5 493 et al., 2013).

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9 494 Given the increasing importance in the textile industry of metal nanoparticles in  
10 495 general, and silver nanoparticles in particular, the concern regarding their environmental  
11 496 and human health risks has also increased in parallel. Because of the great importance  
12 497 of these issues, an exhaustive revision of this topic has not been included in the current  
13 498 paper. We feel that this subject is worthy of an additional and exhaustive specific  
14 499 review. However, for those interested in the topic, information is available in some  
15 500 recent papers (León-Silva et al., 2016; Voelker et al., 2015; Tang et al., 2015; Som et  
16 501 al., 2011; McGillicuddy et al., 2017).

## 23 502 24 25 503 **9. Conclusions**

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29 505 Most chemicals that have been added during the processes of converting fabrics  
30 506 into textiles are rinsed out. However, this does not prevent that residual levels may  
31 507 remain in the finished products, which -in turn- can be released during the use by the  
32 508 consumers. Although there are important control measures for many hazardous  
33 509 compounds employed in the EU, the continuing relocation of textile production to  
34 510 countries with fewer environmental restrictions and work standards, the complex raw  
35 511 material supply chains, as well as the large numbers of operators involved in the  
36 512 different production steps, makes that a strict control on the presence of some toxic  
37 513 chemicals in textiles in general, and in clothes in particular, is indeed very difficult. On  
38 514 the other hand, the rapid changes in fashion trends also lead to fluctuations in the types  
39 515 of prints, dyes, and other kinds of chemicals that are being used during the production  
40 516 processes (Fransson and Molander, 2013; Luongo et al., 2014). In 2012, Brigden and  
41 517 co-workers published an interesting report on the presence and in some cases  
42 518 concentrations of hazardous chemicals (nonyl phenol ethoxylates, carcinogenic amines,  
43 519 and phthalates) in branded textile products on sale in 27 countries. The results suggested  
44 520 the necessity of develop robust policies to force the elimination of such chemicals from  
45 521 manufacturing processes of textiles, and consequently, in the finished products (Brigden  
46 522 et al., 2012).

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523 In the present paper, we have reviewed the presence of various potentially toxic  
524 chemicals in textiles. The review has been mainly focused on those chemicals whose  
525 probabilities of being detected in clothes were rather higher. Among these, flame  
526 retardants have been included. Flame retardants are incorporated into potentially  
527 flammable materials, including textiles, to prevent/inhibit combustion. PBDEs, HBCD,  
528 organophosphorus compounds and bisphenol A have been extensively used in the  
529 textile industry as flame retardants. The concentrations of these substances reported in a  
530 number of studies suggest that dermal exposure may be a non-negligible way of human  
531 exposure to these toxic compounds. Although the dietary intake is the main route of  
532 human exposure to flame retardants, followed by inhalation (Domingo, 2012; van der  
533 Veen and de Boer, 2012; Linares et al., 2015), without a deep knowledge on the  
534 importance of exposure through the skin, the risks of dermal exposure should not be  
535 discarded/dismissed (Abdallah and Harrad, 2018; Abdallah et al., 2015). A similar  
536 conclusion can be also drawn for the presence of trace elements in clothes (Rizzi et al.,  
537 2014). Thus, we have found that the non-carcinogenic risks of Sb were above 10% of  
538 the safety limit ( $HQ > 1$ ) for dermal contact route, due to exposure with skin-contact  
539 clothes (Rovira et al., 2015). With respect to azo dyes and aromatic amines (AAs),  
540 contact allergy to textile dyes is well known (Ryberg et al., 2009; Isaksson et al., 2015;  
541 Malinauskienė et al., 2013; Coman et al., 2014). In addition, the scientific information  
542 above reviewed indicates that, for example, the presence of mutagenic AAs in textile  
543 azo dyes is of much high concern than previously expected (Brüschweiler and Merlot,  
544 2017). Similarly, the occurrence in clothes of other potentially harmful organic  
545 compounds such as quinolines, BTHs, BTRs, or bisphenol A, is also an issue of concern  
546 that requires further research on their presence in textiles (Luongo et al., 2014, 2016a).  
547 Recently, it has been even reported that perfluorinated compounds such as PFOS and  
548 PFOA, with a well known toxicity (Domingo and Nadal, 2017; Lóez-Doval et al., 2014;  
549 Kingsley et al., 2017), migrate from textiles, which could mean a significant direct and  
550 indirect source of human exposure to these chemicals (Supreeyasunthorn et al., 2016).  
551 Also, the information on human dermal exposure from textiles that contain metal  
552 (mainly silver) nanoparticles is still limited (Stefaniak et al., 2014; Tolve et al., 2015;  
553 Bianco et al., 2015, 2016).

554 In summary, the results of the current review focused mainly on the health risk  
555 assessment of human dermal exposure to potentially toxic chemical through skin-  
556 contact textiles/clothes, show a non-negligible presence of various chemicals in some

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557 textiles, which might lead to potential systemic risks. Even under specific circumstances  
558 of exposure, it might mean non-assumable cancer risks for the consumers. Therefore,  
559 we recommended elucidating which are being the chemicals of most concern in terms of  
560 dermal exposure through clothing. Studies should be conducted in order to prevent  
561 potential human health risks for the consumers, including the adult population, but very  
562 especially babies and children.

563

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