

1 **Air quality, health impacts and burden of disease due to air pollution (PM<sub>10</sub>, PM<sub>2.5</sub>, NO<sub>2</sub> and O<sub>3</sub>):**  
2 **Application of AirQ+ model to the Camp de Tarragona County (Catalonia, Spain)**

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22 **ABSTRACT**

23 The purpose of this study was to assess the impact to human health of air pollutants, through the  
24 integration of different technics: data statistics (spatial and temporal trends), population  
25 attributable fraction using AIRQ+ model developed by the WHO, and burden of disease using  
26 Disability-Adjusted Life Years (DALYs). The levels of SO<sub>2</sub>, NO, NO<sub>2</sub>, O<sub>3</sub>, H<sub>2</sub>S, benzene, PM<sub>10</sub>, PM<sub>2.5</sub>, CO,  
27 benzo(a)pyrene and metals, obtained between 2005 and 2017 from the air quality monitoring  
28 network across Camp de Tarragona County, were temporally and spatially determined. Health  
29 impacts were evaluated using the AIRQ+ model. Finally, the burden of disease was assessed through  
30 the calculation of Years of Lost life (YLL) and Years Lost due to Disability (YLD). In general terms, air  
31 quality was good according to European quality standards, but it did not fulfil the WHO guidelines,  
32 especially for O<sub>3</sub>, PM<sub>10</sub> and PM<sub>2.5</sub>. Several decreasing (NO, NO<sub>2</sub>, SO<sub>2</sub>, PM<sub>10</sub> and benzene) and an  
33 increasing (O<sub>3</sub>) temporal trend were found. Correlation between unemployment rate and air  
34 pollutant levels were found, pointing that the economic crisis (2008-2014) was a factor influencing  
35 the air pollutant levels. Reduction of air pollutant levels (PM<sub>2.5</sub>) to WHO guidelines in the Camp de  
36 Tarragona County would decrease the adult mortality between 23 and 297 cases per year, which  
37 means between 0.5 and 7% of all mortality in the area. In this County, for lung cancer, ischemic  
38 heart disease, stroke, and chronic obstructive pulmonary disease due to levels of PM<sub>2.5</sub> above the  
39 WHO threshold limits, DAYLs were 240 years. This means around 80 DALYs for 100,000 persons  
40 every year -between 2005 and 2017. Population attributable fraction (PAF) and burden of disease  
41 (DALYs) methodologies are suitable tools for regional and national policymakers, who must take  
42 decisions to prevent and to control air pollution and to analyse the cost-effectiveness of  
43 interventions.

44 *Keywords:* Air quality, AIRQ+ model, Population attributable fraction, burden of disease, DALYs,  
45 Camp de Tarragona County

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47 **Highlights**

- 48 • Pollutants levels from an air quality network were temporally and spatially studied.
- 49 • Mortality and health effects due to pollutant levels above WHO standards were assessed.
- 50 • If PM<sub>2.5</sub> fulfil WHO limits, mortality would decrease between 23 and 297 per year.
- 51 • DAYLs due to pollutant levels above WHO standards were 80 years for 100,000 individuals.
- 52 • Integration of AirQ+ and DALYs is suitable to support decisions to reduce air pollution.

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## 54 1. Introduction

55 Air pollution is one of the most challenging environmental problems, which must be faced at local,  
56 regional and global scale. Industrial activities, including energy production, road traffic, and  
57 household combustion, among others, have been identified as important emission sources of a wide  
58 range of pollutants (EEA, 2018; Prüss-Ustün et al., 2016). It is well known that air pollution is  
59 associated to adverse health effects such as respiratory and cardiovascular diseases, cancer and  
60 even death (Abdo et al., 2016; Ai et al. 2019; Brook, 2007; Chen et al., 2008, 2018; Demetriou et al.,  
61 2012; Farhat et al., 2011; Mafrici et al., 2008; Raaschou-Nielsen and Reynolds, 2006; Raaschou-  
62 Nielsen et al., 2013; Song et al., 2014; Sun et al., 2016; Sunyer, 2001; Wang et al. , 2019; Wu et al.,  
63 2019; Zhang et al. 2020). According to the World Health Organization (WHO), in 2012, 3.0 million of  
64 deaths were attributable to outdoor air pollution, while around 6.5 million of deaths (11.6% all world  
65 deaths) would attributable to indoor and outdoor air pollution, if they were taken into account  
66 together (WHO, 2016a). In a recent study (Stanaway et al. 2018), the attributable number of deaths  
67 to air pollution (PM and O<sub>3</sub>), for both indoor and outdoor environments, was set at 4.9 million in  
68 2017.

69 To support decisions of policymakers, it is necessary to quantify the effects of air pollution and the  
70 potential changes due to mitigation policies implementation. To measure health outcomes due to  
71 air pollution exposure, WHO developed the AIRQ+ model. This model has been applied  
72 internationally in Europe, but also worldwide (Al-Hemoud et al., 2018). AIRQ+ measures the  
73 contribution of air pollution to the mortality/morbidity using the population attributable fraction  
74 (PAF). PAF is the proportional reduction in a population morbidity or mortality, which would occur  
75 if exposure to air pollution was reduced to an alternative exposure scenario (WHO, 2016b), e.g.  
76 WHO air quality threshold limits (annual mean limits 10 and 20 µg/m<sup>3</sup> for PM<sub>2.5</sub> and PM<sub>10</sub>,  
77 respectively (WHO, 2006)). As complementary tool, environmental burden of disease using  
78 disability-adjusted life years (DALYs) can assess the health impact (mortality and morbidity) that  
79 may be attributed to environmental factors such as air pollution (Prüss-Ustün et al., 2016). The  
80 DALYs, as a way to compare overall health and life expectancy of different populations, take into  
81 account the years of life lost due to a premature death (YLL), as well as the years of healthy life lost  
82 due to a disease (YLD) (Devleeschauwer et al., 2014).

83 The current study has been focused on the Camp de Tarragona County (Catalonia, NE Spain), where  
84 since 1960s one of the most important petrochemical industrial complex of south-Europe is placed.  
85 In addition to the chemical and petrochemical companies (oil refinery, chlor-alkaly plant, plastic  
86 manufacturers, and chemical industries), a hazardous and a municipal solid waste incinerator (HWI  
87 and MSWI), an important industrial harbour, as well as a commercial airport, are also present in the  
88 same area. Because of the considerable number of potentially contaminant activities, the public  
89 concern for the air quality in the zone has promoted the constitution of the “Air quality panel of the  
90 Camp de Tarragona” by the Parliament of Catalonia, which is aimed at improving the air quality in  
91 the area and at assessing human health effects associated to it (TQACT, 2019).

92 The aim of present study was to assess the impact to human health of air pollutants in the Camp de  
93 Tarragona County between 2005 and 2017, through the integration of different technics (data  
94 statistics, population attributable fraction and burden of disease. To achieve this goal, the levels of  
95 a number of pollutants (SO<sub>2</sub>, NO, NO<sub>2</sub>, O<sub>3</sub>, H<sub>2</sub>S, benzene, PM<sub>10</sub>, PM<sub>2.5</sub> and CO), measured by an air  
96 quality monitoring network across the Camp de Tarragona County (from 2005 to 2017), were  
97 spatially and temporally analysed. In addition, the health impact of some pollutants (NO<sub>2</sub>, O<sub>3</sub>, PM<sub>10</sub>  
98 and PM<sub>2.5</sub>) was evaluated using the AirQ+ model version 1.3, developed by the WHO (2016b).  
99 Population attributable fraction (PAF) of mortality due to pollutant (PM<sub>10</sub>, PM<sub>2.5</sub>, NO<sub>2</sub> and O<sub>3</sub>) levels  
100 above the WHO thresholds was also estimated. Finally, the results obtained by PAF methodology  
101 were used to evaluate the burden of disease (DALYs) attributable to air pollution (PM<sub>10</sub>, PM<sub>2.5</sub>, NO<sub>2</sub>  
102 and O<sub>3</sub>), above the WHO threshold scenario, for the population living in the Camp de Tarragona  
103 County.

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## 105 **2. Materials and methods**

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### 107 *2.1. Area of study*

108 The Camp de Tarragona County (Fig. 1) is subdivided into three counties: Tarragonès, Alt Camp, and  
109 Baix Camp (South of Catalonia, NE Spain). There are two main cities in the area: Reus (103,615  
110 inhabitants in 2017) and Tarragona (131,094 inhabitants in 2017) (IDESCAT, 2019). The total  
111 population in this region ranged from 406,000 inhabitants in 2005 to 483,000 inhabitants in 2017  
112 spread in 2703 km<sup>2</sup> (IDESCAT, 2019). The largest petrochemical complex of southern Europe is  
113 located in this zone since 1960s. The petrochemical complex is divided in two sub-complexes, the  
114 north (petrochemical industries) complex and the south (chemical industries) complex. In addition,  
115 there are other potential pollutant activities in the area: a MSWI, a HWI, an industrial harbour, and  
116 a commercial airport. Obviously, there are also other diffuse emission sources, such as heating  
117 systems, agriculture burning, and a heavy traffic. Climate in the region is typical of Mediterranean  
118 basin. The annual mean temperature (2017) was 16°C, with a minimum and maximum temperature  
119 of -7°C and 36°C, respectively, and a total accumulated precipitation of 340 mm, more than 70%  
120 during the periods of March-April and September-October (Meteocat, 2019). Two different wind  
121 regimes are present in the area. One of them with predominant winds blowing from north, following  
122 Francolí River valley and affecting Alcover, Perafort, Sant Salvador, Constantí, Tarragona and Laboral  
123 stations, and the second one with predominant winds blowing from North-west, affecting Reus and  
124 Vilaseca stations (Meteocat, 2019). Moreover, winds regime blowing from south, from the sea, is  
125 important in the area during the summer.

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### 127 *2.2. Levels of pollutants*

128 The levels of SO<sub>2</sub>, NO, NO<sub>2</sub>, O<sub>3</sub>, H<sub>2</sub>S, benzene, PM<sub>10</sub>, PM<sub>2.5</sub>, As, metals (Cd, Ni and Pb), benzo[*a*]pyrene  
129 and CO were obtained from the Network of Vigilance and Forecast of the Quality of the Air (XVPCA)

130 of the Autonomous Government of Catalonia (Generalitat de Catalunya). Data are available online  
131 (AIR QUALITY, 2019). Table 1 summarizes the main characteristics of each air quality monitoring  
132 station. According to the XPVCA (XVPCA, 2019), the sampling and determination of pollutants were  
133 carried out following these standards: EN14212 for SO<sub>2</sub>; EN14211 for NO and NO<sub>x</sub>; EN12341 for  
134 PM<sub>10</sub>; EN14907 for PM<sub>2.5</sub>; EN14626 for CO; EN14625 for O<sub>3</sub>; EN14622 for benzene; EN12341,  
135 EN14902 and EN15852 for metals (As, Cd, Ni and Pb), and EN12341 and UNE15549 for  
136 benzo[*a*]pyrene. Daily and hourly mean levels of pollutants were obtained from January 1, 2005 to  
137 December 31, 2017, from the nine stations placed in the Camp de Tarragona (Alcover, Sant Salvador  
138 neighbourhood, Perafort, Bonavista neighbourhood, Constantí, Vilaseca, Reus, Tarragona and  
139 Laboral educational campus). The location and the pollutants analysed in each station are shown in  
140 Table 1 and Fig. 1.

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### 142 2.3. Health risk assessment of air pollution (Population attributable fraction).

143 Health effects due to exposure to air pollution were quantified using the AirQ+ model (version 1.3)  
144 developed by the WHO Regional Office for Europe. AirQ+ model was developed to calculate the  
145 magnitude of the burden and impacts of air pollution on health in a certain population. Health  
146 effects considered in the current study have been the following: long-term mortality (due to all  
147 natural causes), chronic obstructive pulmonary disease (COPD), lung cancer (LC), ischemic heart  
148 disease (IHD) and stroke. The model performs calculations that allow quantification of the health  
149 effects due to exposure to air pollution. The model is based on the attributable proportion (AP)  
150 defined as the portion of a health effect attributable to exposure to air pollution by a population  
151 (equation 1):

$$152 \quad AP = \frac{\sum((RR-1) \cdot P)}{\sum(RR \cdot P)} \quad (\text{equation 1}).$$

153 where RR is the relative risk for a given health endpoint in a determined exposure to an air pollutant,  
154 being P the fraction of population under exposure to the air pollutant.

155 The amount of health effect attributable to the exposure (IE), and the number of cases attributable  
156 to the exposure (NE), can be calculated by means of equations 2 and 3, respectively:

$$157 \quad IE = I \cdot AP \quad (\text{equation 2}).$$

$$158 \quad NE = IE \cdot N_p \quad (\text{equation 3}).$$

159 where I is the baseline frequency of health effect in the population under study, while N<sub>p</sub> is the size  
160 of the population. The parameters used for calculations are summarized in Table 2.

161 Air pollutants reductions from 2005-2017 levels to the ideal scenario “WHO threshold scenario”  
162 were considered. WHO threshold scenario consists in O<sub>3</sub> levels lower than 100 µg/m<sup>3</sup> in 8-hourly  
163 mean, and annual means of 20 and 10 µg/m<sup>3</sup> for PM<sub>10</sub> and PM<sub>2.5</sub>, respectively. For NO<sub>2</sub>, a mean  
164 annual value of 20 µg/m<sup>3</sup> was taken into account following the recommendations of the HRAPIE

165 project (WHO, 2013). Population data from the Camp de Tarragona -grouped by age- for the period  
166 2005-2017 were obtained from official statistics of Catalonia (IDESCAT, 2019). The incidence of  
167 mortality was obtained from the Health Department of Catalonia (Departament de Salut, 2019).  
168 Mortality incidence due to all natural causes and for each disease (chronic obstructive pulmonary  
169 disease (COPD), mortality due to lung cancer (LC), mortality due to ischemic heart diseases (IHD),  
170 and mortality due to strokes) were obtained from the Catalan Health System Observatory (CHSO,  
171 2019). Similar levels of NO<sub>2</sub> and O<sub>3</sub> were found in urban and industrial areas. Health impacts were  
172 calculated in urban/industrial areas, using a mean value of the air quality stations. Health outcomes  
173 were also calculated in rural areas for O<sub>3</sub>, but not for NO<sub>2</sub>, because its levels were below the  
174 threshold limit (20 µg/m<sup>3</sup>). Regarding PM<sub>2.5</sub>, it was only measured in urban and industrial sites,  
175 showing similar concentrations. As the levels of PM<sub>2.5</sub> -from 2005 to 2007- were not measured by  
176 air quality stations, they were calculated by multiplying PM<sub>10</sub> levels and PM<sub>2.5</sub>/PM<sub>10</sub> ratio, which was  
177 previously set at 0.60. NO<sub>2</sub> concentrations were used to analyse the attribution factor in the  
178 mortality due to natural causes, while O<sub>3</sub> concentrations were used to analyse the attribution factor  
179 in the mortality because of respiratory diseases. PM<sub>2.5</sub> levels were used to analyse the attribution  
180 factor in mortality due to all natural causes (for adults above 30-years old), mortality due to chronic  
181 obstructive pulmonary disease (COPD), mortality due to lung cancer (LC), mortality due to ischemic  
182 heart diseases (IHD), and mortality due to strokes. To assess human health impacts using AIRQ+  
183 model, a ratio of 0.60 was used in order to convert PM<sub>10</sub> levels to PM<sub>2.5</sub> at the stations where data  
184 were lacking

#### 185 2.4. Burden of disease (DALYs).

186 Burden of disease was assessed using Disability-Adjusted Life Years (DALYs), which are defined as  
187 the loss of healthy life years (Devleeschauwer et al., 2014). They are calculated as the sum of the  
188 Years of Life Lost (YLL) due to premature deaths, and the Years Lost due to Disability (YLD) for people  
189 living with the health consequences of the respective disease (equations 4-6).

$$190 \text{ DALYs} = \text{YLL} + \text{YLD} \quad (\text{equation 4})$$

$$191 \text{ YLL} = N_d \cdot L \quad (\text{equation 5})$$

$$192 \text{ YLD} = I_c \cdot DW \cdot D \quad (\text{equation 6})$$

193 where N<sub>d</sub> is the number of deaths, L is the standard life expectancy at the age of death in years, I<sub>c</sub>  
194 is the number of incident cases, DW is the disability weight factor, which indicates the severity of  
195 the disease from 0 (perfect health) to 1 (death) (WHO, 2019), and D is the average duration of the  
196 case until remission or death (years). Since no information about adverse health outcomes onset  
197 and their duration was found, mean age of diagnosis, and survival rate, were obtained from the  
198 scientific literature (Almagro et al., 2002; Chirlaque et al., 2018; Clua-Espuny, 2014; Miravittles et  
199 al., 2009; Moran et al. 2014; Salmerón et al. 2012; Taylor et al., 2019). These data were used  
200 according to the following origin criteria: Catalan population, Spanish population, and West-Europe  
201 population.

202

## 203 2.5. Statistics

204 The statistical software package XLSTAT (Version 2015.2.02.18681) and IBM SPSS statistics were  
205 used to perform the statistical analysis of data. To evaluate significant differences between groups,  
206 a Levene test was performed to evaluate variances. ANOVA (parametric data) or Kruskal-Wallis  
207 (non-parametric data) test were subsequently applied. Significance level was set at  $p < 0.05$ .  
208 Correlations between variables were assessed using the Pearson's coefficient. Principal components  
209 analyses (PCA) with Varimax rotation were applied to reduce the number of variables and to  
210 facilitate the patterns recognition. Agglomerative hierarchical clustering (dendrograms) were  
211 performed with the Ward method using dissimilarities (Euclidian distances). Annual mean levels  
212 were used to determine temporal trends with Mann-Kendall test at significance level of 0.05.

213

## 214 3. Results

215

### 216 3.1. Levels of pollutants

217 The levels of SO<sub>2</sub>, NO, NO<sub>2</sub>, O<sub>3</sub>, H<sub>2</sub>S, benzene, PM<sub>10</sub>, PM<sub>2.5</sub> and CO registered in the nine stations  
218 located in the Camp de Tarragona from 2005 to 2017 are shown in Table 2 and Fig. 2. Annual air  
219 mean pollutants levels trends at a significance level of 0.05 were summarized at Table S1  
220 (Supplementary materials).

221 Mean annual levels of SO<sub>2</sub> in the nine stations, presented a clear decrease tendency ( $p < 0.05$ ) from  
222 2005 (4.9  $\mu\text{g}/\text{m}^3$ ) to 2017 (2.8  $\mu\text{g}/\text{m}^3$ ). A similar significant decreasing trend ( $p < 0.05$ ) was observed  
223 for the four areas: urban, rural, and north complex, meanwhile no significant ( $p > 0.05$ ) trends were  
224 noted in the South complex. However, between 2015 and 2017 a slight increase was noticed in the  
225 north and south industrial complexes. European air quality standards set a SO<sub>2</sub> daily average limit  
226 of 125  $\mu\text{g}/\text{m}^3$ , a limit that cannot be reached more than 3 times/year (European Union Parliament  
227 and Council, 2008). In the Camp de Tarragona and for the period studied, there were between 2-5  
228 exceedances between 2010 and 2017 (especially in the Constantí station), and up to 8 exceedances  
229 in the period 2005-2007. WHO air quality Guidelines for SO<sub>2</sub> establishes a daily mean of 20  $\mu\text{g}/\text{m}^3$   
230 (WHO, 2006). In recent years (2012-2017), the WHO guideline was exceeded up to 1 time (except  
231 for the Alcover station -placed in a rural area-, where it was exceeded 6 times in 2016). In contrast,  
232 between 2005 and 2009, it was exceeded up to 20 times/year, mainly in the urban, north and south  
233 complexes.

234 Regarding NO<sub>2</sub>, annual mean levels measured in all stations showed a decrease trend ( $p < 0.05$ ) from  
235 23.5  $\mu\text{g}/\text{m}^3$  in 2005, to 18.3  $\mu\text{g}/\text{m}^3$  in 2017. Similar significant ( $p < 0.05$ ) decreasing trends were found  
236 for the four areas. However, higher levels were detected in the urban and south complex (more  
237 affected by urban environment) than the north complex. The rural area presented the lowest NO<sub>2</sub>  
238 levels. A similar trend than that of NO<sub>2</sub> was also observed for NO, with a significant decreasing trend  
239 ( $p < 0.05$ ) from 7.3  $\mu\text{g}/\text{m}^3$  in 2005 to 4.9  $\mu\text{g}/\text{m}^3$  in 2017. The same decreasing trends were found in all  
240 the areas, with the exception of the rural zones. where no significant ( $p > 0.05$ ) temporal trend was

241 observed. Similar differences than those for NO<sub>2</sub> were observed for the four areas (Urban>South  
242 complex>North complex>Rural). European and WHO air quality standards set NO<sub>2</sub> annual and daily  
243 mean limits of 40 µg/m<sup>3</sup> and 200 µg/m<sup>3</sup>, respectively (European Union Parliament and Council, 2008;  
244 WHO, 2006). For the period evaluated, the annual limit was not reached in any station of the Camp  
245 de Tarragona. In turn, the daily limit was reached few times (1-3) per year in the South complex  
246 (Bonavista and Laboral stations), but it never surpassed the 18 exceedances/year allowed by the  
247 legislation.

248 Ozone (O<sub>3</sub>) was analysed only in five stations (Alcover, Constantí, Vilaseca, Reus and Tarragona).  
249 Ozone presented a significant (p<0.05) increasing trend over the years from 53.6 µg/m<sup>3</sup> in 2005 to  
250 59.8 µg/m<sup>3</sup> in 2017. This increasing trend was also significant (p<0.05) for rural, urban and North  
251 complex, but did not reach the level of significance (p>0.05) in the South complex between 2005  
252 and 2007. Rural area (Alcover station) showed higher O<sub>3</sub> levels than those found in the rest of  
253 stations (Constantí, Vilaseca, Reus and Tarragona), which showed similar levels among them.  
254 According to the European legislation, the maximum daily 8-hours mean (120 µg/m<sup>3</sup>) was exceeded  
255 between 16 and 20 times in the Alcover station, and below 8 in the other four stations. However,  
256 this number of exceedances does not surpass the 25 exceedances allowed by the EU air quality  
257 standards (European Union Parliament and Council, 2002). Taking into account WHO (2006)  
258 guidelines, the 8-hour limit (100 µg/m<sup>3</sup>) was reached 25 times in Alcover.

259 For H<sub>2</sub>S, no significant (p>0.05) temporal trends were observed. No spatial differences were noticed  
260 despite the high annual levels registered in 2007 and 2009 in the North complex. The concentrations  
261 were between 1 and 2 µg/m<sup>3</sup>. Nowadays, there is no European or WHO directives that establish  
262 maximum concentrations for this pollutant. However, Spanish legislation (Real Decreto 102, 2011)  
263 set at 40 µg/m<sup>3</sup> daily limit, and 100 µg/m<sup>3</sup> half-hourly limit. Some exceedance in half-hourly limit,  
264 especially in Constantí, Perafort and Laboral stations- before 2011- were noticed. However, in  
265 general terms these limits were accomplished.

266 Benzene levels did not show significant trends (p>0.05) along all period, with the exception of the  
267 North complex (Constantí), where a significant (p<0.05) decreasing trend was found between 2005  
268 and 2017. Between 2005 and 2010, in the Constantí station, benzene registered higher levels  
269 (p<0.05) (between 3.7 and 6.2 µg/m<sup>3</sup>) than those during in the period 2011-2017. Without  
270 considering this period and station, benzene levels were far below the air quality standard annual  
271 limit set in 5 µg/m<sup>3</sup> (European Union Parliament and Council, 2008).

272 Regarding PM<sub>10</sub>, a clear decrease was observed between 2005 (46.5 µg/m<sup>3</sup>) and 2013 (16.9 µg/m<sup>3</sup>),  
273 followed by periods with constant concentrations: 19.1-23.4 µg/m<sup>3</sup>. PM<sub>10</sub> presented a significant  
274 (p<0.05) decreasing trend between 2005 and 2017, if all stations were considered and for each area  
275 (rural, urban, North and South complex). No differences (p>0.05) were noticed in PM<sub>10</sub> levels  
276 regarding the location of each station. PM<sub>2.5</sub> levels were measured since 2008. Mean annual PM<sub>2.5</sub>  
277 concentrations did not show a significant (p>0.05) trend, ranging from 10.7 to 11.8 µg/m<sup>3</sup> between  
278 2008 and 2017, respectively. In both, North and South complex, similar levels (p>0.05), and no  
279 significant (p<0.05) trends, were observed. The annual PM<sub>2.5</sub>/PM<sub>10</sub> ratio was similar in all stations

280 and ranged from 0.40 to 0.71. Mean ratio for the 2008-2017 period, when both PM were analysed,  
281 was set between 0.58 and 0.64, depending on the station. Regarding European air quality standard,  
282 annual PM<sub>10</sub> limit and daily limit are 40 and 50 µg/m<sup>3</sup>, respectively. The daily limit cannot be  
283 exceeded more than 35 times/year (European Union Parliament and Council, 2008). Several stations  
284 exceeded air quality standard annual limits for PM<sub>10</sub> in 2005 and 2006. Thus, Vilaseca station in  
285 2015, and Reus and Constantí stations in 2006, exceeded more than 35 times the daily limit (50  
286 µg/m<sup>3</sup>). Otherwise, the WHO established air quality guidelines for annual mean PM<sub>10</sub> levels of 20  
287 µg/m<sup>3</sup> (WHO, 2006). This level was exceeded in almost all stations in the Camp de Tarragona. For  
288 PM<sub>2.5</sub>, European air quality standard is set at annual mean levels of 25 µg/m<sup>3</sup> (European Union  
289 Parliament and Council, 2008), while the WHO air quality guideline established annual and daily  
290 mean levels of 10 and 25 µg/m<sup>3</sup>, respectively (WHO, 2006). Similarly, to PM<sub>10</sub>, PM<sub>2.5</sub> concentrations  
291 in the Camp de Tarragona were also below the European standard, but they did not fulfil the WHO  
292 guidelines for annual and daily mean levels.

293 Only three stations (Alcover, Reus and Tarragona) measured CO for all the studied period (2005-  
294 2017). No significant ( $p > 0.05$ ) temporal trends were noted despite that the mean levels decreased  
295 from 0.4 mg/m<sup>3</sup> in 2005 to 0.2 mg/m<sup>3</sup> in 2017. Similar levels were found in these three stations. No  
296 exceedances of the 8-hourly limits (10 mg/m<sup>3</sup>) were detected between 2005 and 2017 in these three  
297 stations (European Union Parliament and Council, 2008).

298 In addition to these pollutants, the concentrations of benzo(*a*)pyrene were analysed  
299 discontinuously (14% of days) in one station (Constantí) since 2008, while metals (As, Cd, Ni and Pb)  
300 were analysed discontinuously (14% of days) in Vilaseca, Reus and Constantí stations.  
301 Benzo(*a*)pyrene registered an annual mean value between 0.12 and 0.18 ng/m<sup>3</sup>, which is below the  
302 annual mean limit of 1 ng/m<sup>3</sup> established by the European legislation. Nowadays, the WHO has not  
303 set air quality guidelines for benzo(*a*)pyrene. Regarding metals, in Constantí and Vilaseca stations,  
304 the mean annual levels in the period 2005-2017 ranged between 0.5 and 2.1 ng/m<sup>3</sup> for As, 0.1 to 3.1  
305 ng/m<sup>3</sup> for Cd, 2.4 to 7.4 ng/m<sup>3</sup> for Ni, and 2.4 to 12.4 ng/m<sup>3</sup> for Pb. According to the European  
306 legislation (European Union Parliament and Council, 2008, 2004), the annual mean levels of these  
307 metals cannot exceed 500 ng/m<sup>3</sup>, 20 ng/m<sup>3</sup>, 5 ng/m<sup>3</sup> and 6 ng/m<sup>3</sup> for Pb, Ni, Cd, and As, respectively.  
308 As for benzo(*a*)pyrene, metal levels were below the European annual threshold, taken into account  
309 that they were analysed only 14% of days.

310 High positive Pearson's correlation values were found between NO and NO<sub>2</sub> (0.820;  $p < 0.01$ ) and  
311 between NO<sub>2</sub> and NO, and PM<sub>10</sub> (0.551 and 0.768, respectively, both at  $p < 0.01$ ), while there were  
312 and negative correlations between NO and O<sub>3</sub> (-0.616;  $p < 0.01$ ) and NO<sub>2</sub> and O<sub>3</sub> (-0.788;  $p < 0.01$ )  
313 (Table 3). In addition, moderate but significant ( $p < 0.01$ ) negative correlations were found between  
314 O<sub>3</sub> and PM<sub>10</sub>. Finally, moderate although significant ( $p < 0.01$ ) positive correlations were found  
315 between nitrogen oxides (NO and NO<sub>2</sub>) and particulate matter (PM<sub>10</sub>).

316 Figure 3 depicts principal components analysis (PCA) and dendrograms of 2005-2017 annual mean  
317 levels of pollutants measured in the Camp de Tarragona stations. As not all pollutants were  
318 measured in all stations, therefore, three PCA and dendrograms were carried out to cover all the

319 stations and most analysed pollutants. PM<sub>2.5</sub> was not included because it was analysed only in few  
320 stations, showing a high linear dependency with PM<sub>10</sub>. Moreover, CO was not taken into account  
321 because of similar levels in the three stations measured. Three analysis (PCA plus dendrogram) were  
322 done. The first one (Fig. 3a) included the 9 stations and takes into account NO<sub>2</sub>, NO and H<sub>2</sub>S. The  
323 second one (Fig. 3b), with 6 stations, includes NO<sub>2</sub>, NO, SO<sub>2</sub>, H<sub>2</sub>S and PM<sub>10</sub>. Finally, the last one (Fig.  
324 3c), considers NO<sub>2</sub>, NO, SO<sub>2</sub>, H<sub>2</sub>S and O<sub>3</sub> in 5 stations.

325

### 326 *3.2. Attributable proportion and burden of disease (DALYs).*

327 Population attributable fraction (PAF) methodology was applied to the Camp de Tarragona between  
328 2005 and 2017, using the AirQ+ model.

329 Figure 4a depicts cases of mortality due all natural causes attributable to air pollutants (PM<sub>2.5</sub>, NO<sub>2</sub>  
330 and O<sub>3</sub>) above the WHO threshold scenario in the Camp de Tarragona. Excess cases of mortality due  
331 PM<sub>2.5</sub> exposure ranged from 297 (with a 95% confidence interval between 196 and 389 (196, 389;  
332 CI95%)) in 2005, to 23 (15, 30; CI95%) in 2016. This corresponds to 7% in 2005 and 0.5% in 2016 of  
333 total mortality cases in the area. In the period 2008-2017, when PM<sub>2.5</sub> levels were directly analysed  
334 and not estimated from PM<sub>10</sub>, excess attributable mortality cases ranged from 23 (15, 30; CI95%) to  
335 137 (90, 181; CI95%) in 2016 and 2015, respectively. Regarding NO<sub>2</sub>, in urban and industrial areas,  
336 the excess of mortality cases ranged from 71 (33, 108 CI95%) in 2005 to 1 (1, 1; CI95%) in 2017. No  
337 cases were noticed in 2013, 2014 and 2016, because environmental levels were below the cut off  
338 limit (20 µg/m<sup>3</sup>). The proportion of attributable cases to NO<sub>2</sub> levels above 20 µg/m<sup>3</sup> was 1.4%. The  
339 proportion of attributable cases NO<sub>2</sub> levels was only calculated in urban areas cause in rural areas  
340 annual mean levels were below WHO threshold limit (20 µg/m<sup>3</sup>). Finally, for O<sub>3</sub>, attributable excess  
341 mortality due to respiratory diseases cases were almost constant between 6 (2, 10; CI95%) and 9 (3,  
342 15; CI95%) in urban and rural areas, respectively. More than 80% of attributable mortality cases due  
343 to O<sub>3</sub> belongs to the urban area. Despite in rural areas the O<sub>3</sub> levels were higher than those in urban  
344 and industrial areas, only around 10% of population of the Camp de Tarragona is living in rural areas.  
345 However, the percentages of attributable cases of mortality due to respiratory causes because of  
346 the concentrations of O<sub>3</sub> -above the WHO guidelines- were around 1.75% in urban areas, and 3.70%  
347 in rural areas. From 2005 to 2007, PM<sub>2.5</sub> were not directly measured, but calculated from PM<sub>10</sub>. For  
348 this reason, we have considered only data of PM<sub>2.5</sub> from the last ten years (2008-2017). Mortality  
349 cases in adults older than 30 years, attributable to levels above the annual WHO threshold (10  
350 µg/m<sup>3</sup>, ranged from 23 (15, 30; CI95%) to 137 (90, 181; CI95%), which means percentages of 0.5 and  
351 2.85% of mortality by natural causes. Figure 4b depicts the number of mortality excess cases due  
352 COPD, LC, IHD and stroke attributable to air pollution (PM<sub>2.5</sub>) above the WHO threshold. In the last  
353 ten years (2008-2017), the annual mean excess cases of mortality were 4 (2, 6; CI95%) due to COPD;  
354 2 (1, 4; CI95%) due to LC; 5 (3, 10; CI95%) due to IHD, and 1 (1, 2; CI95%) due stroke.

355 Based on all natural causes mortality above 30-years old, and assuming a mean mortality age of 56  
356 year (mean between 30 and life expectancy in the region, 82), the YLL in the Camp de Tarragona  
357 due to PM<sub>2.5</sub> above the WHO threshold scenario, was between 7842 (5178, 10274; CI95%) years in

358 2005, and 570 (372, 755; CI95%) years in 2016. Consequently, mean YLL per 100,000 individuals was  
359 823 years (541, 1084; CI95%) for the 2005-2017 period, and 483 years (316, 638; CI95%) for the  
360 2008-2017 period, when the concentrations of PM<sub>2.5</sub> were measured. NO<sub>2</sub> levels above the WHO  
361 threshold supposes a mean of 163 YLL/100,000 individuals (77, 251; CI95%) in urban areas of the  
362 Camp de Tarragona County for the 2005-2017 period. Regarding O<sub>3</sub>, the mean YLL per 100,000  
363 individuals due to levels above the WHO threshold scenario was 40 (14, 68; CI95%) for the 2005-  
364 2017 period. Pollutant (NO<sub>2</sub>, PM<sub>2.5</sub> and O<sub>3</sub>) levels above the WHO threshold resulted in an YLL mean  
365 value of 686 (407, 953; CI95%) years/100,000 subjects for the 2008-2017 period. It should be noticed  
366 that summing the mortality caused by these air pollutants could induce to a double counting  
367 because some pollutants here considered are usually highly correlated. Using the mortality cause  
368 for each disease obtained from the AIRQ+ model, as well as survival rates and age of  
369 diagnosis/occurrence found in the literature, YLD were calculated. In the Camp de Tarragona, mean  
370 lost due to disability (YLD) were 19.7 years/100,000 individuals (10.7, 32.5; CI95%), or 60.2 years for  
371 all the Camp de Tarragona population (32.6, 99.5; CI95%) in the 2005-2017 period, for LC, IHD,  
372 stroke and COPD. Taken into account both results, YLL and YLD, DALYs in the Camp de Tarragona for  
373 lung cancer, ischemic heart disease (IHD), stroke and chronic obstructive pulmonary disease (COPD),  
374 due to PM<sub>2.5</sub> levels above the WHO threshold limits, was 240, with 25% coming from YLL. It means  
375 around 80 DALYs for 100,000 subjects (78 years/100,000 subjects) with minimum and maximum  
376 values of 16 years/100,000 subjects in 2016, and 264 years/100,000 subjects in 2005, respectively.

377

## 378 **4. Discussion**

### 379 *4.1. Levels of pollutants*

380 Based on the levels of the pollutants (NO<sub>2</sub>, NO, H<sub>2</sub>S, PM<sub>10</sub>, PM<sub>2.5</sub>, CO, benzene, metals and benzo[*a*  
381 pyrene]) analysed in the Camp de Tarragona, the air quality is good according to the European  
382 thresholds. However, it did not fulfil the WHO guidelines, especially for O<sub>3</sub> in rural area, PM<sub>10</sub> and  
383 PM<sub>2.5</sub>. The high O<sub>3</sub> levels registered in Alcover (rural area) are from local origin, due to the orography  
384 of the zone (Jiménez and Baldasano, 2004). Typical winds in the area, from sea to land, move O<sub>3</sub> or  
385 its precursors (NO<sub>2</sub> and VOCs), produced in urban areas (Tarragona) and in the petrochemical  
386 complex. They could have been accumulated in the Alcover area because of the presence of  
387 northern mountains. For both PM<sub>10</sub> and PM<sub>2.5</sub>, in general terms, EU standard were accomplished  
388 with few exceptions, by contrast, WHO air quality guidelines were not fulfilled in almost all stations.

389 WHO air quality guidelines and EU air quality standard have different roles. WHO air quality  
390 guidelines are based on scientific evidence to protect human health. However, EU standards can  
391 have legal consequences once implemented if they are not accomplished. Moreover, EU air quality  
392 standards take into account not only scientific evidences, but also economic, technical, and political  
393 aspects (Schneider et al., 2004). For the purpose of the current study, WHO air quality guidelines  
394 were taken into account and as commented previously, O<sub>3</sub> in rural area, PM<sub>10</sub> and PM<sub>2.5</sub> surpassed  
395 the WHO guidelines continuously between 2005-2017 in the Camp de Tarragona County.

396 PM<sub>10</sub>/PM<sub>2.5</sub> ratios, ranging from 0.58 to 0.64, were similar to those proposed by the WHO (2013)  
397 (PM<sub>2.5</sub>/PM<sub>10</sub>=0.65), being equivalent to those found in the literature for the Mediterranean  
398 European region (CAFE, 2004; Contini et al., 2014; Ostro, 2004; Querol et al., 2004, 2008).  
399 Consequently, as previously mentioned in the Materials and Methods section, to assess human  
400 health impacts using AIRQ+ model, a ratio of 0.60 was used in order to convert PM<sub>10</sub> levels to PM<sub>2.5</sub>  
401 at the stations where data were lacking.

402 It should be taken into account that between 2005 and 2017 several significant decreasing trends  
403 were found especially for nitrogen oxides, SO<sub>2</sub>, PM<sub>10</sub>, and benzene, meanwhile an increasing trend  
404 was registered for O<sub>3</sub>. Similar trends were registered in other European studies (Borge et al. 2019;  
405 Fenech et al. 2020; Font et al. 2019).

406 As expected, positive correlations between PM<sub>10</sub> and PM<sub>2.5</sub> and between nitrogen oxides (NO<sub>2</sub> and  
407 NO) were found sources. Correlations between nitrogen oxides (NO and NO<sub>2</sub>) and particulate matter  
408 (PM<sub>10</sub>) were probably due to the common sources of these pollutants: traffic, heating systems,  
409 industries, and other combustion processes (Núñez-Alonso et al. 2019). However a negative  
410 correlation were found between O<sub>3</sub> and other pollutants (NO, NO<sub>2</sub> and PM<sub>10</sub>). As suggested by some  
411 authors (Jia et al., 2017; Li et al., 2011), particulate matter absorb solar radiation, and in  
412 consequence, O<sub>3</sub> formation is reduced. Negative correlations between nitrogen oxides (NO and NO<sub>2</sub>)  
413 and O<sub>3</sub> have been reported in previous studies (Beckerman et al., 2008; Fecht et al., 2016; Kumar et  
414 al., 2015). Nitrogen oxides, O<sub>3</sub> and VOC cycles during the day, nights, as well as differences in  
415 weekend and weekday, would explain these correlations (Alghamdi et al., 2014; Han et al., 2011).

416 The period of time (2005-2017) here studied includes the years of recent economic crisis, which in  
417 Catalonia meant an increase of the unemployment rate from 8.9% in 2008 to 16.2% in 2009,  
418 reaching a maximum of 23.1% in 2013. After that, the unemployment rate decreased to 13.4% in  
419 2017. It is a clear indicator of the economic activity in the Camp de Tarragona, being related with  
420 the air quality. For this reason, Pearson's correlation was performed between annual mean values  
421 of air pollutants and unemployed rate as an indicator of the economic crisis (Table S2).  
422 Unemployment rate correlated negatively and significantly (p<0.05) with almost all pollutants NO<sub>2</sub>  
423 (-0.627), NO (-0.850), SO<sub>2</sub> (-0.760), PM<sub>10</sub> (-0.811) and benzene (-0.617), but positively with O<sub>3</sub>  
424 (0.824). Decreased trends were observed in PM<sub>10</sub>, NO<sub>2</sub> (especially in urban areas), NO, SO<sub>2</sub>, H<sub>2</sub>S (in  
425 north complex) and benzene (in north complex). These decreasing trends started in 2009, just when  
426 the economic crisis started in Spain. Some of these levels increased again in 2014, when the  
427 economic crisis began to remit. Similar trends and impacts of economic crisis in air quality and air  
428 pollutants emission were found in European countries and US (Monteiro et al. 2018, Pacca et al.  
429 2020; Squizzato et al. 2019; Tong et al. 2016, Tzima et al. 2018).

430 Regarding spatial analysis (Figure3), Vilaseca and Reus stations, affected from wind blowing NO-SE  
431 direction, appear close together in PCA and dendrograms (Figures 3b and 3c). These two stations  
432 were impacted by the urban environment and were less affected by South and North industrial  
433 complexes because of prevalent wind directions. Tarragona, an urban station, was represented  
434 alone (Figures 3a, 3b and 3c) and showed a high correlation with NO<sub>2</sub>. The Laboral and Bonavista

435 stations appeared together since they are placed together and influenced by the South complex,  
436 urban background and the harbour (Figures 3a and 3b). Perafort and Alcover stations, both rural,  
437 are influenced by the industrial activities and appeared close together (Fig. 3c), with a negative  
438 correlation with nitrogen oxides. Alcover station appeared highly correlated with O<sub>3</sub> (Fig. 3a). As  
439 above commented, Alcover is an O<sub>3</sub> hotspot due to prevalent winds and orography. Finally,  
440 Constantí station did not show similarities with the rest of stations. Constantí station showed a high  
441 positive correlation with SO<sub>2</sub> and H<sub>2</sub>S (Fig 3). This station is placed in the Francolí River valley, which  
442 is highly affected by the North complex, the MSWI, Tarragona city and Harbour, since prevalent  
443 winds blow from North and South.

444 In addition to pollutant previously reported, benzo(a)pyrene, the only polycyclic aromatic  
445 hydrocarbon (PAH) analysed, was measured only in Constantí station each year during the 14% of  
446 days. Similarly, metal (As, Cd, Ni and Pb) levels, were only analysed in three stations (Reus, Constantí  
447 and Vilaseca) at 15% of the days/year. In turn, other pollutants such as VOCs, (i.e., 1,3-butadiene,  
448 acrylonitrile and vinyl chloride), metals (V), or other PAHs, characteristics of industrial sources  
449 present in Camp de Tarragona (Chen et al., 2016; Domínguez-Morueco et al. 2015;  
450 Kampeerawipakorn et al., 2017; Kwon et al., 2016; Visschedijk et al., 2013) are not measured by the  
451 air quality network. However, there are previous data and monitoring studies in the area, which are  
452 focused on specific pollutants such PAHs and VOCs (Domínguez-Morueco et al., 2019, 2017; Ramírez  
453 et al., 2012, 2011), or on specific contaminants (PCDD/Fs, PCBs, metals) and industrial sources such  
454 as HWI (Bordonaba et al., 2011; Mari et al., 2013, 2014; Nadal et al., 2009a; Vilavert et al., 2010),  
455 MSWI (Vilavert et al., 2012, 2014, 2015) and chemical and petrochemical complexes (Domínguez-  
456 Morueco et al., 2017, 2019; Gallego et al. 2018; Nadal et al., 2009b, 2011; Ras et al., 2009). Based  
457 on these results, specific pollutants, such as metals, COVs and PAHs, which are representative of  
458 industrial activities in the area, should be measured with a higher frequency in order to improve the  
459 Camp de Tarragona air quality monitoring network.

460

#### 461 *4.2. Attributable proportion and burden of disease (DALYs).*

462 In the present study, mortality due all-natural causes attributable to PM<sub>2.5</sub> above the WHO threshold  
463 scenario in the Camp de Tarragona ranged from 6.9 to 41.7 cases/100,000 individuals. These results  
464 were higher than in other Mediterranean cities such as Perpignan, Marseille, Nice and Livorno (3.78,  
465 7.07, 9.71, and, 10.67 cases/100,000 individuals, respectively), but similar or even lower than those  
466 found in other Mediterranean cities such as Naples or Rome (55.20 and 34.75 cases/100,000  
467 individuals, respectively) (Sicard et al. 2019). However, for mortality due to respiratory diseases  
468 attributable to O<sub>3</sub> above the WHO threshold scenario, the same Mediterranean cities (ranging from  
469 1.77 to 5.17 cases/100,000 individuals for Rome and Livorno, respectively) showed slightly higher  
470 values than those of the current study (range 1.34 to 1.84 cases/100,000 individuals) (Sicard et al.  
471 2019).

472 Results for YLL (823, 163 and 40 YLL/100,000 individuals for PM<sub>2.5</sub>, NO<sub>2</sub> and O<sub>3</sub>, respectively) are in  
473 agreement with the European air quality report that, for Spain, estimated 658, 209 and 43  
474 YLL/100,000 individuals due to PM<sub>2.5</sub>, NO<sub>2</sub> and O<sub>3</sub>, respectively (EEA, 2018).

475 Due to the lack of data for the region, to calculate YLD, attributable to air pollutant levels above the  
476 WHO guidelines for the population of the Camp de Tarragona, was not possible. However, to  
477 calculate the DAYLs (YLL+YLD) of COPD, LC, IHD and stroke due to air pollutant levels above WHO  
478 guidelines (Fig. 5), several approaches were done (survival rate, average age disease detection, and  
479 time with the diseases were obtained from the scientific literature) (Almagro et al., 2002; Chirlaque  
480 et al., 2018; Clua-Espuny, 2014; Miravittles et al., 2009; Moran et al. 2014; Salmerón et al. 2012;  
481 Taylor et al., 2019). In the Camp de Tarragona, for lung cancer, ischemic heart disease (IHD), stroke  
482 and chronic obstructive pulmonary disease (COPD) due to PM<sub>2.5</sub> levels above the WHO threshold  
483 limits, DAYLs was 240 years, with a 25% coming from YLL. This means around 80 DALYs for 100,000  
484 individuals (78 years/100,000 individuals) annually between 2005 and 2017. According to the WHO  
485 (2016c), in Spain in 2012, DALYS from due ambient air pollution was set in 286 years/100,00  
486 individuals, with uncertainty intervals (97.5% CI) between 53 and 469 years/100,00 individuals. The  
487 results of present study, even in the low range of results reported by the WHO, are of the same  
488 order of magnitude.

489 This study present some limitations, for example, acute health effects for O<sub>3</sub> or PM outbreak  
490 episodes were not taken into account. In the present investigation, only health effects of chronic  
491 exposure to air pollutant levels were assessed. It is well known that health effects due to acute  
492 exposure are not negligible in terms of health outcomes (Kloog et al. 2012), which should be taken  
493 into account in future research. However, the lack of health outcomes data with enough temporal  
494 resolution makes impossible to assess acute health effects to air pollutants peak levels. It should be  
495 taken into account that PM shape and chemical components differs in each location and period, and  
496 consequently, health effects may be different (Sicard et al. 2019). In addition, concentrations were  
497 measured in air quality monitoring stations for all the territory, and they did not take into account  
498 the real pollutant levels indoor (where usually people expend more time), or other  
499 microenvironments. The use and integration of agent based modelling, remote sensing, and  
500 personal sensors could help to solve these limitations (Reis et al., 2015). Finally, individual  
501 differences among subgroups, such as elderly, children or chronic patients, are an important issue  
502 to take into account and it should be also considered in further studies.

503

## 504 **5. Conclusions**

505 In general terms, according to the pollutant levels analysed by air quality network in the Camp de  
506 Tarragona County, the air quality was good based on the European air quality standards. However,  
507 it did not fulfil the WHO guidelines, especially for O<sub>3</sub>, PM<sub>10</sub> and PM<sub>2.5</sub>. An economic indicator such  
508 as the unemployment rate significantly correlates (negatively) with several air pollutant (NO<sub>2</sub>, NO,  
509 SO<sub>2</sub> and PM<sub>10</sub>) levels that remark the influence of the economic crisis (2009-2014) in air pollutant  
510 levels. Reduction of air pollutants levels (PM<sub>2.5</sub>) in the Camp de Tarragona to WHO guidelines should  
511 reduce the adult mortality between 23 and 297 cases per year. It means between 0.5 and 7% of all

512 mortality in the area. Lack of data (metals, PAHs and VOCs), or incomplete data (benzo(a)pyrene  
513 and metals were only analysed in 14% and 15% of the days, respectively) are the main weak points  
514 of the air quality network in the Camp de Tarragona. Moreover, specific pollutants such as metals,  
515 VOCs and PAHs, which are representative of industrial activities in the zone, should be more  
516 frequently measured to improve the air quality monitoring network. The integration of population  
517 attributable fraction (PAF) and burden of disease (DALYs) methodologies, as well as statistical tools  
518 to find temporal and spatial trends, is a suitable approach to assess the impact to human health.  
519 This approach should help to regional and national policymakers, who must take/support decisions  
520 to prevent and control air pollution and to analyse the cost-effectiveness of interventions.

521

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525

## 526 **References**

- 527 Abdo, N., Khader, Y.S., Abdelrahman, M., Graboski-Bauer, A., Malkawi, M., Al-Sharif, M., et al., 2016.  
528 Respiratory health outcomes and air pollution in the Eastern Mediterranean Region: A  
529 systematic review. *Rev. Environ. Health* 31, 259–280. doi:10.1515/reveh-2015-0076
- 530 Al-Hemoud, A., Gasana, J., Al-Dabbous, A.N., Al-Shatti, A., Al-Khayat, A., 2018. Disability adjusted life  
531 years (Dalys) in terms of years of life lost (yll) due to premature adult mortalities and  
532 postneonatal infant mortalities attributed to pm2.5 and pm10 exposures in kuwait.  
533 *International Journal of Environmental Research and Public Health* 15 (11), art. no. 609,  
534 doi:10.3390/ijerph15112609
- 535 Ai, S., Wang, C., Qian, Z.M., Cui, Y., Liu, Y., Acharya, B.K., Sun, X., Hinyard, L., Jansson, D.R., Qin, L.,  
536 Lin, H. 2019. Hourly associations between ambient air pollution and emergency ambulance  
537 calls in one central Chinese city: Implications for hourly air quality standards. *Sci. Total Environ.*  
538 696, art. no. 133956. doi: 10.1016/j.scitotenv.2019.133956
- 539 AIR QUALITY, 2019. Generalitat de Catalunya. Geoinformation. Air Quality Information. Available at:  
540 <http://dtes.gencat.cat/icqa/start.do?lang=en> (accessed 10.01.19).
- 541 Alghamdi, M.A., Khoder, M., Harrison, R.M., Hyvärinen, A.-P., Hussein, T., Al-Jeelani, H., et al., 2014.  
542 Temporal variations of O3 and NOx in the urban background atmosphere of the coastal city  
543 Jeddah, Saudi Arabia. *Atmos. Environ.* 94, 205–214. doi:10.1016/j.atmosenv.2014.03.029
- 544 Almagro, P., Calbo, E., Ochoa De Echagüen, A., Barreiro, B., Quintana, S., Heredia, J.L., Garau, J.  
545 Mortality after hospitalization for COPD (2002) *Chest*, 121 (5), pp. 1441-1448.
- 546 Beckerman, B., Jerrett, M., Brook, J.R., Verma, D.K., Arain, M.A., Finkelstein, M.M., 2008. Correlation  
547 of nitrogen dioxide with other traffic pollutants near a major expressway. *Atmos. Environ.* 42,  
548 275–290. doi:10.1016/j.atmosenv.2007.09.042
- 549 Borge, R., Requia, W.J., Yagüe, C., Jhun, I., Koutrakis, P., 2019. Impact of weather changes on air  
550 quality and related mortality in Spain over a 25 year period [1993–2017]. *Environ. Int.* 133, art.  
551 no. 105272. doi:10.1016/j.envint.2019.105272
- 552 Bordonaba, J.G., Vilavert, L., Nadal, M., Schuhmacher, M., Domingo, J.L., 2011. Monitoring

553 environmental levels of trace elements near a hazardous waste incinerator human health risks  
554 after a decade of regular operations. *Biol. Trace Elem. Res.* 144, 1419–1429.

555 Brook, R.D., 2007. Is air pollution a cause of cardiovascular disease? Updated review and  
556 controversies. *Rev. Environ. Health* 22, 115–137.

557 CAFE, 2004. Second Position Paper on Particulate Matter. Clean Air For Europe programme Working  
558 Group on Particulate Matter.

559 Chen, G., Guo, Y., Abramson, M.J., Williams, G., Li, S., 2018. Exposure to low concentrations of air  
560 pollutants and adverse birth outcomes in Brisbane, Australia, 2003–2013. *Sci. Total Environ.*  
561 622-623, 721-726. doi: 10.1016/j.scitotenv.2017.12.050

562 Chen, H., Goldberg, M.S., Viileneuve, P.J., 2008. A systematic review of the relation between long-  
563 term exposure to ambient air pollution and chronic diseases. *Rev. Environ. Health* 23, 243–  
564 297.

565 Chen, M.-J., Lin, C.-H., Lai, C.-H., Cheng, L.-H., Yang, Y.-H., Huang, L.-J., et al., 2016. Excess lifetime  
566 cancer risk assessment of volatile organic compounds emitted from a petrochemical industrial  
567 complex. *Aerosol Air Qual. Res.* 16, 1954–1966. doi:10.4209/aaqr.2015.05.0372

568 Chirlaque, M.D., Salmerón, D., Galceran, J., Ameijide, A., Mateos, A., Torrella, A., Jiménez, R.,  
569 Larrañaga, et al., 2018. Cancer survival in adult patients in Spain. Results from nine population-  
570 based cancer registries. *Clin. Transl. Oncol.* 20, 201-211.

571 CHSO, 2019. Catalan Health System Observatory. Open data. Generalitat de Catalunya. Available at:  
572 [http://observatorisalut.gencat.cat/en/indicadors\\_i\\_publicacions/publicacions/sistema\\_sanitari/activitat/activitat\\_serveis\\_sanitaris\\_publics/els\\_conjunts\\_minims\\_basics\\_de\\_dades\\_d\\_activitat\\_assistencial/dades\\_obertes/index.html](http://observatorisalut.gencat.cat/en/indicadors_i_publicacions/publicacions/sistema_sanitari/activitat/activitat_serveis_sanitaris_publics/els_conjunts_minims_basics_de_dades_d_activitat_assistencial/dades_obertes/index.html) (accessed 10.01.19).

575 Clua-Espuny, J.L., Garcés-Redondo, M., Lucas-Noll, J., Panisello-Tafalla, A., Queralt-Tomas, LL., et al.,  
576 2014. Stroke Epidemiology, Survival and Disability in A Mediterranean Population According  
577 Malmgren’s Criteria. *Ebrietus Cohort. Ann. Vasc. Med. Res.* 1, 1004.

578 Contini, D., Cesari, D., Donateo, A., Chirizzi, D., Belosi, F., 2014. Characterization Of PM<inf>10</inf>  
579 And PM<inf>2.5</inf> and their metals content in different typologies of sites in South-Eastern  
580 Italy. *Atmosphere (Basel)*. 5, 435–453. doi:10.3390/atmos5020435

581 Demetriou, C.A., Raaschou-Nielsen, O., Loft, S., Møller, P., Vermeulen, R., et al., 2012. Biomarkers  
582 of ambient air pollution and lung cancer: A systematic review. *Occup. Environ. Med.* 69, 619–  
583 627. doi:10.1136/oemed-2011-100566

584 Departament de Salut, 2019. Mortalitat (In Catalan). Generalitat de Catalunya. Available at:  
585 [http://salutweb.gencat.cat/ca/el\\_departament/estadistiques\\_sanitaries/dades\\_de\\_salut\\_i\\_serveis\\_sanitaris/mortalitat/](http://salutweb.gencat.cat/ca/el_departament/estadistiques_sanitaries/dades_de_salut_i_serveis_sanitaris/mortalitat/) (accessed 10.01.19).

587 Devleeschauwer, B., Havelaar, A.H., Maertens de Noordhout, C., Haagsma, J.A., Praet, N., et al,  
588 2014. DALY calculation in practice: a stepwise approach. *Int. J. Public Health* 59, 571–4.  
589 doi:10.1007/s00038-014-0553-y

590 Domínguez-Morueco, N., Ratola, N., Sierra, J., Nadal, M., Jiménez-Guerrero, P., 2019. Combining  
591 monitoring and modelling approaches for BaP characterization over a petrochemical area. *Sci.*  
592 *Total Environ.* 658, 424-438. doi:10.1016/j.scitotenv.2018.12.202

593 Domínguez-Morueco, N., Augusto, S., Trabalón, L., Pocurull, E., Borrull, F., Schuhmacher, M., et al.,  
594 2015. Monitoring PAHs in the petrochemical area of Tarragona County, Spain: comparing  
595 passive air samplers with lichen transplants. *Environ. Sci. Pollut. Res. Int.* 24, 11890-900.  
596 doi:10.1007/s11356-015-5612-2

597 EEA, European Environmental Agency, 2018. Air quality in Europe — 2018 report. European  
598 Environment Agency, Luxembourg. doi: 10.2800/777411

599 European Union Parliament and Council, 2008. Directive 2008/50/EC of the European Parliament  
600 and of the Council of 21 May 2008 on ambient air quality and cleaner air for Europe. *Off. J. Eur.*

601 Union L 152, 1–44.

602 European Union Parliament and Council, 2004. Directive 2004/107/EC of the European Parliament  
603 and of the Council of 15 December 2004 relating to arsenic, cadmium, mercury, nickel and  
604 polycyclic aromatic hydrocarbons in ambient air. *Off. J. Eur. Communities L 23*, 3–16.

605 European Union Parliament and Council, 2002. Directive 2002/3/EC of the European Parliament and  
606 of the Council of 12 February 2002 relating to ozone in ambient air. *Off. J. Eur. Communities*  
607 *L67*, 14–30.

608 Farhat, S.C.L., Silva, C.A., Orione, M.A.M., Campos, L.M.A., Sallum, A.M.E., Braga, A.L.F., 2011. Air  
609 pollution in autoimmune rheumatic diseases: A review. *Autoimmun. Rev.* 11, 14–21.  
610 doi:10.1016/j.autrev.2011.06.008

611 Fecht, D., Hansell, A.L., Morley, D., Dajnak, D., Vienneau, D., Beevers, S., et al., 2016. Spatial and  
612 temporal associations of road traffic noise and air pollution in London: Implications for  
613 epidemiological studies. *Environ. Int.* 88, 235–242. doi:10.1016/j.envint.2015.12.001

614 Fenech, S., Aquilina, N.J., 2020. Trends in ambient ozone, nitrogen dioxide, and particulate matter  
615 concentrations over the Maltese Islands and the corresponding health impacts. *Sci. Total*  
616 *Environ.* 700, 134527. doi: 10.1016/j.scitotenv.2019.134527 .

617 Font, A., Guiseppin, L., Blangiardo, M., Gherzi, V., Fuller, G.W., 2019. A tale of two cities: is air  
618 pollution improving in Paris and London? *Environ. Pollut.* 249, 1–12.  
619 doi:10.1016/j.envpol.2019.01.040

620 Gallego, E., Roca, F.J., Perales, J.F., Gadea, E., 2018. Outdoor air 1,3-butadiene monitoring near a  
621 petrochemical industry (Tarragona region) and in several Catalan urban areas using active  
622 multi-sorbent bed tubes and analysis through TD-GC/MS. *Sci. Total Environ.* 618, 1440–1448.

623 Han, S., Bian, H., Feng, Y., Liu, A., Li, X., Zeng, F., Zhang, X., 2011. Analysis of the relationship between  
624 O<sub>3</sub>, NO and NO<sub>2</sub> in Tianjin, China. *Aerosol Air Qual. Res.* 11, 128–139.

625 IDESCAT, 2019. Institut d’Estadística de Catalunya. Official statistics website of Catalonia. . Available  
626 at: <http://www.idescat.cat/en/> (accessed 10.01.19).

627 Jia, M., Zhao, T., Cheng, X., Gong, S., Zhang, X., et al., 2017. Inverse Relations of PM<sub>2.5</sub>  
628 and O<sub>3</sub> in air compound pollution between cold and hot seasons over an urban area  
629 of East China. *Atmosphere (Basel)*. 8, 1–12. doi:10.3390/atmos8030059

630 Jiménez, P., Baldasano, J.M., 2004. Ozone response to precursor controls in very complex terrains:  
631 Use of photochemical indicators to assess O<sub>3</sub>-NO<sub>x</sub>-VOC sensitivity in the  
632 northeastern Iberian Peninsula. *J. Geophys. Res. D Atmos.* 109, D20309 1–20.  
633 doi:10.1029/2004JD004985

634 Kampeerawipakorn, O., Navasumrit, P., Settachan, D., Promvijit, J., Hunsonti, P., Parnlob, V., et al.,  
635 2017. Health risk evaluation in a population exposed to chemical releases from a  
636 petrochemical complex in Thailand. *Environ. Res.* 152, 207–213.  
637 doi:10.1016/j.envres.2016.10.004

638 Kloog, I., Coull, B.A., Zanobetti, A., Koutrakis, P., Schwartz, J.D., 2012. Acute and chronic effects of  
639 particles on hospital admissions in New-England. *PLoS ONE* 7(4), e34664.  
640 doi:10.1371/journal.pone.0034664

641 Kumar, A., Singh, D., Singh, B.P., Singh, M., Anandam, K., Kumar, K., et al., 2015. Spatial and temporal  
642 variability of surface ozone and nitrogen oxides in urban and rural ambient air of Delhi-NCR,  
643 India. *Air Qual. Atmos. Heal.* 8, 391–399. doi:10.1007/s11869-014-0309-0

644 Kwon, J., Weisel, C.P., Morandi, M.T., Stock, T.H., 2016. Source proximity and meteorological effects  
645 on residential outdoor VOCs in urban areas: Results from the Houston and Los Angeles RIOPA  
646 studies. *Sci. Total Environ.* 573, 954–964. doi:10.1016/j.scitotenv.2016.08.186

647 Li, G., Bei, N., Tie, X., Molina, L.T., 2011. Aerosol effects on the photochemistry in Mexico City during  
648 MCMA-2006/MILAGRO campaign. *Atmos. Chem. Phys.* 11, 5169–5182. doi:10.5194/acp-11-

649 5169-2011

650 Mafrici, A., Proietti, R., Klugmann, S., 2008. Air pollution exposure as an emerging risk factor for  
651 cardiovascular disease: A literature review | L'inquinamento atmosferico quale emergente  
652 fattore di rischio per le malattie cardiovascolari: Una revisione ragionata della letteratura. *G.*  
653 *Ital. Cardiol.* 9, 90–103.

654 Mari, M., Nadal, M., Schuhmacher, M., Barbería, E., García, F., Domingo, J.L., 2014. Human exposure  
655 to metals: Levels in autopsy tissues of individuals living near a hazardous waste incinerator.  
656 *Biol. Trace Elem. Res.* 159, 15–21. doi:10.1007/s12011-014-9957-z

657 Mari, M., Nadal, M., Schuhmacher, M., Domingo, J.L., 2013. Body burden monitoring of dioxins and  
658 other organic substances in workers at a hazardous waste incinerator. *Int. J. Hyg. Environ.*  
659 *Health* 216, 728–734. doi:10.1016/j.ijheh.2013.01.003

660 Meteocat, 2019. Xarxa d'Estacions Meteorològiques Automàtiques. Anuaris Available at:  
661 [http://www.meteo.cat/wpweb/climatologia/serveis-i-dades-climatiques/anuaris-de-dades-](http://www.meteo.cat/wpweb/climatologia/serveis-i-dades-climatiques/anuaris-de-dades-meteorologiques/xarxa-destacions-meteorologiques-automatiques/)  
662 [meteorologiques/xarxa-destacions-meteorologiques-automatiques/](http://www.meteo.cat/wpweb/climatologia/serveis-i-dades-climatiques/anuaris-de-dades-meteorologiques/xarxa-destacions-meteorologiques-automatiques/) (accessed 10.01.19).

663 Miravittles, M., Soriano, J.B., García-Río, F., Muñoz, L., Duran-Tauleria, E., Sanchez, G., Sobradillo, V.,  
664 Ancochea, J., 2009. Prevalence of COPD in Spain: Impact of undiagnosed COPD on quality of  
665 life and daily life activities. *Thorax* 64, 863-868.

666 Monteiro, A., Russo, M., Gama, C., Lopes, M., Borrego, C., 2018 How economic crisis influence air  
667 quality over Portugal (Lisbon and Porto)? *Atmos. Pollut. Res.* 9, 439-445.

668 Moran, A.E., Forouzanfar, M.H., Roth, G.A., Mensah, G.A., Ezzati, M., Flaxman, A., Murray, C.J.L.,  
669 Naghavi, M., 2014. The global burden of ischemic heart disease in 1990 and 2010: The global  
670 burden of disease 2010 study. *Circulation* 129, 1493-1501.

671 Nadal, M., Domingo, J.L., García, F., Schuhmacher, M., 2009a. Levels of PCDD/F in adipose tissue on  
672 non-occupationally exposed subjects living near a hazardous waste incinerator in Catalonia,  
673 Spain. *Chemosphere* 74, 1471–1476.

674 Nadal, M., Mari, M., Schuhmacher, M., Domingo, J.L., 2009b. Multi-compartmental environmental  
675 surveillance of a petrochemical area: Levels of micropollutants. *Environ. Int.* 35, 227–235.

676 Nadal, M., Schuhmacher, M., Domingo, J.L., 2011. Long-term environmental monitoring of  
677 persistent organic pollutants and metals in a chemical/petrochemical area: Human health  
678 risks. *Environ. Pollut.* 159, 1769–1777. doi:10.1016/j.envpol.2011.04.007

679 Núñez-Alonso, D., Pérez-Arribas, L.V., Manzoor, S., Cáceres, J.O., 2019. Statistical Tools for Air  
680 Pollution Assessment: Multivariate and Spatial Analysis Studies in the Madrid Region. *J.*  
681 *Anal. Methods Chem.* 2019, 9753927. doi:10.1155/2019/9753927

682 Ostro, B., 2004. Outdoor air pollution: assessing the environmental burden of disease at national  
683 and local levels. WHO Environmental Burden of Disease Series, Geneva.

684 Pacca, L., Antonarakis, A., Schröder, P., Antoniadis, A., 2020. The effect of financial crises on air  
685 pollutant emissions: An assessment of the short vs. medium-term effects *Sci. Total Environ.*  
686 98, 133614

687 Prüss-Ustün, A., Wolf, J., Cordolán, C., Bos, R., Neira, M., 2016. Preventing disease through healthy  
688 environments: a global assessment of the burden of disease from environmental risks. World  
689 Health Organization, Geneva.

690 Querol, X., Alastuey, A., Moreno, T., Viana, M.M., Castillo, S., Pey, J., et al., 2008. Spatial and  
691 temporal variations in airborne particulate matter (PM10 and PM2.5) across Spain 1999–2005.  
692 *Fifth Int. Conf. Urban Air Qual.* 42, 3964–3979. doi:DOI: 10.1016/j.atmosenv.2006.10.071

693 Querol, X., Alastuey, A., Ruiz, C.R., Artiñano, B., Hansson, H.C., Harrison, R.M., et al., 2004. Speciation  
694 and origin of PM10 and PM2.5 in selected European cities. *Atmos. Environ.* 38, 6547–6555.  
695 doi:10.1016/j.atmosenv.2004.08.037

696 Raaschou-Nielsen, O., Andersen, Z.J., Beelen, R., Samoli, E., Stafoggia, M., Weinmayr, G., et al., 2013.

697 Air pollution and lung cancer incidence in 17 European cohorts: Prospective analyses from the  
698 European Study of Cohorts for Air Pollution Effects (ESCAPE). *Lancet Oncol.* 14, 813–822.  
699 doi:10.1016/S1470-2045(13)70279-1

700 Raaschou-Nielsen, O., Reynolds, P., 2006. Air pollution and childhood cancer: A review of the  
701 epidemiological literature. *Int. J. Cancer* 118, 2920–2929. doi:10.1002/ijc.21787

702 Ramírez, N., Cuadras, A., Rovira, E., Borrull, F., Marcé, R.M., 2012. Chronic risk assessment of  
703 exposure to volatile organic compounds in the atmosphere near the largest Mediterranean  
704 industrial site. *Environ. Int.* 39, 200–209. doi:10.1016/j.envint.2011.11.002

705 Ramírez, N., Cuadras, A., Rovira, E., Marcé, R.M., Borrull, F., 2011. Risk assessment related to  
706 atmospheric polycyclic aromatic hydrocarbons in gas and particle phases near industrial sites.  
707 *Environ. Health Perspect.* 119, 1110–1116. doi:10.1289/ehp.1002855

708 Ras, M.R., Marcé, R.M., Cuadras, A., Mari, M., Nadal, M., Borrull, F., 2009. Atmospheric levels of  
709 polycyclic aromatic hydrocarbons in gas and particulate phases from Tarragona Region (NE  
710 Spain). *Int. J. Environ. Anal. Chem.* 89, 543–556. doi:10.1080/03067310802610276

711 Real Decreto 102, 2011, de 28 de enero, relativo a la mejora de la calidad del aire. Boletín oficial del  
712 Estado 25, 9574-9624. Available at: <https://www.boe.es/boe/dias/2011/01/29/pdfs/BOE-A-2011-1645.pdf> (accessed 1.05.19).

714 Reis, S., Seto, E., Northcross, A., Quinn, N.W.T., Convertino, M., Jones, R.L., Maier, H.R., Schlink, U.,  
715 Steinle, S., Vieno, M., Wimberly, M.C., 2015. Integrating modelling and smart sensors for  
716 environmental and human health *Environmental Modelling and Software*, 74, 238-246.  
717 doi:10.1016/j.envsoft.2015.06.003

718 Salmerón, D., Chirlaque, M.D., Izarzugaza, M.I., Sánchez, M.J., Marcos-Gragera, R., Ardanaz, E.,  
719 Galceran, J., Mateos, A., Navarro, C., 2012. Lung cancer prognosis in Spain: The role of  
720 histology, age and sex. *Respir. Med.* 106, 1301-1308.

721 Schneider, J., Nagl, C., Read, B., 2014. EU Air Quality Policy and WHO Guideline Values for Health.  
722 European Parliament's Committee on Environment, Public Health and Food Safety, 1-68. ISBN:  
723 978-92-823-6017-0 doi:10.2861/70034

724 Sicard, P., Khaniabadi, Y.O., Perez, S., Gualtieri, M., De Marco, A., 2019. Effect of O3, PM10 and  
725 PM2.5 on cardiovascular and respiratory diseases in cities of France, Iran and Italy.  
726 *Environmental Science and Pollution Research* In Press. doi: 10.1007/s11356-019-06445-8

727 Song, Q., Christiani, D.C., XiaorongWang, Ren, J., 2014. The global contribution of outdoor air  
728 pollution to the incidence, prevalence, mortality and hospital admission for chronic obstructive  
729 pulmonary disease: a systematic review and meta-analysis. *Int. J. Environ. Res. Public Health*  
730 11, 11822–11832. doi:10.3390/ijerph11111822

731 Squizzato, S., Masiol, M., Rich, D.Q., Hopke, P.K., 2019. PM2.5 and gaseous  
732 pollutants in New York State during 2005–2016: Spatial variability,  
733 temporal trends, and economic influences *Atmospheric Environ.* 183, 209–  
734 224.

735 Stanaway, J.D., Afshin, A., Gakidou, E., Lim, S.S., Abate, D., Abate, K.H., Abbafati, C., et al. 2018.  
736 Global, regional, and national comparative risk assessment of 84 behavioural, environmental  
737 and occupational, and metabolic risks or clusters of risks for 195 countries and territories,  
738 1990-2017: A systematic analysis for the Global Burden of Disease Study 2017. *The Lancet*, 392  
739 (10159), 1923-1994. doi: 10.1016/S0140-6736(18)32225-6

740 Sun, G., Hazlewood, G., Bernatsky, S., Kaplan, G.G., Eksteen, B., Barnabe, C., 2016. Association  
741 between Air Pollution and the Development of Rheumatic Disease: A Systematic Review. *Int.*  
742 *J. Rheumatol.* 2016, 5356307. doi:10.1155/2016/5356307

743 Sunyer, J., 2001. Urban air pollution and chronic obstructive pulmonary disease: A review. *Eur.*  
744 *Respir. J.* 17, 1024–1033. doi:10.1183/09031936.01.17510240

745 Taylor, C.J., Ordóñez-Mena, J.M., Roalfe, A.K., Lay-Flurrie, S., Jones, N.R., Marshall, T., Hobbs, F.D.R.,

746 2019. Trends in survival after a diagnosis of heart failure in the United Kingdom 2000-2017:  
747 population based cohort study. *BMJ (Online)* 364, 1223.

748 Tong, D., Pan, L., Chen, W., Lamsal, L., Lee, P., Tang, Y., Kim, H., Kondragunta, S., Stajner, I., 2016.  
749 Impact of the 2008 Global Recession on air quality over the United States: Implications for  
750 surface ozone levels from changes in NO<sub>x</sub> emissions. *Geophys. Res. Lett.* 43, 9280-9288.

751 TQACT, 2019. Taula de Qualitat de l'Aire al Camp de Tarragona (In Catalan). Available at: /  
752 (accessed 10.01.19).

753 Tzima, K., Analitis, A., Katsouyanni, K., Samoli, E., 2018. Has the risk of mortality related to short-  
754 term exposure to particles changed over the past years in Athens, Greece? *Environ. Int.* 113,  
755 306-312.

756 Vilavert, L., Nadal, M., Schuhmacher, M., Domingo, J.L., 2015. Two Decades of Environmental  
757 Surveillance in the Vicinity of a Waste Incinerator: Human Health Risks Associated with Metals  
758 and PCDD/Fs. *Arch. Environ. Contam. Toxicol.* 69, 241–253. doi:10.1007/s00244-015-0168-1

759 Vilavert, L., Nadal, M., Schuhmacher, M., Domingo, J.L., 2014. Seasonal surveillance of airborne  
760 PCDD/Fs, PCBs and PCNs using passive samplers to assess human health risks. *Sci. Total  
761 Environ.* 466-467, 733–40. doi:10.1016/j.scitotenv.2013.07.124

762 Vilavert, L., Nadal, M., Schuhmacher, M., Domingo, J.L., 2012. Long-term monitoring of dioxins and  
763 furans near a municipal solid waste incinerator: Human health risks. *Waste Manag. Res.* 30,  
764 908–916.

765 Vilavert, L., Nadal, M., Mari, M., Schuhmacher, M., Domingo, J.L., 2010. Monitoring temporal trends  
766 in environmental levels of polychlorinated dibenzo-p-dioxins and dibenzofurans: Results from  
767 a 10-year surveillance program of a hazardous waste incinerator. *Arch. Environ. Contam.  
768 Toxicol.* 59, 521–531.

769 Visschedijk, A.H.J., Denier Van Der Gon, H.A.C., Hulskotte, J.H.J., Quass, U., 2013. Anthropogenic  
770 Vanadium emissions to air and ambient air concentrations in North-West Europe. *E3S Web of  
771 Conferences*, 1,03004

772 Wang, W., Liu, C., Ying, Z., Lei, X., Wang, C., Huo, J., Zhao, Q., Zhang, Y., Duan, Y., Chen, R., Fu, Q.,  
773 Zhang, H., Kan, H., 2019. Particulate air pollution and ischemic stroke hospitalization: How the  
774 associations vary by constituents in Shanghai, China. *Sci. Total Environ.* 695, art. no. 133780.  
775 doi: 10.1016/j.scitotenv.2019.133780.

776 WHO, 2019. World Health Organization. Health statistics and information systems. Metrics:  
777 Disability-Adjusted Life year (DALY). Available at:  
778 [http://www.who.int/healthinfo/global\\_burden\\_disease/metrics\\_daly/en/](http://www.who.int/healthinfo/global_burden_disease/metrics_daly/en/)  
779 (accessed 10.01.19).

780 WHO, 2016a. World health statistics 2016: monitoring health for the SDGs, sustainable  
781 development goals. World Health Organization.

782 WHO, 2016b. Health risk assessment of air pollution – general principles. WHO Regional Office for  
783 Europe, Copenhagen.

784 WHO, 2016c. Ambient air pollution: A global assessment of exposure and burden of disease. World  
785 Health Organization Press, Geneva.

786 WHO, 2013. Health risks of air pollution in Europe –HRAPIE project: recommendations for  
787 concentration–response functions for cost–benefit analysis of particulate matter, ozone and  
788 nitrogen dioxide. WHO Regional Office for Europe.

789 WHO, 2006. WHO Air quality guidelines for particulate matter, ozone, nitrogen dioxide and sulfur  
790 dioxide. Global update 2005. Summary of risk assessment. World Health Organization Press,  
791 Geneva.

792 Wu, Z., Zhang, Y., Zhang, L., Huang, M., Zhong, L., Chen, D., Wang, X., 2019. Trends of outdoor air  
793 pollution and the impact on premature mortality in the Pearl River Delta region of southern

794 China during 2006–2015. *Sci. Total Environ.* 690, 248-260. doi:10.1016/j.scitotenv.2019.06.401  
795 XVPCA, 2019. Generalitat de Catalunya. Departament de Territori i Sostenibilitat. Avaluació de la  
796 Qualitat de l'aire (In Catalan). Available at:  
797 [http://mediambient.gencat.cat/ca/05\\_ambits\\_dactuacio/atmosfera/qualitat\\_de\\_laيرة/avalua](http://mediambient.gencat.cat/ca/05_ambits_dactuacio/atmosfera/qualitat_de_laيرة/avaluacio/index.html)  
798 [cio/index.html](http://mediambient.gencat.cat/ca/05_ambits_dactuacio/atmosfera/qualitat_de_laيرة/avaluacio/index.html) (accessed 10.01.19).  
799 Zhang, H., Dong, H., Ren, M., Liang, Q., Shen, X., Wang, Q., Yu, L., Lin, H., Luo, Q., Chen, W., Knibbs,  
800 L.D., Jalaludin, B., Wang, Q., Huang, C., 2020. Ambient air pollution exposure and gestational  
801 diabetes mellitus in Guangzhou, China: A prospective cohort study. *Sci. Total Environ.* 699, art.  
802 no. 134390. doi: 10.1016/j.scitotenv.2019.134390  
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807 Table1. Air quality monitoring stations in the Camp de Tarragona

| Station       | Coordinates<br>UTM 31 North | Altitude<br>(m) | Pollutants analysed  | Area          |
|---------------|-----------------------------|-----------------|--|---------------|
| Sant Salvador | 352405, 4558155             | 57              | SO <sub>2</sub> , NO, NO <sub>2</sub> , H <sub>2</sub> S   | North complex |
| Perafort      | 352230, 4561950             | 97              | SO <sub>2</sub> , NO, NO <sub>2</sub> , H <sub>2</sub> S, Bz   | Rural         |
| Bonavista     | 348300, 4553400             | 39              | SO <sub>2</sub> , NO, NO <sub>2</sub> , H <sub>2</sub> S, PM <sub>10</sub>   | South complex |
| Constantí     | 350550, 4557690             | 56              | SO <sub>2</sub> , NO, NO <sub>2</sub> , H <sub>2</sub> S, O <sub>3</sub> , Bz,<br>PM <sub>10</sub> , PM <sub>2.5</sub> , metals, B[a]p | North complex |
| Vilaseca      | 344920, 4553050             | 41              | SO <sub>2</sub> , NO, NO <sub>2</sub> , H <sub>2</sub> S, O <sub>3</sub> , PM <sub>10</sub> ,<br>metals                                | South complex |
| Reus          | 342355, 4557402             | 102             | SO <sub>2</sub> <sup>a</sup> , NO, NO <sub>2</sub> , H <sub>2</sub> S, O <sub>3</sub> , PM <sub>10</sub> ,<br>CO, metals <sup>b</sup>  | Urban         |
| Tarragona     | 352474, 4553482             | 13              | SO <sub>2</sub> , NO, NO <sub>2</sub> , H <sub>2</sub> S, O <sub>3</sub> , Bz,<br>PM <sub>10</sub> , CO                                | Urban         |
| Laboral       | 349010, 4552030             | 5               | SO <sub>2</sub> , NO, NO <sub>2</sub> , H <sub>2</sub> S, Bz, PM <sub>10</sub> ,<br>PM <sub>2.5</sub>                                  | South complex |
| Alcover       | 347665, 4571462             | 238             | SO <sub>2</sub> , NO, NO <sub>2</sub> , H <sub>2</sub> S, O <sub>3</sub> , CO  | Rural         |

Bz: Benzene; B[a]p: Benzo(a)pyrene, Metals: As, Cd, Ni, and Pb.  
<sup>a</sup> Data available from 2005 to 2012; <sup>b</sup> Data available from 2005 to 2015

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809 Table 2. Yearly means (standard deviations) of pollutants in the Camp de Tarragona from 2005 to  
 810 2017 (AIRQUALITY, 2016) and European air quality standards (European Union Parliament and  
 811 Council, 2008) and WHO guidelines (WHO, 2006).

|                   |                | NO <sub>2</sub>  | NO        | SO <sub>2</sub> | PM <sub>10</sub> | PM <sub>2.5</sub> | O <sub>3</sub>   | H <sub>2</sub> S | CO              | Bz        |
|-------------------|----------------|------------------|-----------|-----------------|------------------|-------------------|------------------|------------------|-----------------|-----------|
| <b>2005</b>       |                | 23.5 (6.5)       | 7.3 (3.7) | 4.9 (1.7)       | 46.5 (7.8)       | NA                | 53.6 (6.4)       | 1.4 (0.5)        | 0.4 (0.0)       | 1.1 (0.1) |
| <b>2006</b>       |                | 21.0 (5.7)       | 7.0 (2.5) | 3.7 (1.0)       | 38.5 (3.5)       | NA                | 55.0 (7.3)       | 1.6 (0.4)        | 0.3 (0.1)       | 1.5 (0.4) |
| <b>2007</b>       |                | 21.6 (6.6)       | 6.5 (3.1) | 4.7 (1.7)       | 32.5 (2.1)       | NA                | 54.6 (7.7)       | 1.7 (0.6)        | 0.3 (0.1)       | 1.4 (0.6) |
| <b>2008</b>       |                | 20.1 (5.6)       | 6.4 (2.6) | 3.2 (1.2)       | 29.3 (3.1)       | 10.7 (2.4)        | 53.8 (2.5)       | 1.6 (0.4)        | 0.3 (0.1)       | 1.3 (0.5) |
| <b>2009</b>       |                | 21.6 (5.8)       | 5.8 (2.6) | 4.0 (1.7)       | 27.4 (1.2)       | 12.5 (0.7)        | 55.9 (5.9)       | 1.7 (0.8)        | 0.3 (0.1)       | 1.1 (0.4) |
| <b>2010</b>       |                | 21.3 (6.1)       | 5.3 (2.4) | 3.1 (1.5)       | 22.0 (1.8)       | 12.2 (2.5)        | 58.7 (6.2)       | 1.6 (0.3)        | 0.3 (0.1)       | 1.1 (0.6) |
| <b>2011</b>       |                | 20.2 (5.2)       | 5.4 (2.5) | 3.7 (1.5)       | 25.5 (1.3)       | 13.9 (1.9)        | 57.1 (7.0)       | 1.4 (0.3)        | 0.3 (0.1)       | 1.3 (0.4) |
| <b>2012</b>       |                | 19.1 (5.6)       | 5.1 (2.2) | 2.4 (0.8)       | 22.4 (2.6)       | 14.0 (2.5)        | 58.9 (6.6)       | 1.4 (0.3)        | 0.3 (0.0)       | 1.0 (0.4) |
| <b>2013</b>       |                | 17.8 (4.6)       | 4.9 (2.0) | 2.2 (0.5)       | 16.9 (2.3)       | 10.8 (1.6)        | 61.7 (6.8)       | 1.4 (0.3)        | 0.3 (0.0)       | 1.2 (0.5) |
| <b>2014</b>       |                | 18.0 (4.3)       | 5.0 (1.7) | 2.3 (0.5)       | 19.4 (3.1)       | 11.3 (0.5)        | 58.0 (6.7)       | 1.4 (0.3)        | 0.2 (0.0)       | 1.2 (0.4) |
| <b>2015</b>       |                | 19.5 (4.9)       | 5.7 (2.0) | 2.8 (1.1)       | 23.4 (2.0)       | 14.7 (1.2)        | 58.3 (7.0)       | 1.5 (0.3)        | 0.2 (0.0)       | 1.2 (0.4) |
| <b>2016</b>       |                | 17.9 (4.7)       | 5.3 (1.8) | 3.0 (1.1)       | 19.1 (2.8)       | 10.8 (0.6)        | 57.5 (6.8)       | 1.6 (0.3)        | 0.3 (0.0)       | 1.1 (0.2) |
| <b>2017</b>       |                | 18.3 (5.1)       | 4.9 (1.7) | 2.8 (0.9)       | 20.7 (1.4)       | 11.8 (0.9)        | 59.8 (6.8)       | 1.7 (0.3)        | 0.2 (0.0)       | 1.2 (0.4) |
| <b>WHO</b>        | <b>annual</b>  | 40               | -         | -               | 20               | 10                | -                | -                | -               | -         |
| <b>guidelines</b> | <b>24-hour</b> | 200 <sup>a</sup> | -         | 20              | 50               | 25                | 100 <sup>b</sup> | -                | 10 <sup>b</sup> | -         |
| <b>EU</b>         | <b>annual</b>  | 40               | -         | -               | 40               | 25                | -                | -                | -               | 5         |
| <b>standards</b>  | <b>24-hour</b> | 200 <sup>a</sup> | -         | 125             | 50               | -                 | 120 <sup>b</sup> | -                | 10 <sup>b</sup> | -         |

Bz: Benzene; NA: Not Available. <sup>a</sup> 1 hour mean; <sup>b</sup> 8 hour-mean. Levels of NO<sub>2</sub>, NO, SO<sub>2</sub>, PM<sub>10</sub>, PM<sub>2.5</sub>, O<sub>3</sub>, H<sub>2</sub>S and benzene in µg/m<sup>3</sup>. Levels of CO in mg/m<sup>3</sup>.

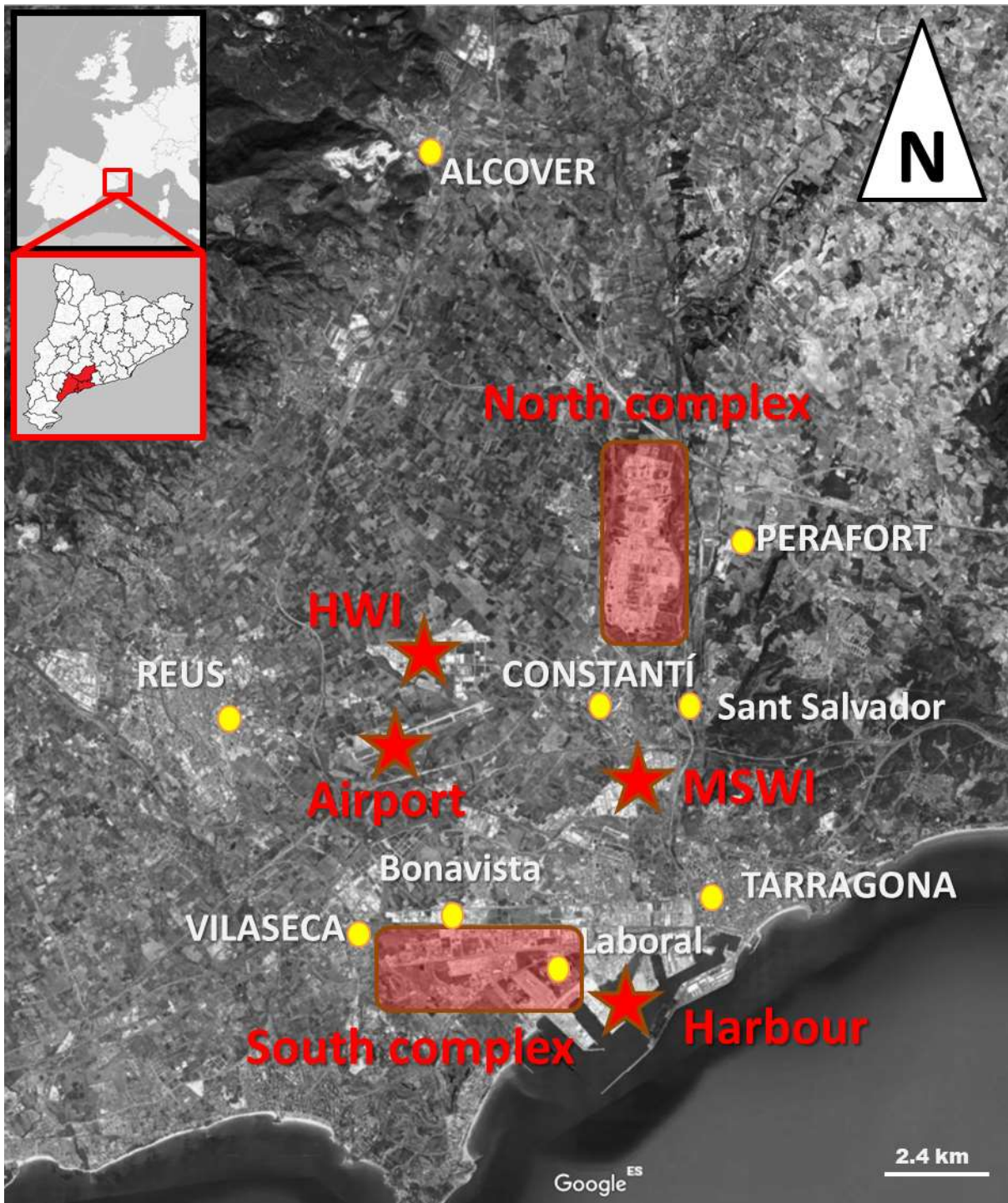
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813 Table 3. Pearson's correlations of annual mean pollutants analysed in all stations of air quality  
 814 network.

|                         | <b>NO<sub>2</sub></b> | <b>NO</b> | <b>SO<sub>2</sub></b> | <b>PM<sub>10</sub></b> | <b>PM<sub>2.5</sub></b> | <b>O<sub>3</sub></b> | <b>H<sub>2</sub>S</b> | <b>CO</b> | <b>Bz</b> |
|-------------------------|-----------------------|-----------|-----------------------|------------------------|-------------------------|----------------------|-----------------------|-----------|-----------|
| <b>NO<sub>2</sub></b>   | 1.000                 | 0.820**   | -0.132                | 0.551**                | 0.063                   | -0.788**             | 0.128                 | 0.289*    | 0.169     |
| <b>NO</b>               |                       | 1.000     | -0.034                | 0.768**                | -0.024                  | -0.616**             | -0.053                | 0.395*    | -0.038    |
| <b>SO<sub>2</sub></b>   |                       |           | 1.000                 | 0.281                  | -0.112                  | 0.084                | 0.171                 | 0.289     | 0.257     |
| <b>PM<sub>10</sub></b>  |                       |           |                       | 1.000                  | 0.431*                  | -0.340**             | 0.040                 | -0.118    | 0.395     |
| <b>PM<sub>2.5</sub></b> |                       |           |                       |                        | 1.000                   | 0.130                | -0.018                | 0.024     | -0.114    |
| <b>O<sub>3</sub></b>    |                       |           |                       |                        |                         | 1.000                | -0.264*               | -0.477**  | -0.033    |
| <b>H<sub>2</sub>S</b>   |                       |           |                       |                        |                         |                      | 1.000                 | -0.073    | 0.343*    |
| <b>CO</b>               |                       |           |                       |                        |                         |                      |                       | 1.000     | -0.187    |
| <b>Benzene</b>          |                       |           |                       |                        |                         |                      |                       |           | 1.000     |

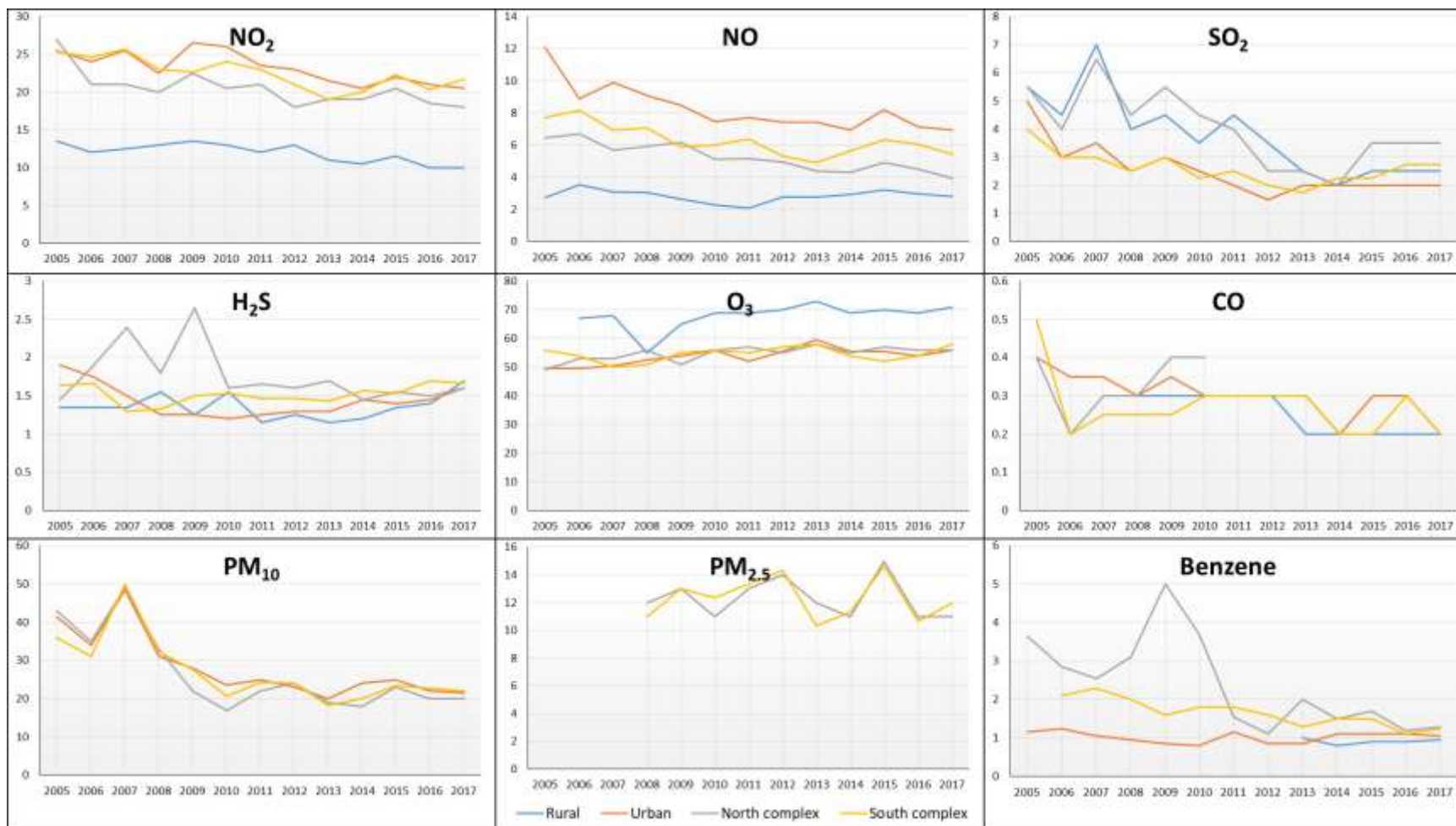
Bz: Benzene; \* Significant correlation at p<0.05, \*\*Significant correlation at p<0.01

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**Fig. 1.** Camp de Tarragona map with emission sources marked in red and air quality network stations marked in yellow circles. Principal municipalities are in capital letters.

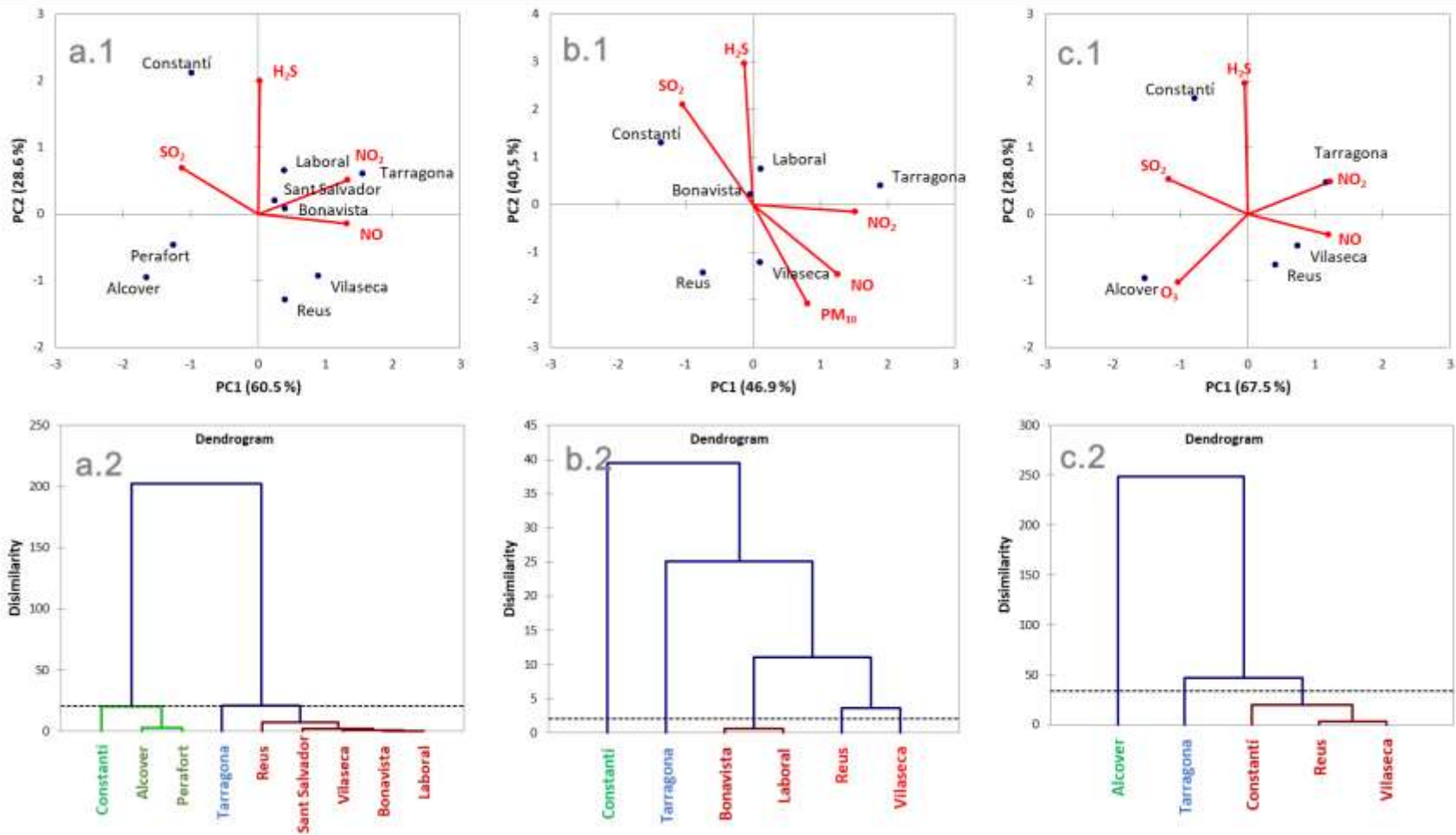


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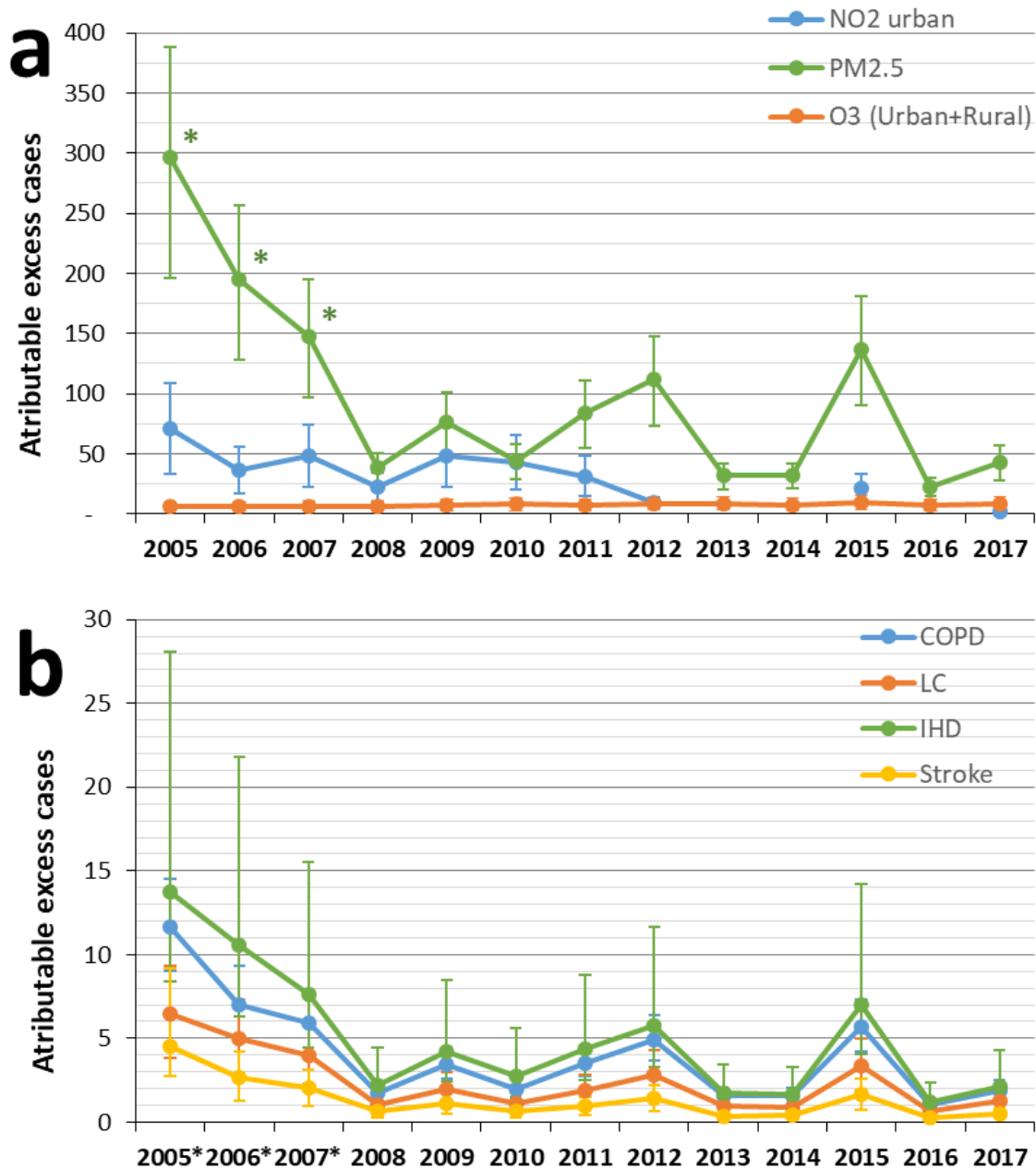
821

**Fig. 2.** Mean annual levels between 2005 and 2017 in the Camp de Tarragona of air pollutants measured in 9 stations clustered in four areas (rural, urban, north and south industrial complex). Levels of SO<sub>2</sub>, NO, NO<sub>2</sub>, O<sub>3</sub>, H<sub>2</sub>S, Benzene, PM<sub>10</sub> and PM<sub>2.5</sub> in µg/m<sup>3</sup>. Levels of CO in mg/m<sup>3</sup>.



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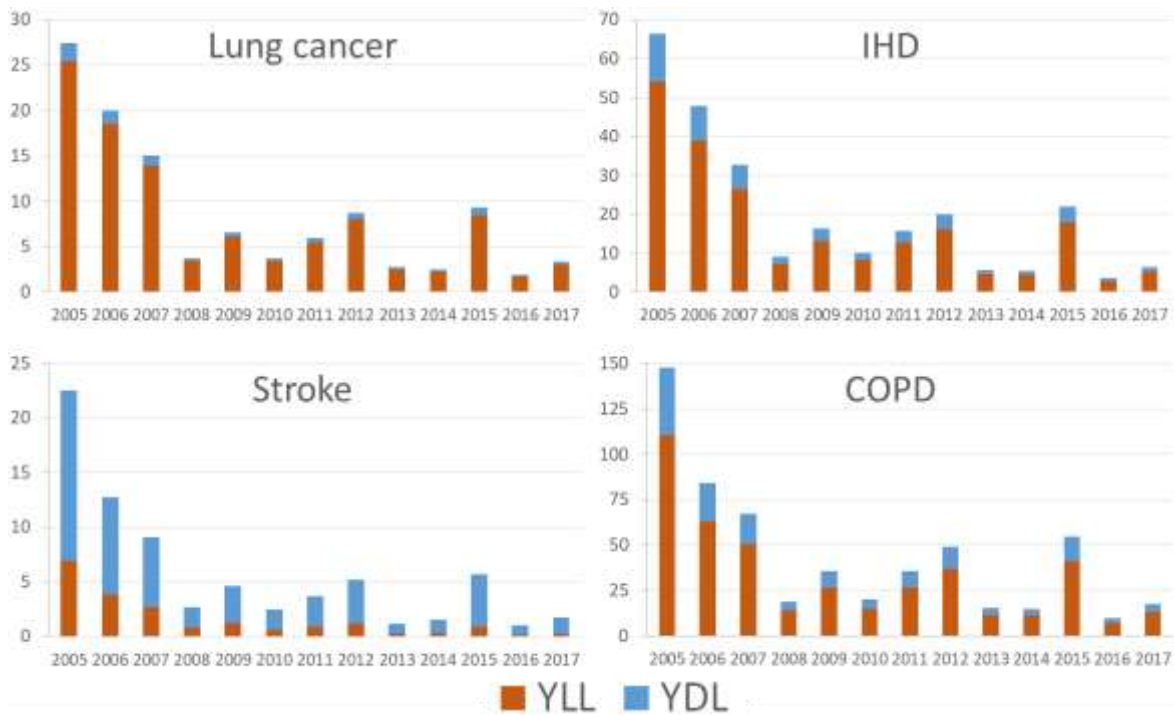
823 **Fig. 3.** Principal components analysis (1) and dendrograms (2) of annual mean pollutant levels in stations of the Camp de Tarragona. As not the  
 824 same pollutants were measured in each station, different pollutants were included in different graphs:  $\text{NO}_2$ ,  $\text{NO}$ ,  $\text{SO}_2$  and  $\text{H}_2\text{S}$  in (a);  $\text{NO}_2$ ,  $\text{NO}$ ,  $\text{SO}_2$ ,  
 825  $\text{H}_2\text{S}$  and  $\text{PM}_{10}$  in (b), and  $\text{NO}$ ,  $\text{NO}_2$ ,  $\text{SO}_2$ ,  $\text{H}_2\text{S}$  and  $\text{O}_3$  in (c).



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827 **Fig. 4.** Attributable excess cases of mortality due to: a) all natural causes and b) chronic obstructive  
 828 pulmonary disease (COPD), lung cancer (LC), ischemic heart disease (IHD) and stroke, attributable  
 829 to air pollution above the WHO threshold scenario in the Camp de Tarragona (2005-2017). Points  
 830 and error bars are central values and 95% confidence intervals, respectively. \* Indicates that PM<sub>2.5</sub>  
 831 levels from 2005 to 2007 were calculated from PM<sub>10</sub>.

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834 **Fig. 5.** Disability-Adjusted Life Years (YLDs plus YLLs), per 100,000 inhabitants for lung cancer,  
 835 ischemic heart disease (IHD), stroke, and chronic obstructive pulmonary disease (COPD) due to PM<sub>2.5</sub>  
 836 levels above the WHO threshold limits in the Camp de Tarragona during the 2005-2017 period.