

1 **PREPARATION OF A POLAR MONOLITHIC COATING FOR STIR BAR**
2 **SORPTIVE EXTRACTION OF EMERGING CONTAMINANTS FROM**
3 **ENVIRONMENTAL WATERS**

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34 **Abstract**

35

36 A new polar monolith based on poly(poly(ethylene glycol) methacrylate-co-
37 pentaerythritol triacrylate) (poly(PEGMA-co-PETRA)) was first synthesised, after the
38 optimisation of the polymerisation conditions, and applied as a coating for the stir bar
39 sorptive extraction (SBSE) of a group of pharmaceuticals and personal care
40 products (PPCPs) from environmental water samples.

41

42 Several parameters affecting extraction and liquid desorption in SBSE were
43 investigated to achieve the optimal sorption efficiencies for the studied analytes.
44 Under the optimised experimental conditions, a rapid, simple and sensitive SBSE
45 performance was provided by the in-house monolithic stir bar. Moreover, the in-
46 house coating was able to extract and desorb most of the studied analytes more
47 effectively and quickly, due to its polar behaviour and suitable mechanical and
48 physical properties, in comparison with the recently commercialised polar stir bars
49 (EG Silicone Twister[®] and Acrylate Twister[®]).

50

51 The analytical methodology, including SBSE followed by liquid chromatography
52 coupled to tandem mass spectrometry (LC-MS/MS), was validated and successfully
53 applied for the determination of a group of PPCPs in wastewater samples.

54 **1. Introduction**

55

56 Over the last few years, stir bar sorptive extraction (SBSE) has gained popularity as
57 an extraction technique for the determination of emerging organic contaminants
58 (EOCs) in complex matrices due to its environmentally friendly performance,
59 simplicity and sensitivity [1-3].

60

61 Although SBSE has successfully been applied for the determination of organic
62 contaminants in environmental, biological and food matrices at trace levels prior to
63 a chromatographic technique [4-7], its main drawback is the availability of
64 commercial coatings. Until very recently, polydimethylsiloxane (PDMS) was the only
65 commercially available coating for SBSE, known by the name PDMS Twister[®] from
66 Gerstel, being suitable for the extraction of those analytes with more apolar
67 characteristics (i.e. $\log K_{o/w} > 3$) [3,8]. In order to overcome the limitations of PDMS
68 for the extraction of more polar compounds, very recently, two new commercial
69 coatings have been introduced by Gerstel, known as EG Silicone Twister[®] and
70 Acrylate Twister[®]. Both are PDMS-based, the former modified with poly(ethylene)
71 glycol (PEG), and the latter with polyacrylate (PA). Few studies have reported the
72 use of these new coatings for SBSE to determine volatile organic compounds
73 (VOCs) in cosmetic samples [9], benzothiazole [10] and pharmaceuticals and
74 personal care products (PPCPs) [11] in wastewaters. All of these studies
75 demonstrated a certain improvement on the extraction efficiencies of the most polar
76 analytes in comparison with the classical PDMS coating. Nevertheless, new
77 approaches in the development of SBSE coatings with polar functionalities are
78 required to improve the extraction of more polar analytes and to extend the range of
79 SBSE applications.

80

81 Up to now, many efforts have focused on the preparation of novel polar coatings for
82 SBSE and several synthetic strategies have been developed. The first of these was
83 sol-gel technology, which allowed novel materials to be obtained based on PDMS
84 modified with β -cyclodextrin to determine estrogens and bisphenol A in water

85 samples [12] or based on PDMS modified with poly(vinylalcohol) to extract
86 organophosphorus pesticides in honey [13], among others. All of these coatings,
87 chemically attached to a glass bar, were shown to be more thermally and
88 chemically stable as well as more effective in terms of extraction efficiencies of both
89 polar and apolar target analytes than with PDMS coating [1]. However, the sol-gel
90 process involved more laborious steps during the polymerisation than other
91 strategies, such as the preparation of monoliths.

92

93 Along the same lines, the second approach was the development of monolithic
94 materials, which have become popular in recent years due to their high permeability
95 and simplicity in preparation. For these reasons, a number of monolithic coatings
96 with different chemical behaviours have been synthesised for SBSE applications.
97 Some examples of this approach consisted of vinylphthalimide (VPA) and *N,N'*-
98 methylenebisacrylamide (MBAA) to determine benzimidazoles in milk and honey
99 [14] or, alternatively, vinylpyrrolidone (VPD) and divinylbenzene (DVB) to extract
100 PPCPs from environmental waters [15], which enabled the effective extraction of
101 polar compounds that were not retained onto the PDMS coating. In addition, this
102 simple strategy has enabled a porous structure with high surface areas to be
103 developed which results in an enhancement of extraction efficiencies [1,16].

104

105 Monolithic materials are highly versatile in comparison with the classic PDMS stir
106 bar because different interactions can be achieved depending on the monomers
107 used during their syntheses. To the best of our knowledge, there is no report that
108 has previously used poly(ethylene glycol) methacrylate (PEGMA) as a functional
109 monomer and pentaerythritol triacrylate (PETRA) as a cross-linker for the
110 preparation of a monolith. The polar nature of both monomers might significantly
111 increase the polarity of the final monolith. Therefore, the aim of this paper is the
112 development of a polar monolithic coating for SBSE by the copolymerisation of
113 PEGMA and PETRA monomers. Furthermore, its SBSE performance, followed by
114 liquid chromatography-mass spectrometry in tandem (LC-MS/MS), was applied to
115 extract PPCPs from wastewater samples.

116

117 **2. Materials and methods**

118

119 **2.1. Chemicals and reagents**

120

121 For the synthesis of the monolithic coating, PETRA (technical grade) and PEGMA
122 ($M_n \sim 526$) as monomers, and cyclohexanol (99%) and methanol (MeOH) (99.7%) as
123 porogens were purchased from Sigma-Aldrich (Steinheim, Germany); 2-2'-
124 azobis(isobutyronitrile) (AIBN) as initiator was supplied by BDH (Poole, UK). The
125 monomers were purified by passing them through a short column filled with neutral
126 alumina and the initiator was recrystallised at a low temperature from MeOH prior to
127 use.

128

129 For the evaluation of the polar sorptive material, paracetamol, caffeine, antipyrine,
130 benzotriazole, propranolol hydrochloride, pridinol methanesulfonate salt,
131 methylparaben, carbamazepine, propylparaben, 2,4-dihydroxybenzophenone
132 (DHB), benzylparaben, 2,2-dihydroxy-4-methoxybenzophenone (DHMB),
133 diclofenac, benzophenone-3 (BP-3), triclocarban and triclosan were purchased from
134 Sigma-Aldrich (Steinheim, Germany) (>97%). The structures, pK_a and $\log K_{o/w}$
135 values of these analytes are presented in Fig. 1. Individual standard solutions of
136 1000 mg L^{-1} of each compound were prepared in MeOH and a standard mixture
137 solution of 10 mg L^{-1} was prepared weekly in MeOH. The standard mixture solution
138 was diluted daily with ultrapure water to give the required concentration and stored
139 at 4°C .

140

141 For LC analyses, HPLC grade MeOH and acetonitrile (ACN) were supplied by
142 Prolabo (Llinars del Vallès, Spain); formic acid (HCOOH) ($\geq 95\%$), sodium chloride
143 (NaCl) (99%) and sodium hydroxide (NaOH) ($\geq 98\%$) were purchased from Sigma-
144 Aldrich (Steinheim, Germany). Ultrapure water was obtained from a water
145 purification system (Veolia, Sant Cugat del Vallès, Spain).

146

147 **2.2. Stir bar preparation**

148

149 The detailed preparation of the poly(PEGMA-co-PETRA) monolithic coating is
150 described in our previous study [17]. The ratios of the reagents and the
151 polymerisation conditions were as follows: a ratio between monomers of 50/50 (%
152 w/w), a ratio between total monomers and porogen of 40/60 (% w/w), a porogenic
153 solvent consisted of 50% (w/w) cyclohexanol and 50% (w/w) MeOH, AIBN (1 mol%
154 relative to polymerisable double bonds), polymerised in a water bath at 60°C for 24
155 h. After the polymerisation, the monolithic stir bar was washed with MeOH, being
156 stirred overnight. The final poly(PEGMA-co-PETRA) had the following dimensions:
157 12 mm in length and a polymer thickness of 1 mm, which corresponds to a polymer
158 volume of 225 μL .

159

160 **2.3. LC-(ESI)-MS/MS analysis**

161

162 The extracts were analysed with an Agilent 1200 series LC coupled to a 6410
163 series triple quadrupole mass spectrometer with an electrospray ionisation (ESI)
164 interface, an automatic injector (volume injected was 50 μL), a degasser, a
165 quaternary pump and a column oven from Agilent Technologies (Waldbronn,
166 Germany).

167

168 The chromatographic column was a Kromasil 100 C₁₈ (150 x 4.6 mm i.d.) with 5 μm
169 particle size from Teknokroma (Barcelona, Spain). ACN and ultrapure water
170 adjusted to pH 3 with HCOOH were used as the mobile phase. The gradient started
171 increasing from 15% to 55% ACN in 12 min (held for 5 min), then to 95% ACN in 8
172 min and kept constant for 3 min. Finally, it decreased back to the initial conditions in
173 2 min. The chromatographic analyses were performed at 45°C with a flow rate of
174 0.6 mL min⁻¹.

175

176 A flow injection of a standard solution of each compound was used to find the
177 optimal conditions in the ESI source. These conditions were as follows: nebuliser

178 pressure of 45 psi, drying gas (N₂), flow rate of 12 L min⁻¹, source temperature of
179 350°C and a capillary potential of 4000 V. Ionisation mode, cone voltage and
180 collision energies were optimised for each analyte in order to obtain two multiple
181 reaction monitoring (MRM) transitions. The conditions are described in Table 1.

182

183 **2.4. Stir bar sorptive extraction**

184

185 The SBSE procedure was as follows: the stir bar was inserted into a 100-mL flask
186 with 50 mL of sample adjusted to pH 7, containing 15% of NaCl. Samples were
187 stirred at 500 rpm for 1 hour at room temperature (25°C). After extraction, the stir
188 bar was removed from the sample using magnetic tweezers and dried with a lint-
189 free tissue. In terms of the liquid desorption (LD), the stir bar was immersed in a vial
190 containing 5 mL of a mixture of MeOH/ACN (1/1, v/v), stirring at the same speed for
191 10 min. The extracts were then evaporated to dryness under a gentle stream of N₂.
192 Prior to the LC injection, the extracts were redissolved in 1 mL of MeOH/H₂O (1/1,
193 v/v). After each use, the stir bar was cleaned 3 times with 5 mL MeOH/ACN (1/1,
194 v/v), stirring for 10 min. Each stir bar was reused at least 20-30 times with
195 environmental samples.

196

197 All of the environmental water samples (from an influent, secondary effluent and
198 tertiary effluent of a domestic sewage treatment plant (STP)) were collected in pre-
199 cleaned amber glass bottles, filtered using a 0.45 µm nylon membrane (Supelco,
200 Bellefonte, PA, USA), acidified to pH 3 (HCOOH) and stored at 4°C until analysis.

201

202 **3. Results and discussion**

203

204 **3.1. LC-(ESI)MS/MS conditions**

205

206 MS/MS parameters were also optimised by injecting each compound at 250 µg L⁻¹
207 in MeOH/H₂O (1/1, v/v) individually in flow injection analysis (FIA). Depending on
208 their acidic or basic properties, the studied compounds were divided, individually or

209 in groups, into six windows applying negative or positive ESI ionisation. In addition,
210 identification and confirmation MRM transitions were selected for each analyte. The
211 optimised cone voltage and collision energy for each MRM transition are
212 summarised in Table 1.

213

214 All of the selected compounds showed good linearity ($r^2 \geq 0.997$) by direct injection
215 with a linear range of 0.50-200 $\mu\text{g L}^{-1}$, except for paracetamol, antipyrine, DHMB,
216 triclosan (2-200 $\mu\text{g L}^{-1}$) and pridinol (25-200 $\mu\text{g L}^{-1}$). The limits of detection (LODs),
217 calculated as signal-to-noise ratio (S/N) of 3, ranged from 0.15 to 0.50 $\mu\text{g L}^{-1}$,
218 except for pridinol (5 $\mu\text{g L}^{-1}$).

219

220 **3.2. Optimisation of the SBSE procedure**

221

222 To overcome the lack of availability of polar commercial stir bars, over the last few
223 years, several monomers with different polarities have been combined in order to
224 synthesise polar monoliths [1,16]. There are several monomers whose polarity
225 properties might be very beneficial. However, they have not been used in the
226 preparation of a monolith yet. In our previous research [17], a monolith coating for
227 SBSE based on poly(PEGMA-co-PETRA) was first synthesised. The choice of
228 these monomers was dependent on the final chemical and physical properties of
229 the desired monolithic material. Therefore, thanks to the hydroxyl and ester
230 functional groups of both PEGMA and PETRA, the resulting monolith was expected
231 to contribute to the sorption of more polar compounds onto the coating. In addition,
232 the use of PEGMA resulted in a gelatinous polymer, providing mechanical stability
233 and swelling capacity.

234

235 Subsequently, the evaluation of this polar monolithic material as SBSE coating was
236 performed. Bearing in mind the polar structure of the resulting monolith, a group of
237 PPCPs with different polarities was selected to test the extraction efficiency of our
238 in-house coating and compare it with the commercial stir bars [11,15,18]. Then, in
239 order to obtain high extraction efficiencies, several main parameters involved in the

240 extraction and desorption steps were optimised. In this study, desorption was
241 performed by LD because the analytes were then separated by LC.

242

243 **3.2.1. Extraction conditions**

244

245 In terms of the optimisation of the extraction conditions, several factors, including
246 sample pH, ionic strength, extraction time, agitation speed and temperature, have
247 been taken into account for SBSE performance.

248

249 The SBSE procedure was optimised under these initial conditions: 50 mL ultrapure
250 water spiked at $2 \mu\text{g L}^{-1}$ with the analytes' mixture stirring at 500 rpm for 1 hour at
251 room temperature, and the LD was performed using 5 mL of MeOH stirring at the
252 same speed for 10 min.

253

254 The effect of the sample pH on the analytes' response was evaluated at different
255 pH values (3, 5, 7 and 9) because it may influence the analytes' sorption onto the
256 coating. As expected, all of the analytes behaved according to their pK_a values (Fig.
257 1). For instance, diclofenac ($\text{pK}_a = 4.2$) showed a lower response when the pH was
258 different from 3 because it was turned into its anionic form. In contrast, those
259 compounds with pK_a around 9 presented higher signal response when higher pH
260 values were applied (from 3 to 7) because they were uncharged, promoting the
261 hydrophilic interactions with the coating, and their sorption onto it decreased to pH 9
262 because these analytes were almost in their anionic state. In addition,
263 carbamazepine and caffeine showed an increase in their response when the
264 sample pH was higher (from 3 to 9) due to their high pK_a values (> 13). The above
265 trend can be seen in Fig. 2 where the effect of the sample pH for a representative
266 group of analytes is shown. Since most of the target analytes showed better
267 responses at pH 7, this pH value was selected as the optimal sample pH for further
268 research.

269

270 The addition of salts, such as NaCl, has been demonstrated to modify the ionic
271 strength of the sample solution and, as a consequence, increase the response of
272 more polar analytes [1]. Therefore, the ionic strength of the sample was adjusted by
273 the addition of NaCl from 0% to 30% w/v. In the range from 0 to 15%, it was
274 observed that almost all the analytes presented higher sorption onto the coating at
275 higher percentages of NaCl. However, with the highest percentage (i.e., 30%) of
276 NaCl, their signal responses decreased. These results could be explained by the
277 salting-out effect, which provided higher responses for the analytes, and then, when
278 30% of NaCl was added, the electrostatic interactions between the analytes and the
279 salt ions caused a decrease in their signal responses [19]. Therefore, the addition of
280 15% of NaCl was chosen to provide better results for most of the studied analytes.

281

282 Regarding extraction time, the SBSE time was varied between 1 and 8 h. Fig. 3
283 details the effect of the extraction time on the analytes' response for a
284 representative group of compounds. It was observed that the analytes almost
285 reached the equilibrium in just 1 hour, except triclocarban, whose sorption onto the
286 coating was lower than the rest of the compounds due to its apolar behaviour.
287 When longer extraction times were applied, no significant improvement on the
288 analytes' response was observed. As a compromise between the sensitivity and the
289 time consumed, 1 h was selected as the optimal extraction time. This extraction
290 step was performed in a short period of time in comparison with those provided by
291 other in-house or commercial stir bars reported previously, which needed between
292 2 and 6 h to guarantee the achievement of the equilibrium at similar stirring rates or
293 even higher [10,11,15,18,20,21]. As mentioned previously, the presence of PEGMA
294 provides high permeability and swelling capacity to the monolith, which enables the
295 analytes to go more easily into the in-house coating and interact quickly with the
296 polar functionalities, providing a rapid SBSE performance.

297

298 The agitation speed was also studied and two agitation rates (500 and 1000 rpm)
299 were tested. It was observed that similar results were obtained at both agitation
300 speeds. However, it is expected that at a higher agitation speed, the monolithic

301 coating would be damaged. Thus, to keep the integrity of our in-house stir bar, the
302 agitation speed was set at 500 rpm for further analyses.

303

304 Finally, the extraction temperature was maintained constant at 25°C during SBSE
305 experiments, since no marked improvement in the analytes' response was
306 observed at higher temperatures.

307

308 **3.2.2. Liquid desorption conditions**

309

310 To ensure the complete desorption of the analytes, different solvents were tested
311 under the optimised extraction conditions: MeOH, ACN and a mixture of
312 MeOH/ACN (1/1, v/v). When 5 mL of these desorption solvents were evaluated, it
313 was observed that ACN provided a higher signal response than MeOH for all of the
314 compounds. Moreover, stirring with 5 mL of a mixture of MeOH/ACN (1/1, v/v),
315 slightly higher analyte responses were obtained in comparison with ACN.
316 Therefore, a mixture of MeOH/ACN (1/1, v/v) was chosen as the desorption solvent
317 for the subsequent experiments. When the solvent volume was increased up to 10
318 mL, the LD results did not improve. Finally, 5 mL of a mixture of MeOH/ACN (1/1,
319 v/v) was selected to provide the best results and 5 mL was the minimum volume
320 necessary to cover the in-house stir bar sufficiently, immersed in the desorption
321 solvent.

322

323 In order to ensure the complete desorption of the target analytes from the polar
324 coating, the desorption time was also investigated between 5 and 30 min. It was
325 observed that stirring for 10 min was long enough to desorb all of the analytes from
326 the in-house stir bar and it was selected as the optimal desorption time.

327

328 After the optimisation procedure, the SBSE conditions used for a further application
329 in environmental waters were as follows: 50 mL of sample at pH 7, containing 15%
330 of NaCl, extracted at room temperature by stirring at 500 rpm for 1h for the
331 extraction and 5 mL of a mixture of MeOH/ACN (1/1, v/v), stirring at the same

332 speed for 10 min for the LD. Under the optimal conditions, the recovery values (%)
333 for each analyte in ultrapure water are detailed in Table 2. The in-house stir bar was
334 able to extract most of the compounds with recovery values between 13% and 64%,
335 except for paracetamol, caffeine, benzotriazole and antipyrine (2% to 9%).

336

337 **3.4. Comparison to commercial stir bars**

338

339 The SBSE performance of the poly(PEGMA-co-PETRA) monolithic coating was
340 compared with two commercially available polar stir bars (EG Silicone Twister[®] and
341 Acrylate Twister[®]), which have recently been commercialised by Gerstel. These
342 commercial stir bars have been applied for the extraction of EOCs from
343 environmental waters [11]. From that study, it was observed that, in fact, the polar
344 commercial coatings were not as effective as expected under their optimised SBSE
345 conditions. Table 2 shows the low recovery values obtained when both
346 commercially available coatings were applied in SBSE of ultrapure water. While the
347 Acrylate Twister[®] was not able to recover even the most apolar compounds, such
348 as DHMB, BP-3 or triclocarban, the EG Silicone Twister[®] provided slightly better
349 SBSE performance for those compounds with more apolar behaviour. However,
350 both commercial coatings were barely capable of extracting the most polar analytes
351 (%R < 3% for those with log $K_{o/w}$ < 3.4) even after a 4 hours extraction. In contrast,
352 the SBSE using the in-house coating provided higher recovery values for more
353 polar compounds, extracting for only 1 h, especially for those analytes with log $K_{o/w}$
354 between 1.9 and 3.4, achieving recovery values from 13% to 55%. This notable
355 improvement could be explained by the fact that a high number of hydroxyl and
356 ester functional groups are present in the monolithic structure.

357

358 Most of the in-house coatings synthesised in the past have shown progress in terms
359 of extraction efficiencies and sensitivity. However, their use still involved long
360 extraction times (< 2 h) [14,15,18,19]. In this paper, the use of the in-house
361 monolith as the SBSE coating enabled better and less time-consuming extraction of
362 the studied analytes than those using the commercial or in-house stir bars, due to

363 its suitable physical properties and high swelling capacity, promoting the diffusion of
364 the analyte through the monolithic material. Moreover, the high permeability and
365 hydrophilicity of PEG-acrylate or PEG-diacrylate monomers have recently reported
366 when they were used to synthesise monolithic columns applied in normal phase
367 and capillary LC for the separation of small molecules, such as peptides and
368 proteins [22-25].

369

370 **3.5. Application to environmental water samples**

371

372 After SBSE optimisation, the in-house stir bar was applied for the extraction of
373 PPCPs in environmental waters, such as influent and secondary effluent samples
374 from a wastewater treatment plant (WWTP).

375

376 Taking into account the application of SBSE in complex matrices and the use of a
377 ESI interface in LC-MS/MS, the effect of the ion suppression/enhancement was
378 evaluated for influent and effluent wastewaters and calculated as the percentage
379 decrease in the signal intensity obtained by the target analytes spiked at 100 µg/L
380 after the extraction of 50 mL of real sample versus the intensity of the same amount
381 of the analytes in ultrapure water [26]. As expected, higher values of ion
382 suppression/enhancement were obtained in influent than in secondary effluent
383 wastewaters due to the complexity of the matrix. For secondary effluent wastewater
384 samples, this effect ranged from -6% to 27%, except for paracetamol and caffeine,
385 whose ion suppression values were the highest ones, 31% and 43%, respectively.
386 In contrast, the ion suppression/enhancement values for influent wastewaters
387 increased and ranged between -22% and 24%, except for paracetamol, caffeine,
388 methylparaben, propylparaben and triclocarban (29% to 48%). Moreover, similar
389 results with respect to this effect were obtained when these analytes were extracted
390 using commercial stir bars, providing ion suppression/enhancement values between
391 -5% and 25%, except for propylparaben (44%) and triclocarban (54%) [11].
392 Although the resulting effect of ion suppression/enhancement was slightly
393 significant for some compounds, especially the most polar ones, higher matrix

394 effects (from -40% to 65%) were found when using other extraction techniques,
395 such as solid-phase extraction (SPE), in comparison to SBSE, as reported
396 previously [27-29].

397

398 The recovery values of the analytes when the in-house stir bar was used to extract
399 PPCPs from wastewater samples are listed in Table 3. The recoveries obtained for
400 both secondary effluent and influent wastewaters were slightly lower than those
401 achieved in ultrapure water (15% to 54% for the most apolar compounds). It should
402 be pointed out that, taking into account the poor sorption of pridinol onto the
403 coating, it was expected that this compound would not be recovered from real
404 samples.

405

406 Once the applicability of the in-house stir bar was demonstrated in real samples and
407 in order to compensate the ion suppression/enhancement, the method was
408 validated using a matrix-matched calibration with 50 mL of secondary effluent
409 wastewater. First of all, a blank sample was analysed and a number of compounds,
410 such as propranolol, carbamazepine, benzylparaben, BP-3, triclocarban and
411 benzotriazole, were detected. The calibration was then achieved by subtracting the
412 signal of existing analytes in the blank. All of the compounds showed good linearity
413 in a range between 50 and 5000 ng L⁻¹, except for DHMB (100-5000 ng L⁻¹),
414 diclofenac (500-5000 ng L⁻¹) and triclocarban (100-5000 ng L⁻¹), with regression
415 coefficients (r^2) greater than 0.997 (except for propranolol ($r^2 > 0.991$)). The limits of
416 quantification (LOQs) obtained were between 50 and 500 ng L⁻¹ and the LODs
417 ranged from 15 to 20 ng L⁻¹ for all of the compounds, except for diclofenac (50 ng L⁻¹)
418 ¹). Particularly, in the case of benzotriazole, which registered a high signal in the
419 blank, the LOD was tentatively calculated as three times the standard deviation of
420 the analyte signal in the blank ($n=3$). The repeatability and reproducibility between
421 days of three samples spiked at 2000 ng L⁻¹, expressed as % relative standard
422 deviation (% RSD), were lower than 16% and 20%, respectively. In the present
423 work, apart from the development of a rapid SBSE, this analytical methodology
424 enabled a suitable method validation for all of the studied compounds, including

425 those with polar behaviour, which were hardly recovered at all using the
426 commercially available stir bars and for which, consequently, the validation method
427 could not be applied [11].

428

429 The SBSE/LC-MS/MS method was then applied to the analysis of different
430 environmental samples, such as influent, secondary effluent and tertiary effluent
431 wastewaters. It is important to mention that the tertiary effluent wastewaters were
432 included in the following quantification because the effective treatment plant, based
433 on reverse osmosis, should provide clean water samples in order to reuse them
434 again. For this reason, it was also decided to check the occurrence of the studied
435 analytes at this point. The concentrations found for each analyte in the different
436 wastewater samples are shown in Table 4. As can be seen, all of the studied
437 analytes were detected in the analysed samples, except antipyrine, DHMB and
438 triclosan. In influent wastewaters, the highest levels of ng L^{-1} were recorded for
439 paracetamol, caffeine, benzotriazole, propylparaben, diclofenac and BP-3 ($626.9 -$
440 $> 5000 \text{ ng L}^{-1}$). After the secondary treatment, it was observed that most of the
441 studied compounds were effectively removed from the wastewaters, except
442 benzotriazole, propranolol, carbamazepine, diclofenac, BP-3 and triclocarban.
443 However, some analytes, such as benzotriazole, DHB and BP-3, still remained,
444 even after the tertiary treatment ($62.1, 155.9$ and 160.7 ng L^{-1} , respectively), being
445 discharged into groundwaters and surface waters to reuse them. The presence of
446 this group of PPCPs in similar samples was reported previously [18,30,31] with
447 results that are strongly in line with the present study.

448

449 **4. Conclusions**

450

451 After an optimisation procedure including both synthesis and extraction, this was the
452 first time that a polar monolithic coating for SBSE based on poly(PEGMA-co-
453 PETRA) has successfully been applied for the extraction of a group of PPCPs from
454 wastewaters. The presence of PEGMA provides a high degree of permeability in the

455 polymer and promotes a high diffusion of the analytes, allowing rapid extraction in
456 just 1 h.

457

458 Moreover, the results obtained in terms of extraction efficiencies for both polar and
459 apolar compounds were better than those obtained using two new commercially
460 available polar coatings (EG Silicone Twister[®] and Acrylate Twister[®]).

461

462 The combination of SBSE/LC-MS/MS was demonstrated for application in complex
463 matrices, such as influent and effluent wastewaters, since no significant ion
464 suppression/enhancement values were obtained for hardly any of the compounds
465 (from -22% to 24%), except for paracetamol, caffeine, propylparaben and
466 triclocarban.

467

468 Finally, the developed analytical methodology was evaluated in terms of linearity,
469 LODs and precision, achieving low LODs (15-50 ng L⁻¹) for all of the selected
470 compounds, overcoming the poor sorption of polar compounds onto the commercial
471 stir bars. Moreover, the in-house coating for SBSE followed by LC-MS/MS allowed
472 the detection and quantification of a broader range of analytes compared to
473 commercial coatings in environmental water samples.

474

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481 (FI-DGR 2011).

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483 **References**

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531

532 **Figure captions**

533

534 **Fig. 1.** Chemical structures, pK_a and $\log K_{o/w}$ of target analytes.

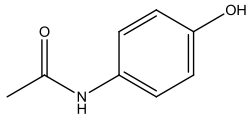
535

536 **Fig. 2.** The effect of sample pH on the analyte response for a representative group
537 of compounds, whose pK_a values are detailed in brackets.

538

539 **Fig. 3.** The effect of extraction time on the analyte response for a representative
540 group of compounds.

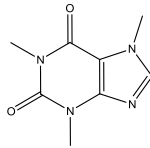
Figure



Paracetamol

pK_a 9.2

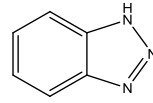
Log $K_{o/w}$ 0.5



Caffeine

pK_a 13.4

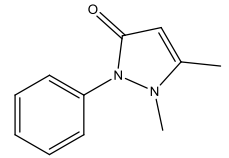
Log $K_{o/w}$ -0.6



Benzotriazole

pK_a 8.5

Log $K_{o/w}$ 0.4



Antipyrine

pK_a 13.3

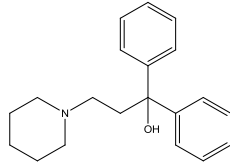
Log $K_{o/w}$ 1.4



Propranolol

pK_a 9.5

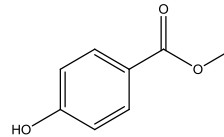
Log $K_{o/w}$ 2.9



Pridinol

pK_a 9.7

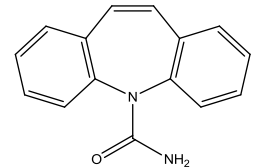
Log $K_{o/w}$ 3.4



Methylparaben

pK_a 8.3

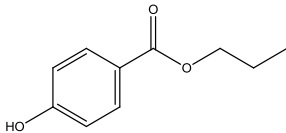
Log $K_{o/w}$ 1.9



Carbamazepine

pK_a 13.7

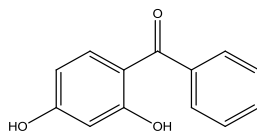
Log $K_{o/w}$ 1.9



Propylparaben

pK_a 8.2

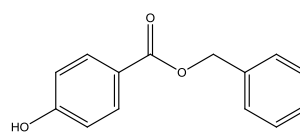
Log $K_{o/w}$ 2.9



DHB

pK_a 7.7

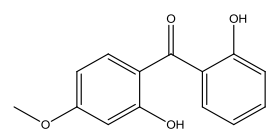
Log $K_{o/w}$ 3.2



Benzylparaben

pK_a 8.2

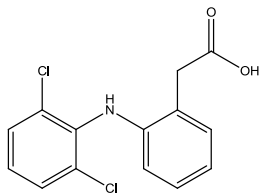
Log $K_{o/w}$ 3.6



DHMB

pK_a 7.1

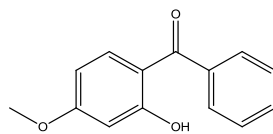
Log $K_{o/w}$ 4.3



Diclofenac

pK_a 4.2

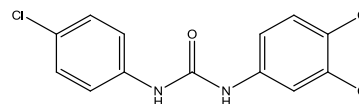
Log $K_{o/w}$ 4.5



BP-3

pK_a 7.6

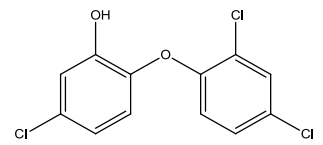
Log $K_{o/w}$ 4.0



Triclocarban

pK_a 12.7

Log $K_{o/w}$ 6.1



Triclosan

pK_a 7.9

Log $K_{o/w}$ 5.3

* pK_a and Log $K_{o/w}$ calculated using Advanced Chemistry Development (ACD/Labs) Software V11.02 (© 1994-2012 ACD/Labs)

Figure

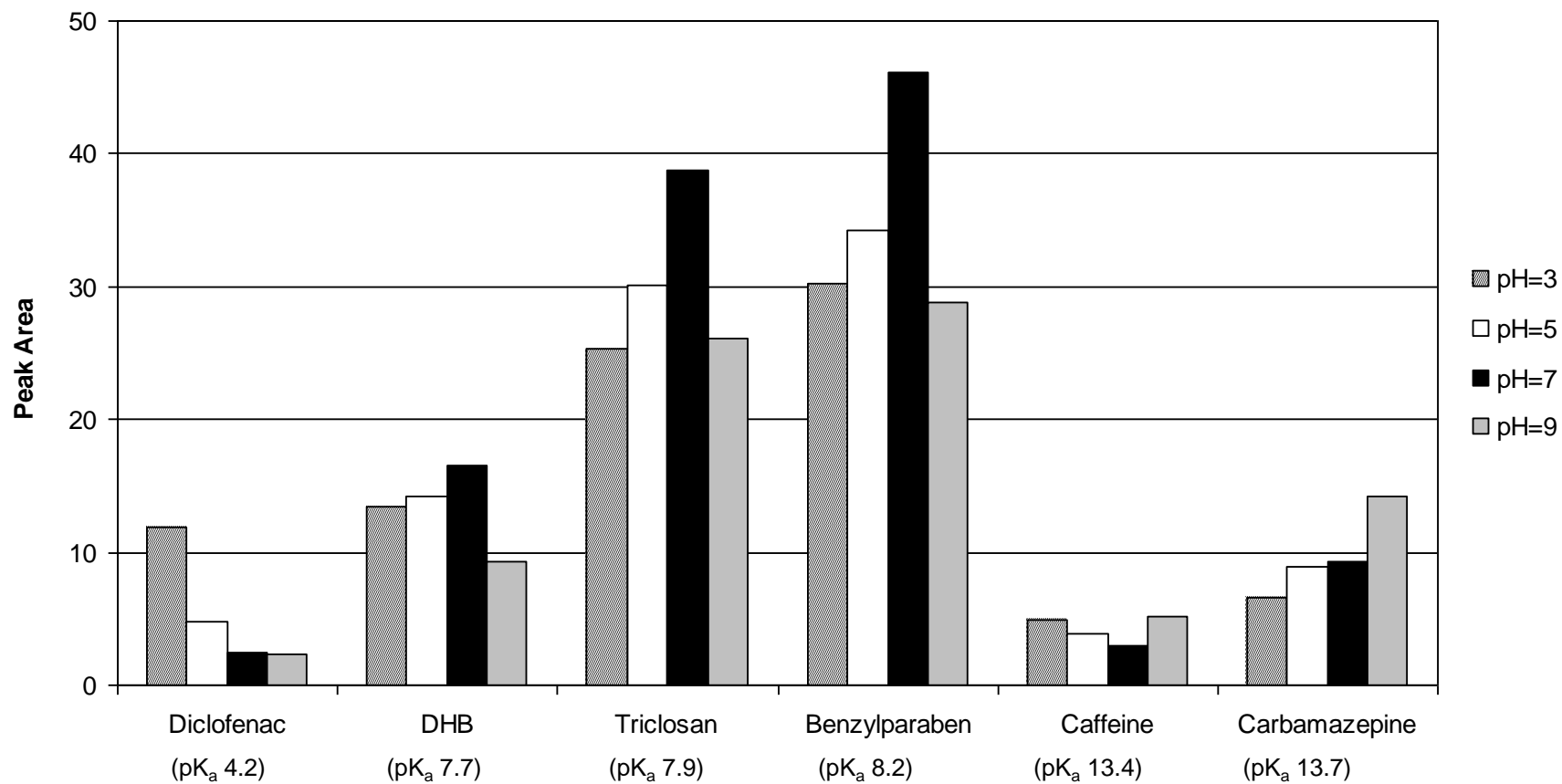


Figure 2

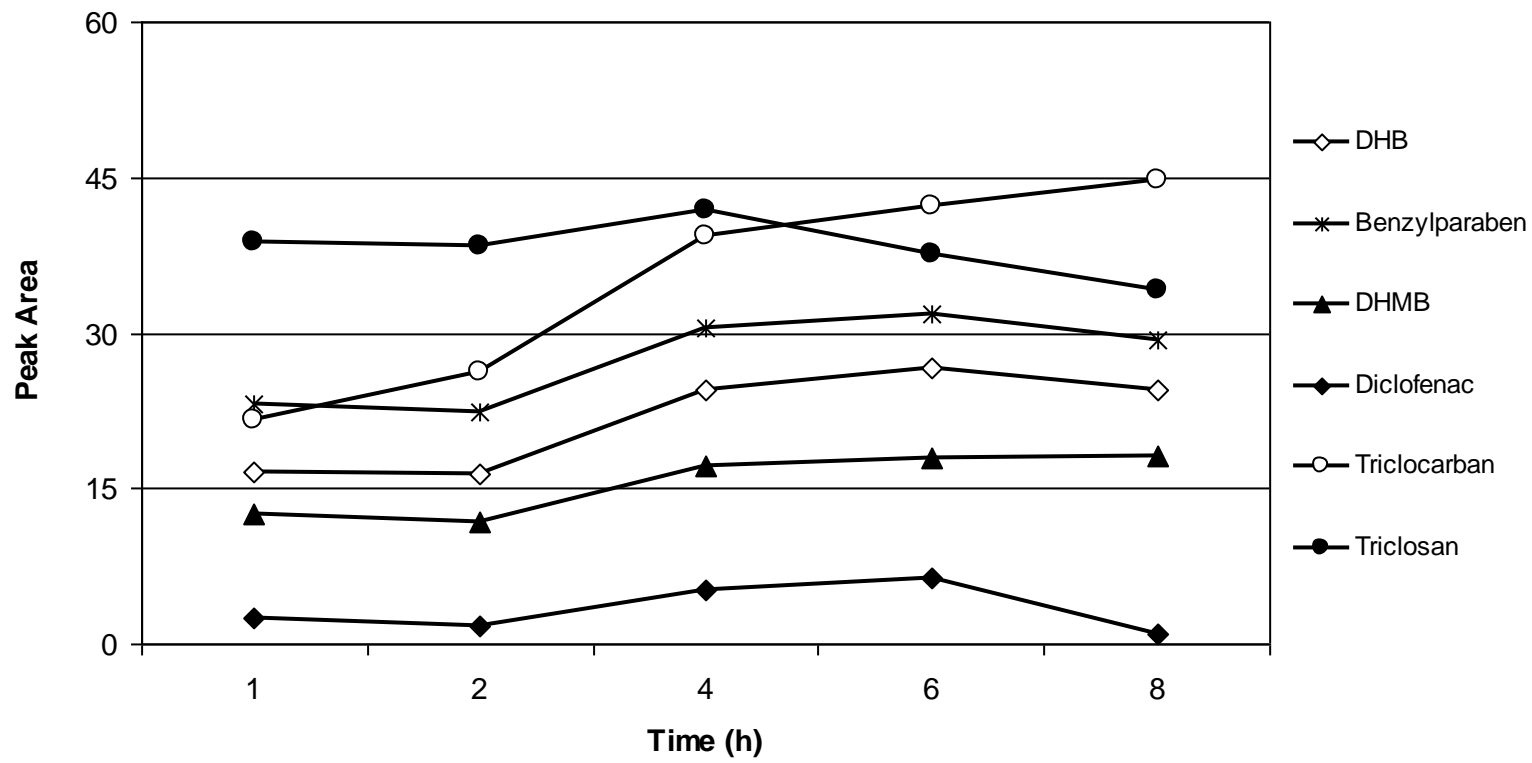


Figure 3

1 **Table 1.** LC-(ESI)MS/MS acquisition parameters in MRM mode for the analysis of
 2 the target analytes.
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Analyte	Precursor ion (m/z)	Cone voltage (V)	Product ion (m/z) (Collision energy (V))		Ionisation mode (ESI)
			Identification	Confirmation	
Paracetamol	152	100	110 (15)	93 (25)	+
Caffeine	195	125	138 (15)	110 (25)	+
Benzotriazole	120	100	65 (25)	92 (15)	+
Antipyrine	189	100	145 (30)	115 (30)	+
Propranolol	260	125	116 (15)	183 (15)	+
Pridinol	296	125	115 (30)	193 (30)	+
Methylparaben	151	80	92 (15)	136 (5)	-
Carbamazepine	237	150	193 (35)	179 (35)	+
Propylparaben	179	100	92 (15)	136 (5)	-
DHB	213	130	135 (15)	169 (5)	-
Benzylparaben	227	100	92 (10)	136 (20)	-
DHMB	243	80	93 (15)	123 (15)	-
Diclofenac	294	75	250 (5)	214 (15)	-
BP-3	229	130	151(15)	105 (15)	+
Triclocarban	313	130	160 (5)	126 (15)	-
Triclosan	287 / 289	8	35 (18)	35 (18)	-

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20 **Table 2.** Recovery values (%) obtained when different coatings were applied in
 21 SBSE of ultrapure water.

22 23 24 25 26 27 28 29 30 31 32 33 34 35	Analyte	Recovery values (%)		
		Poly(PEGMA-co-PETRA) ^{a)}	EG Silicone Twister [®] ^{b)}	Acrylate Twister [®] ^{b)}
24	Paracetamol	5	-	<1
25	Caffeine	2	-	-
26	Benzotriazole	9	<1	<1
27	Antipyrine	2	<1	1
28	Propranolol	19	2	2
29	Pridinol	13	3	2
30	Methylparaben	20	1	2
31	Carbamazepine	25	<1	<1
32	Propylparaben	38	10	2
33	DHB	55	24	9
34	Benzylparaben	64	39	14
35	DHMB	56	26	9
36	Diclofenac	33	<1	<1
37	BP-3	55	45	10
38	Triclocarban	51	59	43
39	Triclosan	55	80	42

^{a)} 50 mL of sample spiked at 2 µg L⁻¹

^{b)} 50 mL of sample spiked at 4 µg L⁻¹ [11]

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50 **Table 3.** Recovery values (%) obtained for the studied analytes when the
 51 poly(PEGMA-co-PETRA) coating was applied in SBSE followed by LC-MS/MS of
 52 50 mL of water samples spiked at 2 µg L⁻¹ with the analyte mixture.

Analyte	Recovery values (%)	
	Secondary effluent WWTP ^{a)}	Influent WWTP ^{b)}
Paracetamol	3	4
Caffeine	1	2
Benzotriazole	8	9
Antipyrine	2	2
Propranolol	15	17
Pridinol	-	-
Methylparaben	12	14
Carbamazepine	19	24
Propylparaben	24	27
DHB	28	37
Benzylparaben	45	50
DHMB	36	42
Diclofenac	25	31
BP-3	44	54
Triclocarban	49	31
Triclosan	43	43

^{a)} % RSD (*n*=3) < 15% for %R > 12%

^{b)} % RSD (*n*=3) < 18% for %R > 14%

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81 **Table 4.** Concentrations of found analytes in influent, secondary effluent and
 82 tertiary effluent wastewater samples when they were analysed using the in-house
 83 stir bar for SBSE followed by LC-MS/MS.

Analyte	Concentration (ng L ⁻¹)		
	Influent WWTP	Secondary effluent WWTP	Tertiary effluent WWTP
Paracetamol	> 5000	n.d.	n.d.
Caffeine	> 5000	<LOD	n.d.
Benzotriazole	3005.4	609.9	155.9
Antipyrine	n.d.	n.d.	n.d.
Propranolol	<LOQ	83.5	<LOQ
Pridinol	n.d.	n.d.	n.d.
Methylparaben	> 5000	n.d.	n.d.
Carbamazepine	n.d.	245.9	n.d.
Propylparaben	2369.4	n.d.	n.d.
DHB	372.1	<LOQ	62.1
Benzylparaben	<LOD	<LOQ	<LOQ
DHMB	n.d.	n.d.	n.d.
Diclofenac	966.0	641.0	<LOD
BP-3	626.9	217.6	160.7
Triclocarban	n.d.	125.0	<LOQ
Triclosan	n.d.	n.d.	n.d.

n.d. = non detected
 % RSD (*n*=3) < 23%

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